

# Tuning the Selectivity of Nitrate Reduction via Fine Composition Control of RuPdNP Catalysts

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Herein, aqueous nitrate ( $\text{NO}_3^-$ ) reduction is used to explore composition-selectivity relationships of randomly alloyed ruthenium-palladium nanoparticle catalysts to provide insights into the factors affecting selectivity during this and other industrially relevant catalytic reactions.  $\text{NO}_3^-$  reduction proceeds through nitrite ( $\text{NO}_2^-$ ) and then nitric oxide (NO), before diverging to form either dinitrogen ( $\text{N}_2$ ) or ammonium ( $\text{NH}_4^+$ ) as final products, with  $\text{N}_2$  preferred in potable water treatment but  $\text{NH}_4^+$  preferred for nitrogen recovery. It is shown that the  $\text{NO}_3^-$  and NO starting feedstocks favor  $\text{NH}_4^+$  formation using Ru-rich catalysts, while Pd-rich catalysts favor  $\text{N}_2$  formation. Conversely, a  $\text{NO}_2^-$  starting feedstock favors  $\text{NH}_4^+$  at  $\approx 50$  atomic-% Ru and selectivity decreases with higher Ru content. Mechanistic differences have been probed using density functional theory (DFT). Results show that, for  $\text{NO}_3^-$  and NO feedstocks, the thermodynamics of the competing pathways for N–H and N–N formation lead to preferential  $\text{NH}_4^+$  or  $\text{N}_2$  production, respectively, while Ru-rich surfaces are susceptible to poisoning by  $\text{NO}_2^-$  feedstock, which displaces H atoms. This leads to a decrease in overall reduction activity and an increase in selectivity toward  $\text{N}_2$  production. Together, these results demonstrate the importance of tailoring both the reaction pathway thermodynamics and initial reactant binding energies to control overall reaction selectivity.

## 1. Introduction

Traditional, colloidal heterogeneous catalysts are crucial in increasing the reaction rates and efficiencies for a great number of industrially relevant reactions. The ever increasing demands upon global energy, natural resources, and chemical products—combined with the need to limit society's impact on the environment—means that controlling catalyst selectivity remains at the forefront of nanoscience and catalysis research.<sup>[1–4]</sup> Catalyst selectivity can be tuned by significantly lowering the barrier of a desired reaction pathway with respect to other potential pathways.<sup>[5]</sup> However, exerting control over catalyst selectivity presents a greater challenge than increasing the relative catalytic activity. This is because it is more challenging to specify which reaction pathway(s) are followed in instances where multiple pathways with disparate energy barriers co-exist. It is comparatively much easier to exert control over a single rate-limiting energy barrier in order to achieve activity improvements. The ability to exert sufficient control over competing reaction pathways is often hindered by the scaling relationships between

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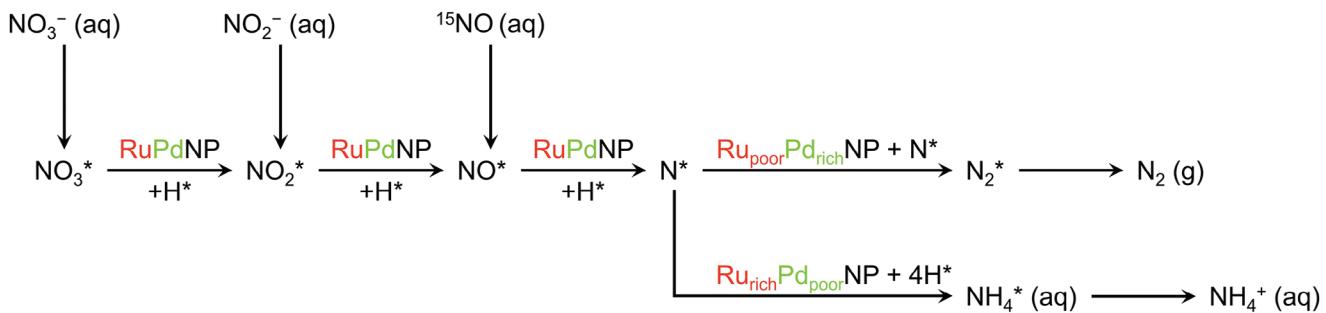
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adsorbate species involved in chemical reactions (i.e., the relationships between binding energies of the different adsorbate species), which limit the tunability of catalyst selectivity.<sup>[6–8]</sup> There are a number of approaches to break these scaling relations and improve catalyst selectivity, including tuning the catalyst surface structure,<sup>[9–12]</sup> or by changing the reaction conditions (pressure, temperature, solvent, etc.).<sup>[13–16]</sup> The reduction of nitrate ( $\text{NO}_3^-$ ) to ammonium ( $\text{NH}_4^+$ ) in drinking water treatment is a prime example of the importance of catalyst selectivity for nutrient recovery.

