

More intensive use and lifetime extension can enable net-zero emissions in China's cement cycle

Tongyuan Wu¹, Thomas S.T. Ng^{2,}, Ji Chen¹, Zhi Cao^{3,4*}*

¹Department of Civil Engineering, The University of Hong Kong, Hong Kong, China

²Department of Architecture and Civil Engineering, City University of Hong Kong, Hong Kong, China

³College of Environmental Science and Engineering, Nankai University, 38 Tongyan Road, Jinnan District, Tianjin, 300350, China

⁴Energy and Materials in Infrastructure and Buildings (EMID), University of Antwerp, Antwerp, Belgium

* Corresponding author

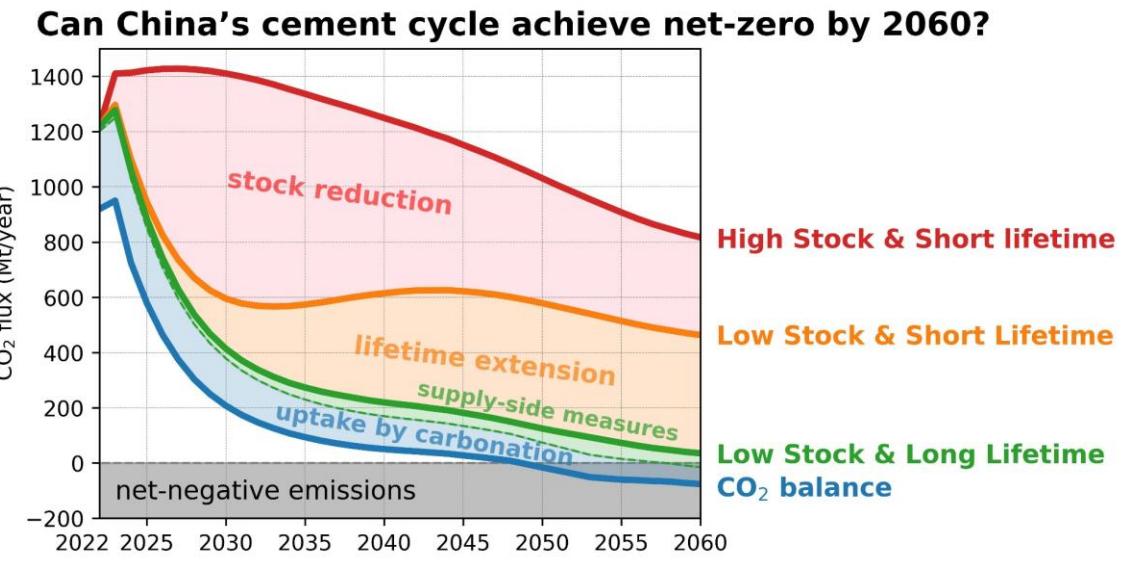
Email: thomasng@cityu.edu.hk; zhicaoie@gmail.com;

1 ***Abstract***

2 Demand-side material efficiency strategies could play a key role in decarbonizing the
3 cement cycle as conventional supply-side measures leave little room for improvement
4 and emerging technologies are still in their infancy and expensive. This study
5 quantitatively evaluates CO₂ reduction opportunities through China's cement cycle by
6 more intensive use and lifetime extension during 2023-2060. More intensive use and
7 lifetime extension can cumulatively save cement consumption by ~58 gigatons (Gt)
8 during 2023-2060, resulting in CO₂ emission reductions of ~33.8 Gt without considering
9 supply-side actions and cement carbonation. If supply-side measures and cement
10 carbonation are considered, China's cement cycle could achieve net-zero CO₂ emissions
11 by 2060 or even mid-century. Besides, it is important to assess social implications,
12 opportunities, and challenges of lower per-capita stocks and longer service-life, when we
13 aim to realize these emission reductions. Also, trade-offs between CO₂ fluxes should be
14 considered in cement decarbonization road mapping because more intensive use and
15 lifetime extension can lower CO₂ uptake by cement carbonation.

16 **Keywords:** cement cycle; dynamic material flow analysis; material efficiency; net-zero

17 *Graphic Abstract*



18

19 ***1. Introduction***

20 Up to 4 gigatons (Gt) of cement was produced to glue gravels and sands together to form
21 concrete or mortar for satisfying the growing demand for buildings, roads, tunnels, water
22 treatments and delivery systems, sewage lines, and dams around the world (Andrew,
23 2019). China has emerged as the largest cement producer in the world since its economic
24 reform. Massive production and use of cement, driven by soaring needs for housing and
25 infrastructure development, has positioned the cement sector as a significant source of
26 CO₂ emissions. The sheer volume of cement production released substantial CO₂
27 emissions, despite a lower emission intensity of around 600 kg CO₂ per tonne of cement
28 compared with metals (Fennell et al., 2021). Annual CO₂ emissions from China's cement
29 sector reached 1.3 Gt in 2020 (Wu et al., 2022). Without fundamental changes, China's
30 cement sector will continue to be a significant source of CO₂ emissions, posing
31 significant risks to climate change goals.

32 Yet, decoupling CO₂ emissions from cement production is a challenging task, because
33 two-thirds of its CO₂ emissions come from carbonate decomposition during calcination.
34 In addition, large-scale deployment of alternative cement is unlikely to occur due to
35 feedstock constraints and high manufacturing costs (Habert et al., 2020; Monteiro et al.,
36 2017). Consequently, beyond conventional supply-side efforts, demand-side measures
37 based on the philosophy ‘using less cement’, i.e., material efficiency strategies, could
38 potentially open up the option space for decarbonizing the cement sector in China
39 (Fennell et al., 2022; Miller et al., 2021).

40 Demand-side material efficiency measures aim to achieve desired levels of service from
41 smaller cement stocks (Hertwich et al., 2019), including more intensive use (Grubler et

42 al., 2018; Pauliuk et al., 2013) and lifetime extension (Cai et al., 2015; Miller, 2020) of
43 in-use product stock, material-efficient design (Cao et al., 2021; Shanks et al., 2019),
44 material substitution (Churkina et al., 2020; Göswein et al., 2022), as well as reusing
45 building components (Eberhardt et al., 2018; Huuhka et al., 2015) and recycling (Wang et
46 al., 2021; Xiao et al., 2012) at the end-of-life. More intensive use of residential and
47 commercial buildings by 20% lower area per person could globally reduce 56.8 Gt CO₂-
48 eq greenhouse gas (GHG) emissions from material production during 2020-2060 (Zhong
49 et al., 2021). Lifetime extension could significantly reduce resource demand to replace
50 demolished buildings and infrastructures by prolonging their service-life (Cai et al., 2015;
51 Miller, 2020). Retrospective analysis showed that a 50% lifetime extension for in-use
52 cement stock could cumulatively lead to a 14% reduction in cement demand and ~0.7 Gt
53 CO₂ emissions from cement production from 1900 to 2015 in the United States (Miller,
54 2020), whereas extending building lifetime to 50 years could reduce annual CO₂
55 emissions by 426 Mt per year in China (Cai et al., 2015). Material-efficient design aims
56 to avoid the load-bearing capacity of structural elements not fully utilized, which is
57 common in the construction industry due to designers' cautiousness (Cao et al., 2021).
58 Although current practices, e.g., performance-based structure design and post-tensioning
59 concrete elements (Shanks et al., 2019), could potentially reduce cementitious binder
60 intensity by ~20%, wide-scale deployment of these design-oriented measures will hinge
61 on how fast designers or contractors will take them up, as well as adequate policy
62 interventions (Cao et al., 2021). Material substitution tries to substitute emission-
63 intensive construction materials (e.g., concrete and steel) by bio-based materials with
64 lower embodied emissions, such as engineered timber (Churkina et al., 2020). An

65 optimistic estimate stated that the widespread implementation of timber construction
66 from 2020 to 2050 has the potential to sequester a minimum of 7 Gt CO₂ (Churkina et al.,
67 2020), whereas the realization of this potential will be constrained by the limited supply
68 capacity from available global forests (Pomponi et al., 2020). As for end-of-life options,
69 efficiently reusing concrete components requires prefabricated or modular construction
70 and standardized design on buildings (Huuhka et al., 2015). Besides, technologies to
71 recycle hydrated cement waste into new cement are still in the development phase and
72 need validation (Gastaldi et al., 2015), though crushed concrete debris could be
73 downcycled as coarse aggregate or roadbed to reduce virgin aggregate consumption (Cao
74 et al., 2021). Whilst previous studies (Pauliuk et al., 2021; Zhong et al., 2021) reveal
75 substantial emissions reductions from employing material efficiency measures in the
76 future building sector globally, how CO₂ fluxes interplay with cement stock dynamics
77 under the impact of a lower per-capita stock through more intensive use and lifetime
78 extension is seldom discussed towards the net-zero future.

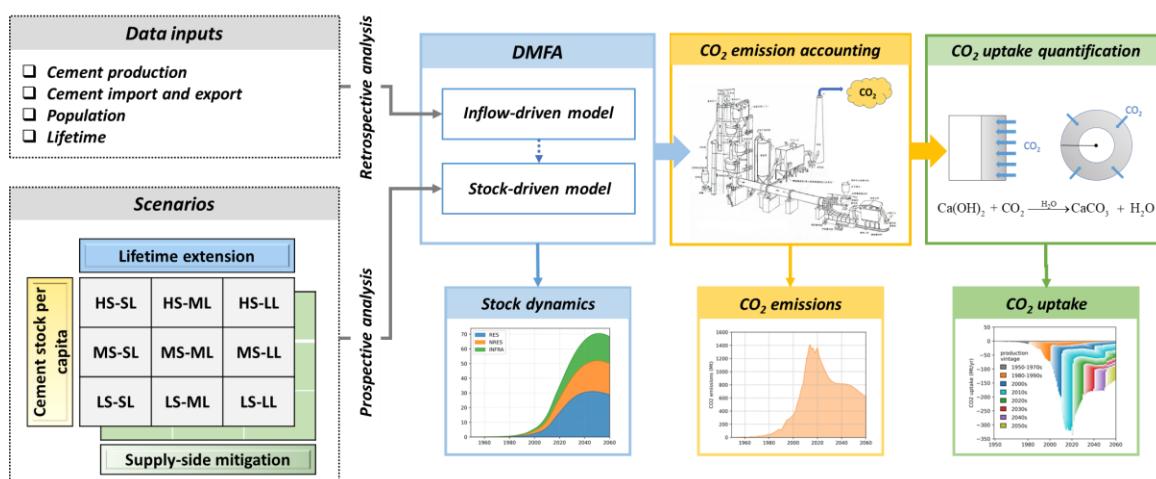
79 Besides as an emitter, cement stock could reabsorb substantial CO₂ by carbonation (Xi et
80 al., 2016), a physicochemical process by which unstable calcium oxide embodied in
81 cement materials will react with atmospheric CO₂ over time (Pade & Guimaraes, 2007).
82 An updated analysis shows that up to 21 Gt CO₂ had been absorbed by cement
83 carbonation during 1930-2019, cumulatively offsetting ~55% of process emissions from
84 production (Guo et al., 2021). In addition to retrospective analysis, a global scenario-
85 based projection highlights that cumulative CO₂ uptake from 2015 to 2100 amounts to
86 81-117 Gt, revealing that the magnitude of this passive sequestration could be greater
87 than the active carbon capture and storage (CCS) prescribed in current technology

88 roadmaps (Cao et al., 2020). Also, for specific countries, cement carbonation could be an
89 important lever to achieve net-zero emissions through the cement and concrete cycle by
90 mid-century (Watari et al., 2022). Whilst the CO₂ uptake by cement carbonation is well
91 recognized in the global carbon cycle (Cao et al., 2020) and carbon budget (Friedlingstein
92 et al., 2022), an explicit understanding of trade-offs between CO₂ emissions and uptake
93 induced by varying stock levels and lifetime scenarios is still lacking, obscuring its role
94 in cement decarbonization roadmaps.

95 Against this background, we developed an integrated modeling framework that allows us
96 to: (1) analyze historical and future cement flows and stocks; (2) consider the interplay
97 between CO₂ fluxes and cement stock dynamics; and (3) quantify net changes in CO₂
98 fluxes associated with China's cement cycle. With this integrated modeling framework,
99 we designed wide-ranging scenarios to explore various decarbonization possibilities for
100 China's cement cycle and evaluate the role of cement stock reduction and lifetime
101 extension.

102 **2. Methods and data**

103 The modeling framework consists of three modules: dynamic material flow analysis
 104 (DMFA), CO₂ emissions accounting, and cement uptake estimation. First, we collected
 105 statistics on China's annual cement production, import, and export to estimate the
 106 historical total and per capita in-use cement stock during 1949-2022 by the inflow-driven
 107 method. With historical total and per capita in-use cement stock, we employed a stock-
 108 driven method to project future cement consumption from 2023 to 2060, considering nine
 109 (3×3) demand-side scenarios to reflect varying levels of lifetime extension and stock
 110 saturation. We then developed a CO₂ emissions accounting model to quantify annual CO₂
 111 emissions associated with cement production. In parallel with CO₂ emissions accounting,
 112 we employed a physicochemical model to estimate CO₂ uptake by cement carbonation
 113 spanning the cement cycle. On top of the demand-side scenarios, we considered supply-
 114 side measures to explore decarbonization opportunities along the cement cycle. By
 115 considering CO₂ emissions and CO₂ uptake, we used net CO₂ emissions to present
 116 simulation outcomes.



117
 118 **Figure 1 | Integrated modeling framework of the material-energy-emission-uptake nexus in the**
 119 **cement cycle (IMAGINE Cement).** Adapted from a previous study (Cao et al., 2021). Note: HS, MS, and

120 LS represent High Stock, Medium Stock, and Low Stock scenarios. And SL, ML, and LL denote Short
121 Lifetime, Moderate Lifetime extension, Substantial Lifetime extension scenarios respectively.
122

123 **2.1 Dynamic material flow analysis**

124 DMFA provides a framework to characterize materials entering and leaving a system
125 over a period of time in a mass-balanced manner (Kapur et al., 2008; Liu & Muller, 2013;
126 Muller et al., 2014). Depending on the goal of DMFA, two methods are usually employed
127 in DMFA: inflow-driven and stock-driven (B. Müller, 2006; Bergsdal et al., 2007). In this
128 study, we employed the inflow-driven method to estimate historical total and per capita
129 in-use cement stocks and the stock-driven method to project future cement consumption
130 from 2023 to 2060.

131 **2.1.1 Inflow-driven method**

132 The underlying equation of the inflow-driven method is described as follows.

133
$$S_t = \sum_{i=1949}^t (I_i - O_i) \quad (1)$$

134 where S_t denotes the material stock within the system boundary in the year t as well as I
135 and O represent material inputs and outputs, respectively. Cement inflow refers to
136 apparent consumption, that is, domestic production plus imports minus exports. Data on
137 cement production, imports, and exports were taken from the data platform of the
138 National Bureau of Statistics (2023). Cement outflow refers to cement leaving from in-
139 use cement stock, which is calculated as follows.

140
$$O_t = \sum_{i=1949}^t I_i \cdot Dr(i-1949) \quad (2)$$

141
$$Dr(T) = \int_{T-1}^T f(x)dx \quad (3)$$

142 where $Dr(T)$ represents the demolition rate for the in-use stock after T years it entered
143 use. It could be calculated from the cumulative distribution function of the lifetime

144 distribution $f(x)$, which describes the probability that a unit will be discarded after the
 145 time T it enters use (B. Müller, 2006). Given that right-skewed distribution is consistent
 146 with the fact that buildings that survive after the demolition peak tend to stay longer than
 147 others in their cohort (Miatto et al., 2017), we chose a two-parameter lognormal
 148 distribution to characterize the lifetime distribution of in-use cement stock (Eq. (4)).

$$149 \quad f(x) = \frac{1}{x\sigma\sqrt{2\pi}} \exp\left(-\frac{(\ln(x)-\mu)^2}{2\sigma^2}\right) \quad (4)$$

150 where μ and σ represent the location and scale parameters respectively.

151 2.1.2 Stock-driven method

152 With estimates of historical per-capita cement stock, we projected future cement inflows
 153 using a stock-driven method (Cao et al., 2020) that considers key drivers, including per-
 154 capita cement stock, population, and lifetime. Annual cement inflows were determined by
 155 summing up net additions to cement stocks and cement outflows (Eq. (5)). Future cement
 156 stocks were determined by multiplying the population by per-capita cement stocks (Eq.
 157 (6)). Due to its abundant resource availability, good workability, long-lasting durability,
 158 and versatility, cement is expected to continue to be a dominating construction material
 159 moving forward. Therefore, a modified Gompertz function (Liu et al., 2012) was used to
 160 simulate the development of per-capita cement stocks, assuming that no abrupt changes
 161 will take place moving forward (Eq. (7)).

$$162 \quad I_t = S_t - S_{t-1} + O_t \quad (5)$$

$$163 \quad S_t = P_t \cdot s_t \quad (6)$$

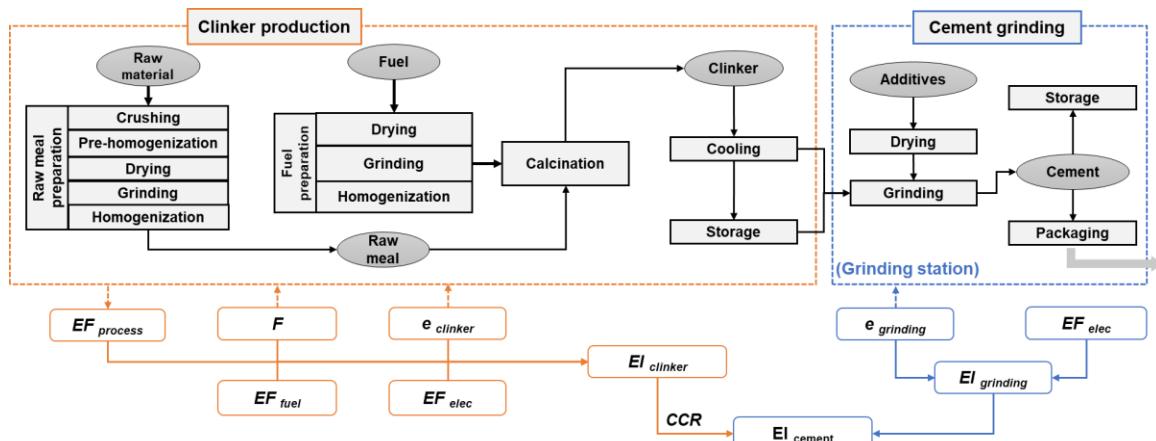
$$164 \quad s_t = \frac{s_{sat}}{1 + \left(\frac{s_{sat}}{s_{2022}} - 1\right) \cdot \exp(A \cdot (1 - \exp(B \cdot (t - 2022))))} \quad (7)$$

165 where P_t denotes population, s_t represents cement stock per capita in the year t , a and b
 166 denote parameters that determine the growth curve and when s_t reaches 98% of the
 167 saturation level s_{sat} at a given time.

168

169 2.2 CO₂ emissions accounting

170 We considered CO₂ emissions associated with the main processing steps in cement
 171 production, including raw materials preparation, clinker production, and cement grinding.
 172 As emissions-related data were drawn from plant-level statistics, we limited the system
 173 boundary of CO₂ emissions accounting to activities that occur within cement plants.



174
 175 **Figure 2 | System boundary of CO₂ emission accounting.** (Notes: EI represents the emission intensity of
 176 a unit process of production. EF denotes the emission factors for fuel combustion or electricity generation.
 177 F and e represent thermal and electrical duty for unit process respectively. CCR denotes the clinker-to-
 178 cement ratio.)

179
 180 CO₂ emissions associated with cement production stem from three sources: carbonate
 181 decomposition, fuel combustion, and electricity use. And the emission intensity, that is
 182 the total CO₂ emission released from the production of the unit product, of cement
 183 (EI_{cement}) and clinker ($EI_{clinker}$) could be calculated in Eq. (8) and Eq. (9), respectively.

184
$$EI_{cement} = EI_{clinker} \cdot CCR + EI_{grinding} \quad (8)$$

185
$$EI_{clinker} = EF_{process} + EI_{fuel} + EI_{elec}^{elec} \quad (9)$$

186 where CCR represents the clinker content in unit product cement, i.e., the clinker-to-
187 cement ratio; EF_{process} denotes the emission factor of limestone decomposition during kiln
188 calcination; EI_{fuel} denotes the emission intensity for fuel combustion during calcination;
189 $EI_{\text{clinker}}^{\text{elec}}$ and EI_{grinding} represents the emission intensity from electricity use by clinker
190 production and cement grinding, respectively. And EI_{fuel} could be broken down in Eq.
191 (10) to consider the deployment of alternative fuel.

$$192 \quad EI_{\text{fuel}} = F \cdot \sum_j w_j \cdot EF_{\text{fuel}}^j \quad (10)$$

$$193 \quad EI_{\text{clinker}}^{\text{elec}} = e_{\text{clinker}} \cdot EF_{\text{elec}} \quad (11)$$

$$194 \quad EI_{\text{grinding}} = e_{\text{grinding}} \cdot EF_{\text{elec}} \quad (12)$$

195 where, F denotes the thermal duty for the unit process of clinker production; w_j denotes
196 the deployment rate of fuel of type j ; EF_{fuel}^j represents emission factor of fuel of type j ;
197 e_{clinker} and e_{grinding} represent electrical duty for the unit process of clinker production and
198 cement grinding, respectively; EF_{elec} denote emission factors for power generation. And
199 the specific values of these parameters are displayed in the Supplementary information.
200 Besides emissions accounting, those cement plants in operation will turnover based on
201 their expected lifetime and the consideration of recent industrial policies (see details in
202 Supplementary Notes B).

203

204 **2.3 Cement carbonation**

205 A physicochemical model was employed to characterize cement carbonation and estimate
206 CO₂ update across the life cycle of cement (Cao et al., 2020; Guo et al., 2021; Watari et
207 al., 2022; Xi et al., 2016). The cement carbonation model succinctly parameterizes the

208 physicochemical carbonation reactions at different life cycle stages (i.e., production, in-
 209 use, demolition, and secondary use) and is populated by region-specific data from field
 210 surveys. We considered four CO₂-absorbing materials: concrete, mortar, construction
 211 waste, and cement kiln dust (CKD). For this model, we made the following two
 212 assumptions. First, cement-based products are characterized by simplified geometries
 213 (see Figure S13). Second, the diffusion front is equivalent to the carbonation front, and
 214 the area behind the front is regarded as fully carbonated (Guo et al., 2021). With these
 215 assumptions, the annual CO₂ uptake by cement and mortar was estimated based on the
 216 annual growth in cumulative carbonation fraction of carbonated cement-based products,
 217 cement content of the material, clinker-to-cement ratio, CaO content of clinker, the
 218 proportion of CaO converted to CaCO₃ within the fully carbonated clinker, and molar
 219 mass ratio of CO₂ and CaO (Eq. (13)). The cumulative carbonation fraction was
 220 determined by the carbonation depth and the thickness of concrete structure components,
 221 mortar layer, or diameter of demolished waste particles, based on the simplified geometry
 222 assumption (Eq. (14)). And carbonation depth was determined using Fick's diffusion law
 223 (Eq. (15)).

$$224 \quad U_t = S_t^{t_1} \cdot \Delta F_t \cdot f_{\text{cement}} \cdot f_{\text{clinker}} \cdot f_{\text{CaO}} \cdot \gamma \cdot M r_{\text{CaO}}^{\text{CO}_2} \quad (13)$$

$$225 \quad F_t = d_t^{t_1} / D \quad (14)$$

$$226 \quad d_t^{t_1} = k \cdot \sqrt{t_1 - t} \quad (15)$$

227 Where U_t is the annual uptake by concrete or mortar products; $S_t^{t_1}$ is the remained in-use
 228 stock that entered into use phase in year t_1 ; F_t is the cumulative carbonated fraction and
 229 ΔF_t is the annual growth in carbonated fraction; f_{cement} is the cement content in concrete
 230 or mortar; f_{clinker} is the clinker content in the cement, i.e., the clinker-to-cement ratio; f_{CaO}

231 is the CaO content in the clinker or CKD; γ is the fraction of CaO that could be
232 converted to CaCO_3 ; $M_{\text{CaO}}^{\text{CO}_2}$ is the molar mass ratio of CO_2 and CaO; k is the
233 carbonation rate coefficient of the material, which is affected by exposure condition, CO_2
234 concentrations, cement additives, and coating and coverings; and $t_1 - t$ is the cumulative
235 exposed time in years; D is the average thickness of concrete structure components or
236 mortar layer. As for construction waste and CKD, we estimated their CO_2 uptake by an
237 annual carbonation fraction of total mass. The reason is that they are small pieces and
238 particles, and we assume that they will be fully carbonated within one to five years (Guo
239 et al., 2021). All the calculation processes and other details of the carbonation model are
240 displayed in Supporting Information 5.