Nitrate is one of the most pervasive contaminants in global surface and groundwaters;<sup>[17–20]</sup> it is found naturally in the environment at concentrations  $\leq 1 \text{ mg-N L}^{-1}$ .<sup>[21]</sup> However,  $\text{NO}_3^-$  levels are often much higher, in large part due to over-application of nitrogen-rich fertilizers coupled with poor nitrogen uptake by crops.<sup>[22–24]</sup> A 12-year study by the US Geological Survey (USGS) National Water Quality Assessment Program (NAWQA) found that 50% of tested drinking water wells had  $\text{NO}_3^-$  levels above  $1 \text{ mg-N L}^{-1}$ , and 8% of wells had  $\text{NO}_3^-$  concentrations above the



**Scheme 1.** Catalytic reaction pathways for  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and  $^{15}\text{NO}$  hydrogenation over RuPdNP catalysts. The \* symbol represents adsorbed states. The end-product selectivity is determined by the divergent pathways stemming from adsorbed N which can either combine with another adsorbed N leading to  $\text{N}_2$  or with adsorbed H to eventually yield  $\text{NH}_4^+$ .

US Environmental Protection Agency (USEPA) maximum contaminant level (MCL) of  $10 \text{ mg-N L}^{-1}$ .<sup>[25]</sup> High levels of agricultural  $\text{NO}_3^-$  pollution pose major environmental concerns due to increased eutrophication and hypoxia in surface waters. Simultaneously, energy is wasted in the manufacture of ammonia-based fertilizers via the energy-intensive Haber-Bosch process.<sup>[26–32]</sup> High levels of  $\text{NO}_3^-$  also pose a major public health risk to populations that rely on this water for drinking:  $\text{NO}_3^-$  can be reduced *in vivo* to nitrite ( $\text{NO}_2^-$ ), which can cause methemoglobinemia and may react further to form potentially carcinogenic *N*-nitroso compounds. Implications of the latter include higher rates of various cancers with long-term exposure to even low levels of  $\text{NO}_3^-$  being shown to increase rates of various cancers.<sup>[33–36]</sup> Currently, the state-of-practice for  $\text{NO}_3^-$  removal from drinking water is ion exchange (IX). However, IX does not fully destroy  $\text{NO}_3^-$ , but merely transfers it from the finished water into a concentrated waste brine which requires further disposal or treatment. Additionally, IX requires large amounts of brine to regenerate exchange resins, resulting in high operating costs. Other methods (e.g., biological denitrification, reverse osmosis) have also been explored, but these strategies have their own drawbacks that hinder their implementation and as such are not the focus of this work.<sup>[37]</sup>