241

242 **2.4 Scenarios and narratives**

243 We designed nine demand-side scenarios to reflect the effects of stock saturation and
244 lifetime extension. More specifically, we constructed three stock-related scenarios (i.e.,
245 **Low Stock**, **Medium Stock**, and **High Stock**) to represent the implementation of demand-
246 side measures that can decouple cement demand from service provision, including more
247 intensive use and material-efficient design (Cao et al., 2021; Cullen & Cooper, 2022;
248 Hertwich et al., 2019; Pauliuk et al., 2021; Zhong et al., 2021) (Table 1). The three stock-
249 related scenarios reflect the deployment levels of these demand-side measures, ranging
250 from 33.5 t/cap to 57.6 t/cap (details see Supporting Information 2). The Low Stock
251 scenario describes a future in which service levels will maintain at current levels by
252 widespread implementation of efficient measures such as more-intensive use, whereas the
253 aggregate cement intensity will grow by ~18% as service quality will be improved by

254 more cement-based products, leading to a per-capita stock that converges to 33.5 t/cap by
 255 2060. Per-capita stock of Medium Stock scenario will grow and converge to 41.3 t/cap by
 256 2060, as both service levels and quality will be improved moderately. And the High
 257 Stock scenario reflects a future in which cement stock per capita in China will grow
 258 significantly to 57.6 t/cap by 2060, as its service levels will catch up with that in
 259 industrialized countries and the aggregate cement intensity will also increase substantially,
 260 due to the widespread implementation of cement-based products.
 261 In addition, we designed three lifetime-related scenarios (i.e., ***Short Lifetime, Moderate***
 262 ***Lifetime Extension, and Substantial Lifetime Extension***) to reflect the implementation
 263 of lifetime extension. In modeling the effects of lifetime extension, we assumed that the
 264 lifetime will gradually increase over time, aiming to avoid abrupt changes in projected
 265 cement inflows. The lifetime of cement-based products is determined by physical, social,
 266 and economic factors (Huuhka & Lahdensivu, 2014; Liu et al., 2014; O'Connor, 2004;
 267 Thomsen & van der Flier, 2011). For the ***Short Lifetime*** scenario, we assumed that the
 268 lifetime remains unchanged during 2020-2060. For the ***Moderate Lifetime Extension***
 269 scenario, we assumed that the lifetime is extended to 50 years during 2020-2060. For the
 270 ***Substantial Lifetime Extension*** scenario, we assumed that the lifetime is extended to 70
 271 years during 2020-2060, as land use rights in China are usually granted for 70
 272 years(1990).

273 **Table 1 | Scenarios and key assumptions.**

Dimension	Scenario	Description
Lifetime	Short Lifetime (SL)	Lifetime remains unchanged during 2020-2060.
	Moderate Lifetime (ML)	Lifetime is extended to 50 years by 2060.
	Substantial Lifetime (LL)	Lifetime is extended to 70 years by 2060.
	High Stock (HS)	Cement stock per capita will grow on historical rates and saturate at 57.6 t/cap by 2060.

Medium Stock (MS)	Cement stock per capita will saturate at 41.3 t/cap by 2060.
Low Stock (LS)	Cement stock per capita will saturate at 33.5 t/cap by 2060.

274

275 For population projections, we adopted the medium growth scenario from United Nations
276 Population Prospects 2022, and the high- and low-variants to characterize the upper and
277 lower bounds(2022).

278 Under each of the nine demand-side scenarios, we considered five supply-side measures
279 to represent decarbonization efforts targeting cement production: (1) thermal efficiency,
280 (2) electrical efficiency, (3) alternative fuels, (4) clinker substitution, and (5) carbon
281 capture and storage (CCS). We collected data from various sources to determine present-
282 day values and values achievable by 2060. More specifically, we retrieved data on
283 thermal and electric efficiency from recent surveys and industrial statistics (2022). We
284 assumed that energy efficiency will improve to the current best levels by 2060 (2018;
285 2021). The share of alternative fuels will increase to 35% by 2050 and 40% by 2060
286 (Dinga & Wen, 2022). As recent policies promote high-grade cement with higher clinker
287 content than before (Wu et al., 2022), the clinker-to-cement ratio (CCR) is expected to
288 gradually decrease to 60% by 2060. Besides, we also considered grid decarbonization,
289 which is reflected by decreases in CO₂ emissions per unit of electricity used due to
290 increasing penetration of renewables (Tan et al., 2022). According to technical maturity
291 and economic costs, we considered the most mature post-combustion technology, i.e.,
292 absorption with mono-ethanolamine (MEA), as well as oxyfuel (OXY) and calcium
293 looping (CAL) technologies (Gardarsdottir et al., 2019; Voldsgaard et al., 2019). More
294 specifically, 88-93% of CO₂ from stack flue gas could be captured by these technologies,
295 with varying energy penalties (Voldsgaard et al., 2019). According to technology readiness

296 (Hills et al., 2016), mature MEA facilities have been applied to retrofit existing plants,
 297 while OXY and CAL are expected to be commercially available for new kilns by 2030
 298 (Hills et al., 2016), respectively. It is expected that around 100 kilns will be equipped
 299 with CCS facilities by 2050, based on the current pilot projects and optimistic
 300 anticipation (ECRA & CSI, 2017; GCCA, 2021). The deployment pace of CCS was
 301 determined by a capacity turnover model (see Supplementary Notes B), in which kilns
 302 with a capacity over 5000 tpd and remain lifetime within 20 years are prioritized for CCS
 303 deployment to maximize its utility and revenue.

304 **Table 2 | Supply-side mitigation measures.**

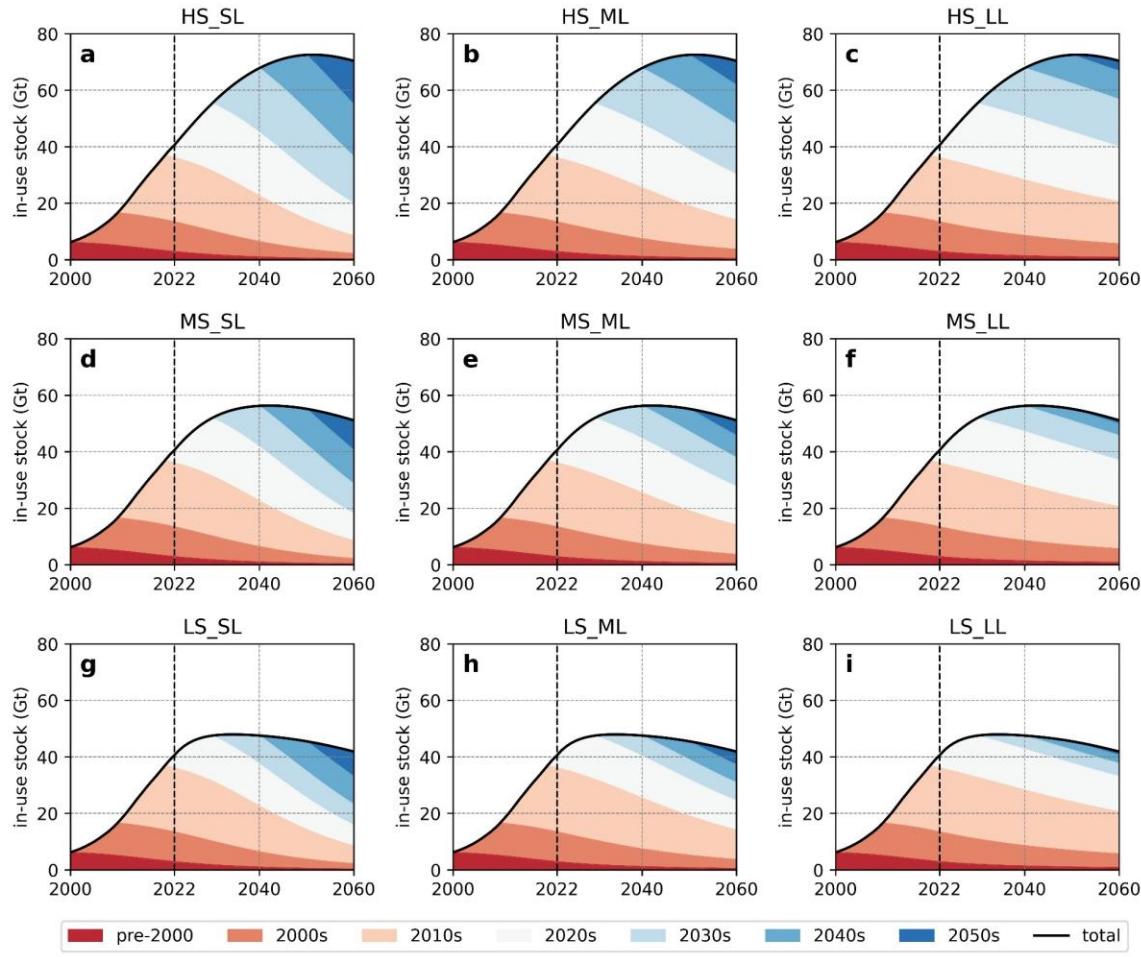
Measures	Description
Thermal efficiency	Thermal efficiency decreases to 3.0 and 2.9 GJ/t clinker by 2050 and 2060, respectively.
Electrical efficiency	Electrical efficiency for clinker production decreases to 57 and 50 kWh/t clinker by 2050 and 2060, respectively. Electrical efficiency for cement grinding decreases to 30 and 28 kWh/t cement by 2050 and 2060, respectively.
Alternative fuels	The share of alternative fuels increases to 35% by 2050 and 40% by 2060, respectively.
Clinker substitution	The clinker-to-cement (CCR) ratio decreases to 60% by 2060.
Carbon capture and storage (CCS)	Around 100 kilns are equipped with CCS by 2050, and five pilot kilns are retrofitted with MEA for demonstration by 2030. CAP and MAL are applicable for new kilns from 2030 and Oxyfuel and CAL are applicable for new kilns from 2040.
Grid decarbonization	CO ₂ emissions per unit of electricity used decrease to 0.07 and 0.03 kg CO ₂ /kWh by 2050 and 2060, respectively.

305
 306 For each supply-side measure, we assumed a linear trend between now and 2060 for
 307 simplification. More details about the methods and data are documented in the
 308 Supplementary Information.

309 **3. Results**

310 **3.1 Trends of cement stocks**

311 As shown in Figure 3, China's in-use cement stock has increased from 6.2 Gt to 40.5 Gt
312 during 2000-2022. Under Medium Stock scenarios (Figures 3d, 3e, and 3f), the in-use
313 stock will peak at 56.3 Gt by 2040 and then decrease to 51.1 Gt by 2060, if in-use stock
314 per capita will continue to rise based on the historical rate. As economic development and
315 urbanization often lead to lower fertility rates, population declines are expected to arrive
316 in the 2030s (Ye, 2022), leading to a slight decline in in-use stock after the peak. If
317 China's service level will catch up with the levels in industrialized countries, the in-use
318 stock will grow substantially and finally peak at 72.5 Gt by 2050, after that gradually
319 decline to 70.4 Gt in 2060 (High Stock scenarios, Figure 3a, 3b, 3c). In contrast, the in-
320 use stock will only increase by 17 % and peak at 47.5 Gt as early as 2030, with the
321 implementation of material efficiency measures (Low Stock scenarios, Figures 3g, 3h, 3i).
322 Across all scenarios, lifetime extension will significantly change the vintage proportion
323 of the in-use stock. Without lifetime extension, in-use cement stock built before 2020 will
324 gradually decrease to 22.7 Gt in 2040 and finally slide to 8.6 Gt in 2060, whereas in-use
325 stock built after 2020 will amount to as high as 45.2 Gt in 2040 and 61.8 Gt in 2060
326 (High Stock and Short Lifetime, Figure 3a). If the lifetime is extended to 70 years, in-use
327 stock built before 2020 will survive longer and gradually slide to 28.3 Gt in 2040 and
328 20.6 Gt in 2060 at a slower rate, and in-use stock built after 2020 could be reduced to as
329 low as 19.2 Gt in 2040 and 21.3 Gt in 2060 (Low Stock and Substantial Lifetime
330 Extension, Figure 3i).

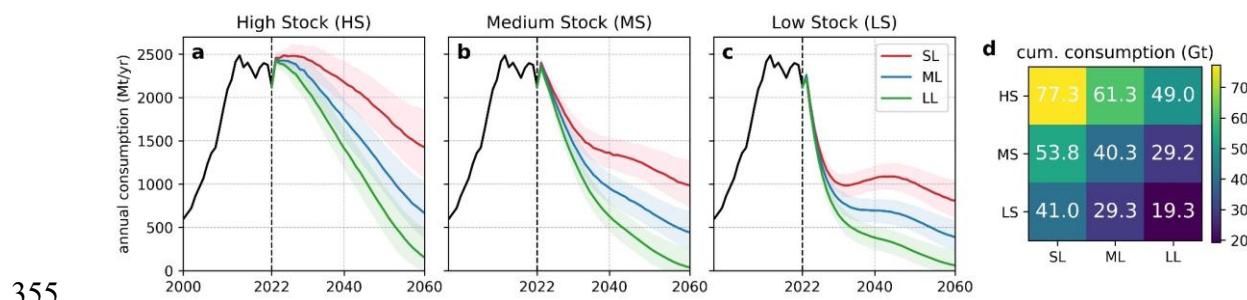


331
332 **Figure 3 | Projection of China's in-use cement stock of all scenarios during 2000-2060.**
333

334 3.2 Trends of cement consumption in China

335 As shown in Figure 4, the gradual saturation of cement stocks leads to declines in cement
336 consumption during 2023-2060. Across all scenarios, China's cement consumption is
337 projected to decrease over the next decades due to the combined effect of a declining
338 population and saturated per-capita cement stocks. This effect could gradually decrease
339 annual cement consumption from 2.1 Gt yr⁻¹ in 2022 to 1.7 Gt yr⁻¹ in 2030 if per-capita
340 stock grows at the historical rate (Medium Stock and Short lifetime, Figure 4b). In High
341 Stock scenarios, annual consumption will rebound to around 2.2 Gt yr⁻¹ in the 2030s after
342 a slight decline in the near term, when much more cement is needed to build up and

343 renovate buildings and infrastructure during 2023-2060 (Figure 4a). Conversely, a quick
 344 peaked per-capita stock by material efficiency improvement will constrain the annual
 345 consumption, leading to a quick and sharp decline to 1.0 Gt yr⁻¹ by 2030, after which
 346 annual consumption will fluctuate and eventually slide to 0.8 Gt yr⁻¹ in 2060 (Low Stock
 347 scenarios, Figure 4c). Lifetime extension can bend the curve of annual cement
 348 consumption across all stock-related scenarios. If the lifetime is extended to 50 and 70
 349 years, China's annual cement consumption will substantially decrease to 387-667 Mt yr⁻¹
 350 and 38-153 Mt yr⁻¹ by 2060, respectively. During 2023-2060, these savings could
 351 cumulatively amount to 21.3 Gt in the Low Stock scenario, 24.6 Gt in the Medium Stock
 352 scenario, and 28.3 Gt in the High Stock scenario. If combined, a lower per-capita stock
 353 and an extended lifetime can bring substantial cement savings, totaling 58 Gt in this
 354 period.



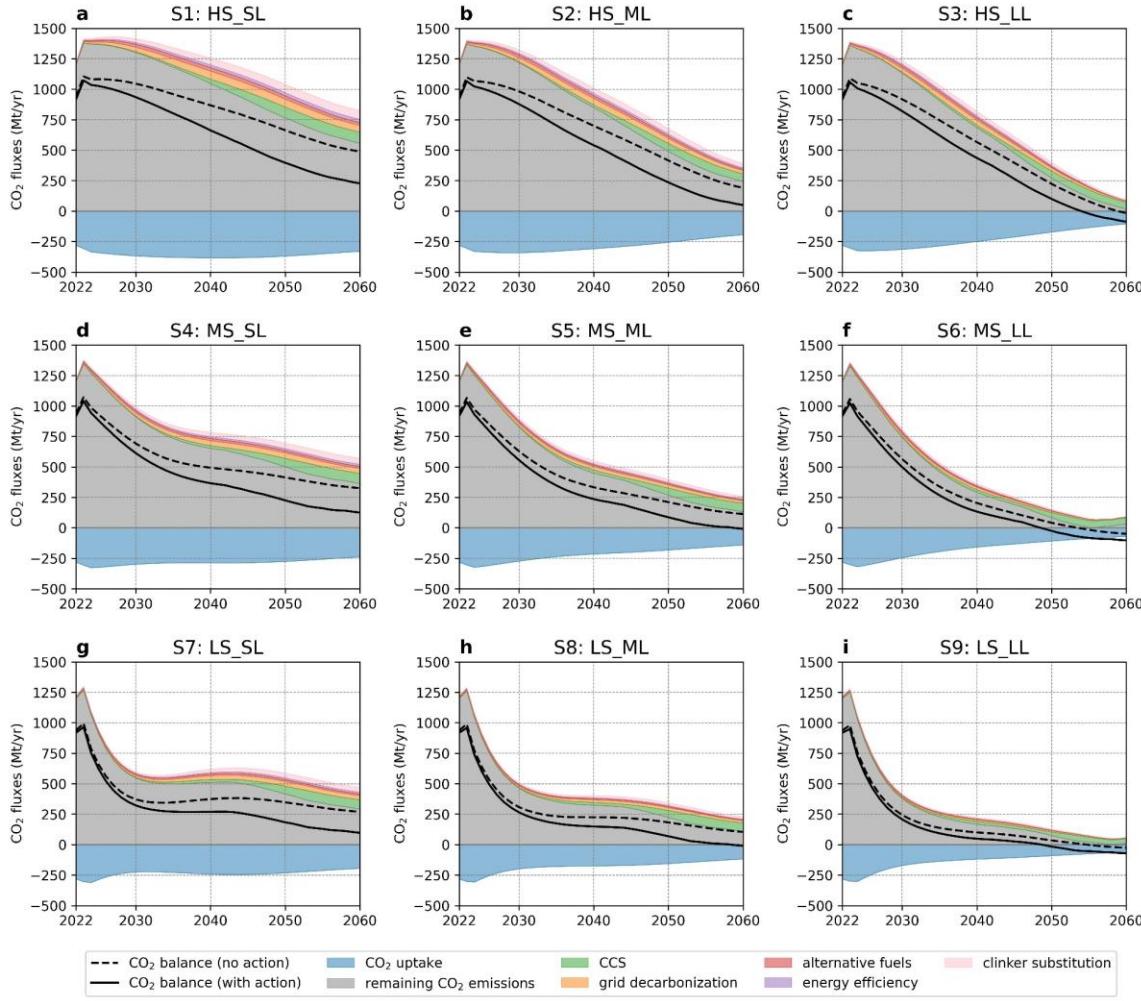
355
 356 **Figure 4 | China's annual cement consumption in the High Stock (a), Medium Stock (b), and Low**
 357 **Stock (c) scenarios in 2000-2060, and the cumulative cement consumption during 2023-2060 of all**
 358 **scenarios (d). Notes: The shaded area represents the simulated results using the high- and low-variant**
 359 **population projections. And SL, ML and LL represent Short Lifetime, Moderate Lifetime Extension,**
 360 **and Substantial Lifetime extension scenarios respectively.**

361

362 3.3 Decarbonization pathways of China's cement cycle

363 Future cement consumption sets important boundary conditions for decarbonizing the
 364 cement cycle. As shown in Figure 5, trends of annual CO₂ emissions are largely
 365 consistent with annual cement consumption across all scenarios. When no supply-side

measures are taken, annual CO₂ emissions will gradually decline to 817 Mt yr⁻¹, 564 Mt yr⁻¹, and 464 Mt yr⁻¹ in 2060, under the High Stock, Medium Stock, and Low Stock scenarios. In parallel, China will see substantial CO₂ uptake by cement-based products, gradually decreasing from ~282 Mt CO₂ yr⁻¹ in 2022 to ~194-329 Mt yr⁻¹ in 2060. When CO₂ uptake is considered but no supply-side measures are taken, the CO₂ balance is projected to decline to 488 Mt yr⁻¹ (High Stock and Short Lifetime), 326 Mt yr⁻¹ (Medium Stock and Short Lifetime), and 269 Mt yr⁻¹ (Low Stock and Short Lifetime) by 2060. If the potential of supply-side measures is fully seized, the CO₂ balance is projected to decrease to 228 Mt yr⁻¹ (High Stock and Short Lifetime), 125 Mt yr⁻¹ (Medium Stock and Short Lifetime), and 95 Mt yr⁻¹ (Low Stock and Short Lifetime) by 2060. Among the five supply-side measures, CCS can bring substantial CO₂ savings, ranging from 80 Mt yr⁻¹ to 91 Mt yr⁻¹ by 2060. The second largest CO₂ saver is clinker substitution, which can deliver CO₂ reductions ranging from 41 to 72 Mt CO₂ yr⁻¹ in 2060.

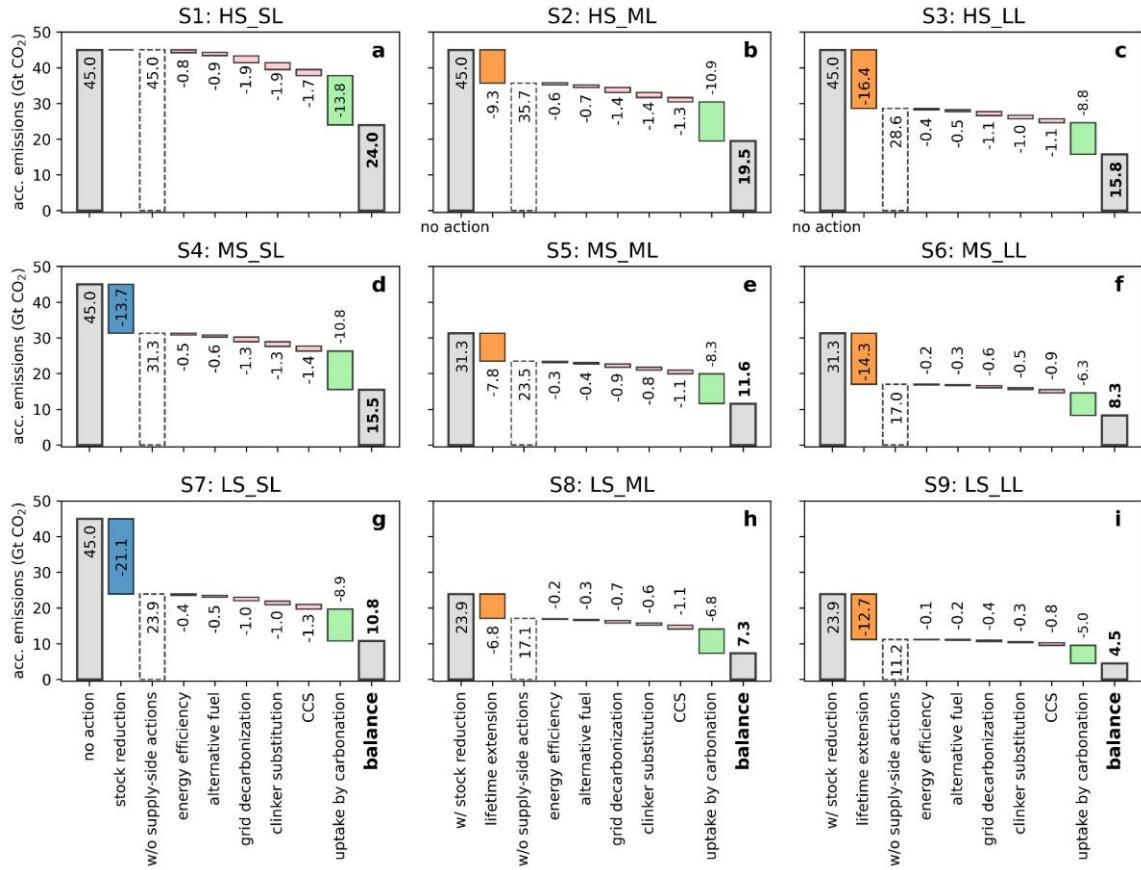


379
380 **Figure 5 | Annual CO₂ fluxes associated with the cement cycle during 2022-2060.**
381

382 If the lifetime is extended to 50 years, no-action CO₂ emissions will decline to 384 Mt yr⁻¹ (High Stock and Moderate Lifetime Extension), 254 Mt yr⁻¹ (Medium Stock and Moderate Lifetime Extension), and 222 Mt yr⁻¹ (Low Stock and Moderate Lifetime Extension) by 2060. While lifetime extension can deliver emission cuts, China will see considerable reductions in CO₂ uptake, as less CO₂ uptake will occur due to reduced cement consumption. When CO₂ uptake is considered and supply-side measures are taken, the CO₂ balance is projected to decline to 47 Mt yr⁻¹ (High Stock and Moderate Lifetime Extension), -7 Mt yr⁻¹ (Medium Stock and Moderate Lifetime Extension), and -11 Mt yr⁻¹

390 (Low Stock and Moderate Lifetime Extension) by 2060, clearly indicating that these
391 demand-side strategies could enable a net-zero cement cycle by 2060 in China. If the
392 lifetime is further extended to 70 years, the CO₂ balance is projected to reach zero by
393 2050 in Medium Stock and Substantial Lifetime Extension and Low Stock and
394 Substantial Lifetime Extension scenarios, leading to a negative-emission cement cycle
395 after 2050.

396 As shown in Figure 6, lower per-capita stocks could cumulatively reduce total CO₂
397 emissions by 13.7 Gt and 21.1 Gt in 2023-2060. Moderate and substantial lifetime
398 extension will cut total emissions by 6.8-9.3 Gt and 12.7-16.4 Gt, whereas these
399 reductions shrink with stock levels. And combining these two demand-side strategies
400 together could totally decrease the cumulative emissions by 33.8 Gt CO₂. Also, the
401 contribution of CO₂ uptake varies by scenario. Take two extreme scenarios as examples:
402 in the High Stock and Short Lifetime scenario, CO₂ uptake can offset ~31% of the no-
403 action CO₂ emissions, while in the Low Stock and Substantial Lifetime Extension
404 scenario, CO₂ uptake can offset ~45% of the CO₂ emissions without supply-side actions.
405 The varying contributions of CO₂ uptake in decarbonizing the cement cycle indicate that
406 CO₂ balance hinges on two boundary conditions: cement stock and lifetime. Cement
407 stock reduction and lifetime extension can slow down the turnover of cement stocks and
408 reduce the total societal throughput of cement, which in turn deliver emission cuts.
409 However, due to the legacy effects of historically consumed and accumulated cement,
410 CO₂ uptake is comparatively less affected by cement stock reduction and lifetime
411 extension. We therefore highlight that designing decarbonization roadmaps for the
412 cement cycle must consider these two critical boundary conditions.



413
414
415

Figure 6 | Cumulative CO₂ emissions and uptake associated with China's cement cycle during 2023-2060. (a)-(i) cumulative CO₂ emissions reductions, uptake, and CO₂ balance of scenarios S1-S9.

416 **4. Discussion**

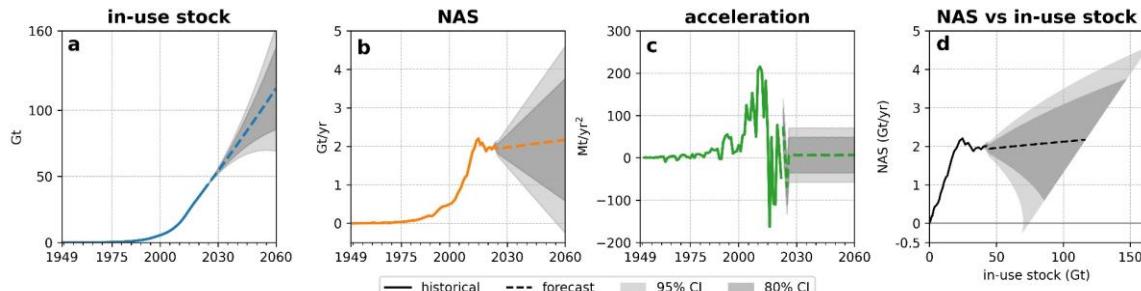
417 **4.1 Robustness of stock saturation assumption**

418 In our scenario analysis, we adopted the ‘stock saturation’ hypothesis, which has been
419 supported by previous studies (Cao et al., 2020; Krausmann et al., 2020; Liu et al., 2012).

420 According to this hypothesis, a convergence of per-capita material stocks is anticipated in
421 the future, following an S-shaped system transition between socio-metabolic regimes
422 (Wiedenhofer et al., 2021). Although empirical evidence of this saturation pattern exists
423 in a few industrialized countries (Cao et al., 2017), we further conducted a forecast using
424 an Autoregressive Integrated Moving Average (ARIMA) model (Fishman et al., 2016) to
425 examine the robustness of the ‘stock saturation’ assumption in this study. The historical
426 acceleration extrapolated constant positive values, suggesting an increasing net addition
427 to stock (NAS) and consequently, a continued growth of the in-use stock (Figure 7).

428 Additionally, Figure 7d reveals that points with zero NAS lie on the edge of the 95%
429 confidence interval of the forecast, which could be interpreted as an indication of ‘stock
430 saturation’ (Wiedenhofer et al., 2021). Presently, the signal of stock saturation is weak as
431 these saturation points appear less likely to occur, but it is expected to strengthen if
432 cement consumption continues to decline in the coming years. Despite the current weak
433 signal, we have chosen to adopt the saturation hypothesis due to its alignment with the
434 developmental trajectories observed in many industrialized countries, as well as China’s
435 emphasis on high-quality development today (Xinhua New Agency, 2023). The concept
436 of material stocks saturation implies that China will achieve a ‘decoupling’ of economic
437 development from resource use and transition towards a more service and consumption-

438 based economy in the future, following a phase of rapid growth that has focused on
 439 essential infrastructure and capital stock formation (Bleischwitz et al., 2018).



440
 441 **Figure 7 | ARIMA forecast on the level (a), NAS (b), and acceleration (c) based on the historical**
 442 **trend of China's cement stock in 1949-2060, as well as the trajectory of stock to NAS ratio (d).** 95%
 443 CI and 80% CI refers to 95% and 80% confidence intervals of forecast respectively. NAS is the annual Net
 444 Addition to Stock, which is equivalent to inflows minus outflows in a certain year, whereas acceleration is
 445 the change in NAS between two consecutive years, or is the second order of difference of in-use stock
 446 (Fishman et al., 2016). The ARIMA model results are displayed in Supporting Information.
 447

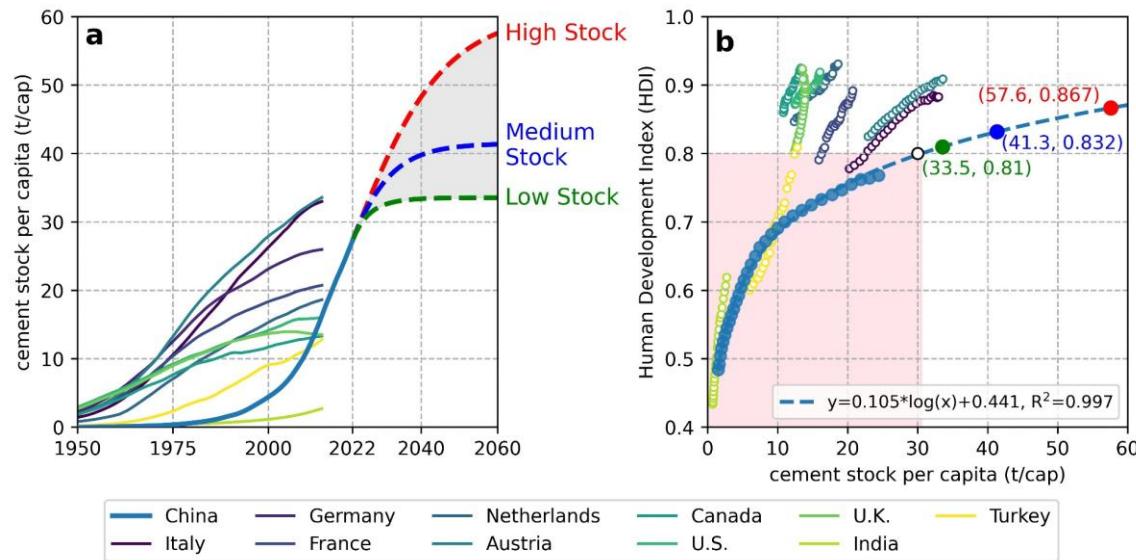
448 **4.2 Implications of a lower per-capita stock**

449 A lower per capita cement stock, achieved through more intensive use, has the potential
 450 to significantly reduce future cement demand and CO₂ emissions. However, it should be
 451 noted that this reduction primarily implies a lower per capita service level rather than a
 452 decrease in cement intensity of end-use products, which cannot be easily achieved
 453 through dematerialization as seen in the past upgrades of smartphones. This raises
 454 important social implications, as essential services would need to be provided with
 455 significantly lower levels of material stocks. Several measures can be taken in the
 456 building sectors to limit per-capita floor area, including smaller dwellings, more shared
 457 spaces, increased multi-family housing (Berrill et al., 2022; Creutzig et al., 2021;
 458 Hertwich et al., 2019), while a more compact city, characterized by higher urban density,
 459 can lead to reduced infrastructure requirements per individual by more efficient
 460 transportation networks and mixed land use (Muller et al., 2013). However, it remains
 461 uncertain whether these strategies can lead to the levels outlined in Low Energy Demand

462 (LED) scenarios (Fishman et al., 2021; Grubler et al., 2018). In the United States, there is
463 potential to reduce per-capita residential floor space by 15% from 69.2 m²/cap to 58.9
464 m²/cap by 2060 through a combination of reducing the size of new single-family houses
465 and increasing the proportion of multi-family houses (Berrill & Hertwich, 2021). In
466 comparison, China appears to utilize floor space more efficiently, with its average
467 residential area reaching a lower level of approximately 41 m²/cap by 2020. Efficiency in
468 the use of floor space also varies between urban and rural areas in China. Affordability
469 issues in high-priced urban areas may limit per capita floor space, while rural areas have
470 the potential for residents to economically build multi-story houses on their own parcels.
471 Additionally, rural-to-urban migration leads to vacant houses for a significant portion of
472 the year, resulting in wasted floor space.

473 A transition towards lower stock necessitates behavioral and lifestyle changes, along
474 with considerations of architectural design preferences. The underlying challenge lies in
475 achieving a reduced stock without compromising societal development and quality of life.
476 In China, the current per capita cement stock surpasses that of many countries due to its
477 historical reliance on cement-intensive building and infrastructure construction (Figure
478 8a). We have found a significant association between China's Human Development
479 Index (HDI) and its per capita cement stock from 1990 to 2021 (Figure 8b). Extrapolating
480 the trend reveals that while all stock scenarios would result in a higher HDI than the
481 present, a lower per capita stock may constrain future HDI growth unless China
482 successfully decouples its development from cement stock. Furthermore, well-being
483 concerns should be carefully addressed, as it appears that limiting shelter services based

484 on ‘sufficiency’ considerations may contribute little to overall well-being, whereas other
 485 demand-side solutions align with high levels of well-being (Creutzig et al., 2021).



486
 487 **Figure 8 | Comparison of China’s future per-capita cement stock scenarios with other countries (a),**
 488 **and the relationship between per-capita stock and Human Development Index (HDI) (b).** The colors
 489 of the circles in (b) are the same as those of the lines in (a). The data for high-income countries are obtained
 490 from Cao et al. (2017). HDI data are collected from [Home | Human Development Reports \(undp.org\)](#).

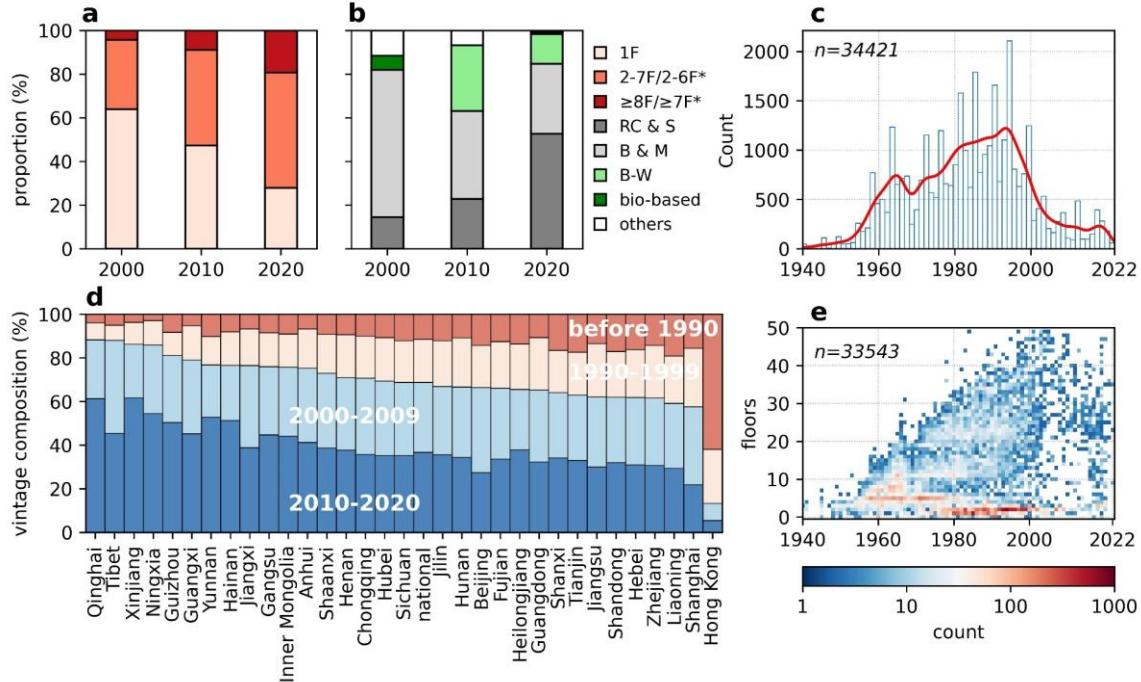
491

492 **4.3 Opportunities and challenges of lifetime extension**

493 Both field investigations and macro-level analyses consistently confirm a significantly
 494 shorter building lifespan in China compared to the designed service-life of 50 years (Cai
 495 et al., 2015; Cao et al., 2019; Liu et al., 2014). Empirical studies have shown that
 496 premature demolitions in China are primarily attributed to external factors rather than
 497 internal degradation of building materials (Liu et al., 2014; O'Connor, 2004). This
 498 phenomenon of “short-lived” buildings in China ultimately reflects a premature
 499 “obsolescence” resulting from rapid urbanization over the past two decades (Thomsen &
 500 van der Flier, 2011). The discrepancy between rising expectations driven by increasing
 501 land values and the actual function and quality of buildings has led to premature
 502 demolitions. However, it is worth noting that this pattern may differ in cases where

503 buildings are situated on valueless land, where out-of-service buildings tend to persist
504 without undergoing demolition (Thomsen & van der Flier, 2011).

505 There have been some improvements observed, as the average building lifespan in
506 China has increased by approximately 0.5 years between 2000 and 2015 (Cao et al.,
507 2019). This extension may be indirectly validated by structural changes in China's
508 building stocks. Over the past two decades, new buildings have become taller and more
509 concrete-intensive than before (Figure 9a-b), suggesting potentially longer service-life
510 due to limitations imposed by China's property rights system. Specifically, high-rise and
511 high-density residential districts, accommodating hundreds to thousands of households in
512 a small piece of land, present daunting challenges for demolition and redevelopment
513 when multiple property rights are involved. In mainland China, most building stocks are
514 within 30 years old, with over 60% constructed during the period from 2000 to 2020. In
515 contrast, the proportion of such buildings is relatively low in Hong Kong Special
516 Administrative Region (SAR) (Figure 9d). Hong Kong experienced a construction boom
517 between the 1960s and 1980s, similar to industrialized countries in the post-war era, and
518 buildings from that period are still in service today (Figure 9c). Additionally, many high-
519 rise buildings in Hong Kong have survived for decades (Figure 9e), potentially
520 supporting the argument that high-rise buildings tend to have longer service-life due to
521 the challenges associated with redevelopment. Therefore, the average building lifespan is
522 expected to increase overall, but the extent of this increase may depend on various
523 practical factors.



524
 525 **Figure 9 | Current composition of China's building stocks.** (a) Residential building stock composition
 526 by building floor. (b) Residential building stock composition by building material. (c) Year built of
 527 private buildings in Hong Kong Special Administrative Region (SAR). (d) Vintage composition of
 528 residential buildings by province as well as HKSAR. (e) Building floors versus year built in HKSAR.
 529 Notes: In (a), “2-6F” and “≥7F” are for 2000 and 2010, whereas “2-7F” and “≥8F” are for 2000 and 2010.
 530 RC refers to reinforced concrete structure, S refers to steel structures, B and M refers to brick and masonry
 531 structures, B-W denotes brick-wood structures, and bio-based mainly refers to structures constructed by
 532 bio-based materials. Source: Data for China's provinces are retrieved from (the 7th National Population
 533 Census in 2020), whereas the data for Hong Kong were collected from *Database on Private Buildings in*
 534 *Hong Kong* of the Home Affairs Department of HKSAR (2023).

535

536 While it is expected that high-rise building stocks will survive for a longer time,
 537 addressing the obsolescence of remaining aged low-rise buildings poses challenges in
 538 terms of regular maintenance, proper retrofitting, and adaptive reuse, especially when
 539 redevelopment appears to be the more economically viable choice given the ongoing
 540 urbanization in China. In Hong Kong, the presence of aged buildings also raises social
 541 concerns, particularly related to safety issues. The reconstruction of these aged and
 542 dilapidated buildings, located on valuable land, is considered a more economically viable
 543 solution compared to renovation and reuse. Consequently, there is a growing emphasis on
 544 ensuring a successful redevelopment process, which may involve implementing measures

545 such as compulsory sales of the land on which the aged buildings are situated.
546 Furthermore, financial considerations also play a significant role. The inflow of more
547 vulnerable residents into the aged buildings can severely hamper maintenance
548 investments (Thomsen & van der Flier, 2011).

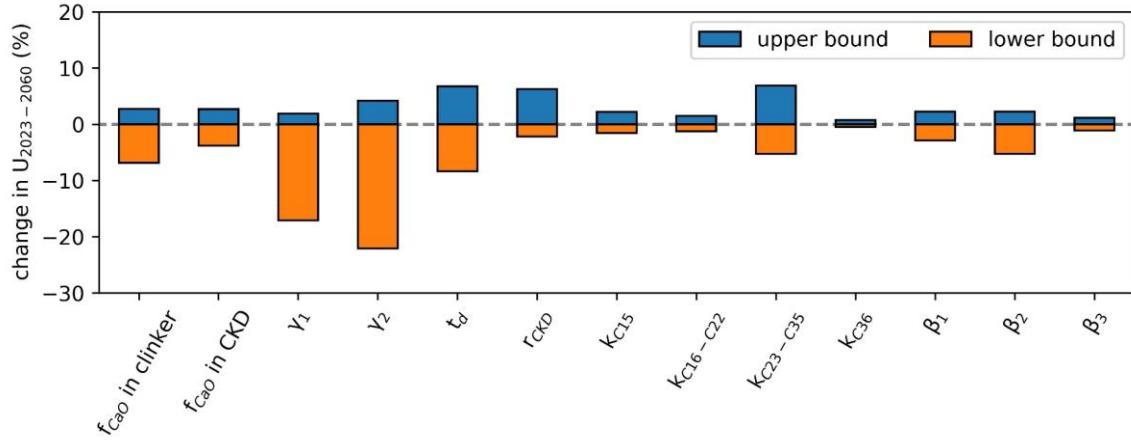
549 From a technical perspective, the extension of service-life may give rise to side-effects
550 that need to be carefully considered in practice. Durable advanced cement-based
551 materials typically have higher greenhouse gas (GHG) emissions during production
552 compared to those with lower durability (Miller, 2020). When the target service-life is
553 significantly shorter than the physical life of durable concrete mixtures, it becomes less
554 desirable due to higher GHG emissions, leading to a waste in durability performance.
555 Another aspect to consider is the decline in performance as cement stocks age. This
556 decline may result in increased heating and cooling loads due to reduced thermal
557 insulation performance, necessitating additional energy for operation (Grussing, 2014).
558 This technological obsolescence may require future energy retrofitting. Moreover, for
559 road infrastructure, changes in cement-based material properties due to aging, such as
560 surface roughness, can lead to increased fuel consumption and emissions from vehicle
561 use (Reger et al., 2014). These considerations also apply to young high-rise buildings
562 currently in existence, as they will undergo aging and face obsolescence in the future.

563 **4.4 Trade-off in CO₂ uptake**

564 Notwithstanding that stock reduction and lifetime extension could unequivocally deliver
565 emissions cut by decreasing cement production, both will inevitably shrink passive CO₂
566 sequestration by cement carbonation because of reduced in-use stock, inflows, and
567 demolished outflows (Figure S18). Specifically, CO₂ uptake by in-use concrete structures

568 will be almost halved during 2023-2060 in the Low Stock-Substantial Lifetime Extension
569 scenario (1.3 Gt) compared with that in the High Stock-Short Lifetime scenario (2.3 Gt).
570 Those materials embodied in new construction activities and cement production will
571 decrease significantly in line with the downward trend of cement inflows, especially for
572 rendering mortar, which will be fully carbonated and reabsorb substantial CO₂ within the
573 first few years after entering use. Besides, avoided outflow by lifetime extension will
574 significantly shrink CO₂ uptake by demolished concrete structures, as crushed concrete
575 debris could dramatically increase the surface area exposed to air, leading to a further
576 uptake within the short demolition stage. Only CO₂ uptake by masonry mortar will
577 slightly increase (by ~0.6 Gt) by lifetime extension because it ensures enough time for
578 CO₂ to penetrate through the thick surface rendering layer to reach the masonry mortar
579 layer and induce its carbonation.

580 Additionally, a sensitivity analysis was conducted to investigate the impacts of
581 parameters' ranges on the future CO₂ uptake (details seem Supporting Information 6). As
582 shown in Figure 10 below, total CO₂ uptake during 2023-2060 will rise with an increase
583 in the CaO contents in clinker and cement kiln dust (CKD) as well as the carbonation
584 rates of concrete mixture. And when the proportion of CaO that could be converted to
585 CaCO₃ within the fully carbonated cement and mortar (γ_1 and γ_2) decrease to their lower
586 bounds, total uptake will significantly shrink by 17% and 22%. Further, decreases in
587 corrector coefficients that reflect the impacts of surrounding CO₂ concentrations, coating
588 and covers, and cement additives (β_1 , β_2 , and β_3) will reduce the total uptake
589 approximately by -5%.



590
591 **Figure 10 | Impacts of parameter changes on total CO₂ uptake by cement carbonation during 2023-
592 2060.**

593
594 **4.5 Policy recommendations**

595 Effective implementation of demand-side strategies holds significant potential for
596 reducing future demand and greenhouse gas (GHG) emissions through the efficient
597 utilization of material stocks. In this regard, governments play a crucial role in directing
598 the attention towards demand-side measures. Encouraging a more compact layout of
599 buildings, which can be specified in industry guidance or building codes, is a key
600 approach to maximizing the efficient use of building floor area. The current shift in
601 architectural design within the real estate market aligns with this objective, as new
602 apartments are being designed to accommodate more rooms without a substantial
603 increase in total floor area. This trend enables the accommodation of larger families while
604 maintaining housing affordability. Additionally, it is pertinent to initiate a discourse on
605 the concept of “sufficiency” in terms of per-capita building floor area (Cohen, 2020),
606 particularly in light of climate ambitions.

607 Urban planners also play a vital role in extending the lifespan of buildings and
608 mitigating premature demolitions. They should consider developing new or revising

609 existing building zone requirements that facilitate mixed uses and repurposing of existing
610 structures, thereby extending their longevity (Zhong et al., 2021). Where feasible, urban
611 sprawl with a premise of minimizing adverse impacts on cropland and forest areas, could
612 be prioritized over frequent urban sprawl, which often leads to premature demolition. It is
613 essential to note that ensuring safety is a fundamental requirement for the prolonged
614 utilization of buildings and infrastructure. To achieve this, mainland China should
615 establish a robust legal and regulatory system that includes regular inspections of aged
616 building structures and components. Valuable insights can be gained from Hong Kong's
617 implementation of the Mandatory Inspection on Building and Mandatory Inspection on
618 Windows regulations.

619 Furthermore, the cement industry faces challenges in promoting efficient material use
620 due to its profit model, which is inherently linked to the volume of materials sold (Watari
621 et al., 2022). Thus, incentives and subsidies may be necessary to encourage the industry
622 to adopt more sustainable practices. Moreover, the issue of over-capacity in China's
623 cement sector may be further exacerbated by a decline in cement demand. Underutilized
624 capacity poses a significant risk to newly constructed and costly cement kilns, as annual
625 cement demand is projected to remain below two billion tons based on the scenarios
626 analysis in this study (Lu et al., 2022).

627 ***5. Conclusions***

628 This study developed an integrated modeling framework of the material-energy-
629 emission-uptake nexus in the cement cycle (IMAGINE Cement) to quantify the emission
630 mitigation potentials from demand-side measures. And the message of this study is clear:
631 net-zero emissions across China's cement cycle are achievable by 2060 through more
632 intensive use and lifetime extension of in-use stock, in conjunction with a moderate
633 portfolio of supply-side measures. Specifically, combined low demand and an extended
634 lifetime of 70 years could save up to 58 Gt cement demand during 2023-2060, which in
635 turn avoids 33.8 Gt CO₂ production emissions in this period. Also, trade-offs between
636 emission fluxes should be carefully considered in the decarbonization pathways because
637 more intensive use and lifetime extension will compromise the CO₂ uptake by cement
638 carbonation. However, besides these substantial emissions reductions, it is necessary to
639 pay attention on the implications, opportunities, and challenges of lower per-capita stocks
640 and longer service-life of cement stocks, as these changes have a profound impact on
641 social development and quality of life in the future.

642

643 ***Data availability***

644 Part of data used for analysis in this study are available in a repository online in
645 accordance with funder data retention policies (<https://doi.org/10.5281/zenodo.8083355>).

646 ***Acknowledgments***

647 We thank anonymous reviewers for their valuable comments and suggestions.

References

- Andrew, R. M. (2019). Global CO₂ emissions from cement production, 1928–2018. *Earth System Science Data*, 11(4), 1675-1710. <https://doi.org/10.5194/essd-11-1675-2019>
- B. Müller, D. (2006). Stock dynamics for forecasting material flows—Case study for housing in The Netherlands. *Ecological Economics*, 59(1), 142-156. <https://doi.org/10.1016/j.ecolecon.2005.09.025>
- Bergsdal, H., Brattebø, H., Bohne, R. A., & Müller, D. B. (2007). Dynamic material flow analysis for Norway's dwelling stock. *Building Research & Information*, 35(5), 557-570. <https://doi.org/10.1080/09613210701287588>
- Berrill, P., & Hertwich, E. G. (2021). Material flows and GHG emissions from housing stock evolution in US counties, 2020–60. *Buildings and Cities*, 2(1), 599-617. <https://doi.org/10.5334/bc.126>
- Berrill, P., Wilson, E. J. H., Reyna, J. L., Fontanini, A. D., & Hertwich, E. G. (2022). Decarbonization pathways for the residential sector in the United States. *Nature Climate Change*, 12(8), 712-718. <https://doi.org/10.1038/s41558-022-01429-y>
- Bleischwitz, R., Nechifor, V., Winning, M., Huang, B., & Geng, Y. (2018). Extrapolation or saturation – Revisiting growth patterns, development stages and decoupling. *Global Environmental Change*, 48, 86-96. <https://doi.org/10.1016/j.gloenvcha.2017.11.008>
- Cai, W., Wan, L., Jiang, Y., Wang, C., & Lin, L. (2015). Short-Lived Buildings in China: Impacts on Water, Energy, and Carbon Emissions. *Environ Sci Technol*, 49(24), 13921-13928. <https://doi.org/10.1021/acs.est.5b02333>
- Cao, Z., Liu, G., Duan, H., Xi, F., Liu, G., & Yang, W. (2019). Unravelling the mystery of Chinese building lifetime: A calibration and verification based on dynamic material flow analysis. *Applied Energy*, 238, 442-452. <https://doi.org/10.1016/j.apenergy.2019.01.106>
- Cao, Z., Masanet, E., Tiwari, A., & Akolawala, S. (2021). *Decarbonizing Concrete: Deep decarbonization pathways for the cement and concrete cycle in the United States, India, and China*.
- Cao, Z., Myers, R. J., Lupton, R. C., Duan, H., Sacchi, R., Zhou, N., Reed Miller, T., Cullen, J. M., Ge, Q., & Liu, G. (2020). The sponge effect and carbon emission mitigation potentials of the global cement cycle. *Nat Commun*, 11(1), 3777. <https://doi.org/10.1038/s41467-020-17583-w>
- Cao, Z., Shen, L., Lovik, A. N., Muller, D. B., & Liu, G. (2017). Elaborating the History of Our Cementing Societies: An in-Use Stock Perspective. *Environ Sci Technol*, 51(19), 11468-11475. <https://doi.org/10.1021/acs.est.7b03077>
- Churkina, G., Organschi, A., Reyer, C. P. O., Ruff, A., Vinke, K., Liu, Z., Reck, B. K., Graedel, T. E., & Schellnhuber, H. J. (2020). Buildings as a global carbon sink. *Nature Sustainability*, 3(4), 269-276. <https://doi.org/10.1038/s41893-019-0462-4>
- Cohen, M. J. (2020). New Conceptions of Sufficient Home Size in High-Income Countries: Are We Approaching a Sustainable Consumption Transition? *Housing, Theory and Society*, 38(2), 173-203. <https://doi.org/10.1080/14036096.2020.1722218>
- Creutzig, F., Niamir, L., Bai, X., Callaghan, M., Cullen, J., Díaz-José, J., Figueroa, M., Grubler, A., Lamb, W. F., Leip, A., Masanet, E., Mata, É., Mattauch, L., Minx, J. C., Mirasgedis, S., Mulugetta, Y., Nugroho, S. B., Pathak, M., Perkins, P., . . . Ürge-Vorsatz, D. (2021). Demand-side solutions to climate change mitigation consistent with high levels of well-being. *Nature Climate Change*, 12(1), 36-46. <https://doi.org/10.1038/s41558-021-01219-y>

- Cullen, J. M., & Cooper, D. R. (2022). Material Flows and Efficiency. *Annual Review of Materials Research*, 52(1), 525-559. <https://doi.org/10.1146/annurev-matsci-070218-125903>
- Department of Economic and Social Affairs of United Nations. (2022). *World Population Prospects 2022*. <https://population.un.org/wpp/>
- Dinga, C. D., & Wen, Z. (2022). China's green deal: Can China's cement industry achieve carbon neutral emissions by 2060? *Renewable and Sustainable Energy Reviews*, 155. <https://doi.org/10.1016/j.rser.2021.111931>
- Eberhardt, L. C. M., Birgisdóttir, H., & Birkved, M. (2018). Life cycle assessment of a Danish office building designed for disassembly. *Building Research & Information*, 47(6), 666-680. <https://doi.org/10.1080/09613218.2018.1517458>
- European Cement Research Academy, & Cement Sustainability Initiative of the World Business Council for Sustainable Development. (2017). *Technology Papers 2017 Development of State of the Art Techniques in Cement Manufacturing: Trying to Look Ahead*.
- Fennell, P., Driver, J., Bataille, C., & Davis, S. J. (2022). Going net zero for cement and steel. *Nature*, 603(7902), 574-577. <Go to ISI>://WOS:000772661800010
- Fennell, P. S., Davis, S. J., & Mohammed, A. (2021). Decarbonizing cement production. *Joule*, 5(6), 1305-1311. <https://doi.org/10.1016/j.joule.2021.04.011>
- Fishman, T., Heeren, N., Pauliuk, S., Berrill, P., Tu, Q., Wolfram, P., & Hertwich, E. G. (2021). A comprehensive set of global scenarios of housing, mobility, and material efficiency for material cycles and energy systems modeling. *Journal of Industrial Ecology*, 25(2), 305-320. <https://doi.org/10.1111/jiec.13122>
- Fishman, T., Schandl, H., & Tanikawa, H. (2016). Stochastic Analysis and Forecasts of the Patterns of Speed, Acceleration, and Levels of Material Stock Accumulation in Society. *Environ Sci Technol*, 50(7), 3729-3737. <https://doi.org/10.1021/acs.est.5b05790>
- Friedlingstein, P., O'Sullivan, M., Jones, M. W., Andrew, R. M., Gregor, L., Hauck, J., Le Quéré, C., Luijckx, I. T., Olsen, A., Peters, G. P., Peters, W., Pongratz, J., Schwingshackl, C., Sitch, S., Canadell, J. G., Ciais, P., Jackson, R. B., Alin, S. R., Alkama, R., . . . Zheng, B. (2022). Global Carbon Budget 2022. *Earth System Science Data*, 14(11), 4811-4900. <https://doi.org/10.5194/essd-14-4811-2022>
- Gardarsdottir, S., De Lena, E., Romano, M., Roussanaly, S., Voldsdund, M., Pérez-Calvo, J.-F., Berstad, D., Fu, C., Anantharaman, R., Sutter, D., Gazzani, M., Mazzotti, M., & Cinti, G. (2019). Comparison of Technologies for CO₂ Capture from Cement Production—Part 2: Cost Analysis. *Energies*, 12(3). <https://doi.org/10.3390/en12030542>
- Gastaldi, D., Canonico, F., Capelli, L., Buzzi, L., Boccaleri, E., & Irico, S. (2015). An investigation on the recycling of hydrated cement from concrete demolition waste. *Cement and Concrete Composites*, 61, 29-35. <https://doi.org/10.1016/j.cemconcomp.2015.04.010>
- Global Cement and Concrete Association. (2021). *The GCCA 2050 Cement and Concrete Industry Roadmap for Net Zero Concrete*.
- Göswein, V., Arehart, J. H., Pittau, F., Pomponi, F., Lamb, S., Zea Escamilla, E., Freire, F., Silversre, J. D., & Habert, G. (2022). Wood in buildings the right answer to the wrong question. IOP Conference Series: Earth and Environmental Science,
- Grubler, A., Wilson, C., Bento, N., Boza-Kiss, B., Krey, V., McCollum, D. L., Rao, N. D., Riahi, K., Rogelj, J., De Stercke, S., Cullen, J., Frank, S., Fricko, O., Guo, F., Gidden, M., Havlík, P., Huppmann, D., Kiese wetter, G., Rafaj, P., . . . Valin, H. (2018). A low energy demand scenario for meeting the 1.5 °C target and sustainable development goals without negative emission technologies. *Nature Energy*, 3(6), 515-527. <https://doi.org/10.1038/s41560-018-0172-6>

- Grussing, M. N. (2014). Life Cycle Asset Management Methodologies for Buildings. *Journal of Infrastructure Systems*, 20(1). [https://doi.org/10.1061/\(asce\)is.1943-555x.0000157](https://doi.org/10.1061/(asce)is.1943-555x.0000157)
- Guo, R., Wang, J., Bing, L., Tong, D., Ciais, P., Davis, S. J., Andrew, R. M., Xi, F., & Liu, Z. (2021). Global CO₂ uptake by cement from 1930 to 2019. *Earth System Science Data*, 13(4), 1791-1805. <https://doi.org/10.5194/essd-13-1791-2021>
- Habert, G., Miller, S. A., John, V. M., Provis, J. L., Favier, A., Horvath, A., & Scrivener, K. L. (2020). Environmental impacts and decarbonization strategies in the cement and concrete industries. *Nature Reviews Earth & Environment*, 1(11), 559-573. <https://doi.org/10.1038/s43017-020-0093-3>
- Hertwich, E. G., Ali, S., Ciacci, L., Fishman, T., Heeren, N., Masanet, E., Asghari, F. N., Olivetti, E., Pauliuk, S., Tu, Q., & Wolfram, P. (2019). Material efficiency strategies to reducing greenhouse gas emissions associated with buildings, vehicles, and electronics—a review. *Environmental Research Letters*, 14(4). <https://doi.org/10.1088/1748-9326/ab0fe3>
- Hills, T., Leeson, D., Florin, N., & Fennell, P. S. (2016). Carbon Capture in the Cement Industry: Technologies, Progress, and Retrofitting. *Environmental Science Technology*, 50(1), 368-377. <https://doi.org/10.1021/acs.est.5b03508>
- Home Affairs Department of HKSAR. (2023). *Database of Private Buildings in Hong Kong*. https://bmis2.buildingmgt.gov.hk/bd_hadbiex/home.jsf
- Huuhka, S., Kaasalainen, T., Hakanen, J. H., & Lahdensivu, J. (2015). Reusing concrete panels from buildings for building: Potential in Finnish 1970s mass housing. *Resources, Conservation and Recycling*, 101, 105-121. <https://doi.org/10.1016/j.resconrec.2015.05.017>
- Huuhka, S., & Lahdensivu, J. (2014). Statistical and geographical study on demolished buildings. *Building Research & Information*, 44(1), 73-96. <https://doi.org/10.1080/09613218.2014.980101>
- International Energy Agency, & Cement Sustainability Initiative of the World Business Council for Sustainable Development. (2018). *Technology Roadmap: Low Carbon Transition in the Cement Industry*.
- Kapur, A., Keoleian, G., Kendall, A., & Kesler, S. E. (2008). Dynamic Modeling of In-Use Cement Stocks in the United States. *Journal of Industrial Ecology*, 12(4), 539-556. <https://doi.org/10.1111/j.1530-9290.2008.00055.x>
- Krausmann, F., Wiedenhofer, D., & Haberl, H. (2020). Growing stocks of buildings, infrastructures and machinery as key challenge for compliance with climate targets. *Global Environmental Change*, 61. <https://doi.org/10.1016/j.gloenvcha.2020.102034>
- Liu, G., Bangs, C. E., & Müller, D. B. (2012). Stock dynamics and emission pathways of the global aluminium cycle. *Nature Climate Change*, 3(4), 338-342. <https://doi.org/10.1038/nclimate1698>
- Liu, G., & Muller, D. B. (2013). Centennial evolution of aluminum in-use stocks on our aluminized planet. *Environ Sci Technol*, 47(9), 4882-4888. <https://doi.org/10.1021/es305108p>
- Liu, G., Xu, K., Zhang, X., & Zhang, G. (2014). Factors influencing the service lifespan of buildings: An improved hedonic model. *Habitat International*, 43, 274-282. <https://doi.org/10.1016/j.habitatint.2014.04.009>
- Lu, Y., Cohen, F., Smith, S. M., & Pfeiffer, A. (2022). Plant conversions and abatement technologies cannot prevent stranding of power plant assets in 2 °C scenarios. *Nat Commun*, 13(1), 806. <https://doi.org/10.1038/s41467-022-28458-7>

- Miatto, A., Schandl, H., & Tanikawa, H. (2017). How important are realistic building lifespan assumptions for material stock and demolition waste accounts? *Resources, Conservation and Recycling*, 122, 143-154. <https://doi.org/10.1016/j.resconrec.2017.01.015>
- Miller, S. A. (2020). The role of cement service-life on the efficient use of resources. *Environmental Research Letters*, 15(2). <https://doi.org/10.1088/1748-9326/ab639d>
- Miller, S. A., Habert, G., Myers, R. J., & Harvey, J. T. (2021). Achieving net zero greenhouse gas emissions in the cement industry via value chain mitigation strategies. *One Earth*, 4(10), 1398-1411. <https://doi.org/10.1016/j.oneear.2021.09.011>
- Monteiro, P. J. M., Miller, S. A., & Horvath, A. (2017). Towards sustainable concrete. *Nat Mater*, 16(7), 698-699. <https://doi.org/10.1038/nmat4930>
- Muller, D. B., Liu, G., Lovik, A. N., Modaresi, R., Pauliuk, S., Steinhoff, F. S., & Brattebo, H. (2013). Carbon emissions of infrastructure development. *Environ Sci Technol*, 47(20), 11739-11746. <https://doi.org/10.1021/es402618m>
- Muller, E., Hilty, L. M., Widmer, R., Schluep, M., & Faulstich, M. (2014). Modeling metal stocks and flows: a review of dynamic material flow analysis methods. *Environ Sci Technol*, 48(4), 2102-2113. <https://doi.org/10.1021/es403506a>
- National Bureau of Statistics of China. (2023). *National Data* <https://data.stats.gov.cn/english/>
- O'Connor, J. (2004). *Survey on actual service lives for North American buildings* Woodframe Housing Durability and Disaster Issues, Las Vegas.
- Office of the Leading Group of the Seventh National Population Census of the State Council. (2022). *China Population Census Yearbook 2020* China Statistics Press. <http://www.stats.gov.cn/sj/pcsj/rkpc/7rp/zk/indexch.htm>
- Pade, C., & Guimaraes, M. (2007). The CO₂ uptake of concrete in a 100 year perspective. *Cement and Concrete Research*, 37(9), 1348-1356. <https://doi.org/10.1016/j.cemconres.2007.06.009>
- Pauliuk, S., Heeren, N., Berrill, P., Fishman, T., Nistad, A., Tu, Q., Wolfram, P., & Hertwich, E. G. (2021). Global scenarios of resource and emission savings from material efficiency in residential buildings and cars. *Nat Commun*, 12(1), 5097. <https://doi.org/10.1038/s41467-021-25300-4>
- Pauliuk, S., Sjöstrand, K., & Müller, D. B. (2013). Transforming the Norwegian Dwelling Stock to Reach the 2 Degrees Celsius Climate Target. *Journal of Industrial Ecology*, 17(4), 542-554. <https://doi.org/10.1111/j.1530-9290.2012.00571.x>
- Pomponi, F., Hart, J., Arehart, J. H., & D'Amico, B. (2020). Buildings as a Global Carbon Sink? A Reality Check on Feasibility Limits. *One Earth*, 3(2), 157-161. <https://doi.org/10.1016/j.oneear.2020.07.018>
- Reger, D., Madanat, S., & Horvath, A. (2014). Economically and environmentally informed policy for road resurfacing tradeoffs between costs and greenhouse gas emissions. *Environmental Research Letters*, 9, 104020.
- Resources and Environment Research Branch of China National Institute of Standardization. (2022). *Interpretation of the national standard GB 16780-2021 "Energy Consumption Quota per Unit Product of Cement"*. Retrieved 2022/11/24 from https://www.cnis.ac.cn/bydt/kydt/202201/t20220104_52642.html
- Shanks, W., Dunant, C. F., Drewniok, M. P., Lupton, R. C., Serrenho, A., & Allwood, J. M. (2019). How much cement can we do without? Lessons from cement material flows in the UK. *Resources, Conservation and Recycling*, 141, 441-454. <https://doi.org/10.1016/j.resconrec.2018.11.002>

- State Administration for Market Regulation of the People's Republic of China, & National Standardization Administration of the People's Republic of China. (2021). The norm of energy consumption per unit product of cement In. Beijing, China
- State Council of the People's Republic of China. (1990). *Temporary Regulations on the transfer of the right to use of state owned land in People's Republic of China*
- Tan, C., Yu, X., & Guan, Y. (2022). A technology-driven pathway to net-zero carbon emissions for China's cement industry. *Applied Energy*, 325. <https://doi.org/10.1016/j.apenergy.2022.119804>
- Thomsen, A., & van der Flier, K. (2011). Understanding obsolescence: a conceptual model for buildings. *Building Research & Information*, 39(4), 352-362. <https://doi.org/10.1080/09613218.2011.576328>
- Voldsdund, M., Gardarsdottir, S., De Lena, E., Pérez-Calvo, J.-F., Jamali, A., Berstad, D., Fu, C., Romano, M., Roussanaly, S., Anantharaman, R., Hoppe, H., Sutter, D., Mazzotti, M., Gazzani, M., Cinti, G., & Jordal, K. (2019). Comparison of Technologies for CO₂ Capture from Cement Production—Part 1: Technical Evaluation. *Energies*, 12(3). <https://doi.org/10.3390/en12030559>
- Wang, B., Yan, L., Fu, Q., & Kasal, B. (2021). A Comprehensive Review on Recycled Aggregate and Recycled Aggregate Concrete. *Resources, Conservation and Recycling*, 171. <https://doi.org/10.1016/j.resconrec.2021.105565>
- Watari, T., Cao, Z., Hata, S., & Nansai, K. (2022). Efficient use of cement and concrete to reduce reliance on supply-side technologies for net-zero emissions. *Nat Commun*, 13(1), 4158. <https://doi.org/10.1038/s41467-022-31806-2>
- Wiedenhofer, D., Fishman, T., Plank, B., Miatto, A., Lauk, C., Haas, W., Haberl, H., & Krausmann, F. (2021). Prospects for a saturation of humanity's resource use? An analysis of material stocks and flows in nine world regions from 1900 to 2035. *Global Environmental Change*, 71. <https://doi.org/10.1016/j.gloenvcha.2021.102410>
- Wu, T., Ng, S. T., & Chen, J. (2022). Deciphering the CO₂ emissions and emission intensity of cement sector in China through decomposition analysis. *Journal of Cleaner Production*, 352. <https://doi.org/10.1016/j.jclepro.2022.131627>
- Xi, F., Davis, S. J., Ciais, P., Crawford-Brown, D., Guan, D., Pade, C., Shi, T., Syddall, M., Lv, J., Ji, L., Bing, L., Wang, J., Wei, W., Yang, K.-H., Lagerblad, B., Galan, I., Andrade, C., Zhang, Y., & Liu, Z. (2016). Substantial global carbon uptake by cement carbonation. *Nature Geoscience*, 9(12), 880-883. <https://doi.org/10.1038/ngeo2840>
- Xiao, J., Li, W., Fan, Y., & Huang, X. (2012). An overview of study on recycled aggregate concrete in China (1996–2011). *Construction and Building Materials*, 31, 364-383. <https://doi.org/10.1016/j.conbuildmat.2011.12.074>
- Xinhua New Agency. (2023). *China's 2023 economic work priorities to facilitate high-quality development.* http://english.www.gov.cn/news/topnews/202303/10/content_WS640a6f01c6d0a757729e7e88.html
- Ye, Y. (2022). When will China's population peak? It depends who you ask. <https://www.nature.com/articles/d41586-022-02304-8>
- Zhong, X., Hu, M., Deetman, S., Steubing, B., Lin, H. X., Hernandez, G. A., Harpprecht, C., Zhang, C., Tukker, A., & Behrens, P. (2021). Global greenhouse gas emissions from residential and commercial building materials and mitigation strategies to 2060. *Nat Commun*, 12(1), 6126. <https://doi.org/10.1038/s41467-021-26212-z>