Both thermal catalytic and electrocatalytic reduction of  $\text{NO}_3^-$  have shown promise as potentially disruptive technologies for the direct removal of  $\text{NO}_3^-$  during water treatment by reduction of  $\text{NO}_3^-$  to create  $\text{N}_2$  or  $\text{NH}_4^+$ . The first two steps of both processes are adsorption of  $\text{NO}_3^-$  to form  $\text{NO}_3^-_{(\text{ads})}$  and then the reduction of  $\text{NO}_3^-_{(\text{ads})}$  to  $\text{NO}_2^-_{(\text{ads})}$ . Generally, the  $\text{NO}_2^-_{(\text{ads})}$  is then subsequently reduced to  $\text{NO}_{(\text{ads})}$ , and then, through a series of intermediates to either  $\text{N}_2$  or  $\text{NH}_4^+$  (Scheme 1).<sup>[38–40]</sup> Typically during electrocatalysis, the reduction process occurs through direct electron transfer from the electrode to the adsorbed N-species. The competition of the various mechanisms during electrocatalytic  $\text{NO}_3^-$  reduction, which have been well summarized in a number of recent reviews, have a direct influence on the end-product selectivity. Traditionally, coinage metals (Cu, Ag, Au) and platinum-group metals (PGM; Rh, Ru, Ir, Pt, Pd) are used as cathodic materials for electrocatalytic  $\text{NO}_3^-$  reduction, though a number of other metals and metal oxides (e.g., Ti,  $\text{TiO}_2$ , Fe, Bi) have also been studied in an attempt to improve catalyst performance or to target certain end-products.<sup>[39,41,42]</sup> Some of these studies have shown that the mechanism of electrocatalytic  $\text{NO}_3^-$  reduction is highly dependent on the actual exposed facets of the metal elec-

trode, and exposing ideal facets can improve performance and alter selectivity.<sup>[43,44]</sup> Further, the addition of secondary metals has been used to improve activity during electrocatalytic  $\text{NO}_3^-$  and  $\text{NO}_2^-$  reduction and to direct end-product selectivity.<sup>[45–48]</sup> However, these studies have only explained why one end-product is more favored than another, and they have not yielded tunable end-product control. More recent work also suggests that while electrocatalytic and thermal catalytic  $\text{NO}_3^-$  reduction may follow similar pathways, there are major mechanistic differences between the two technologies, and the study of one should not exclude the other.<sup>[40,46,49]</sup>

In thermal catalysis, the reduction processes use atomic hydrogen as the electron donor. The atomic hydrogen is usually generated *in situ* by oxidative dissociation of  $\text{H}_2$  using a PGM. The first step of transforming adsorbed  $\text{NO}_3^-$  to  $\text{NO}_2^-$  often requires the presence of a promoter metal (e.g., In, Cu, Sn) to help activate the first N-O bond in  $\text{NO}_3^-$ , and generally this step is not regarded as rate limiting for thermal catalysis.<sup>[50,51]</sup> Subsequently,  $\text{NO}_2^-$  migrates to the PGM where it is further reduced to  $\text{NO}$ , and then, through a series of intermediates, to  $\text{N}_2$  or  $\text{NH}_4^+$ . The promoter is then regenerated by reduction with additional  $\text{H}_2$ .<sup>[6,51–55]</sup> Historically, Pd has been the preferred PGM for thermal catalysis (with In, Cu, or Sn as a promoter) due to its relatively high activity and its high selectivity for  $\text{N}_2$ .<sup>[52,56,57]</sup> Formation of  $\text{NH}_4^+$  has commonly been avoided as it promotes bacterial growth in finished water; this negatively impacts the taste and odor of drinking water, and raises concerns regarding growth of pathogens.<sup>[58–60]</sup> However,  $\text{NH}_4^+$  is a valuable product that can be re-used in agriculture, such that its recovery has more recently garnered attention. Alternatives to Pd-based catalysts (e.g., nickel phosphide ( $\text{Ni}_2\text{P}$ ), ruthenium (Ru), etc.) have emerged as promising alternatives that can achieve near complete selectivity toward  $\text{NH}_4^+$  formation upon  $\text{NO}_3^-$  reduction, and, importantly, they do not require a promoter metal for the first step of  $\text{NO}_3^-$  reduction to  $\text{NO}_2^-$ .<sup>[61,62]</sup> Unfortunately,  $\text{Ni}_2\text{P}$  catalysts can only activate  $\text{NO}_3^-$  in low pH environments (pH 2–4),<sup>[61]</sup> whereas, Ru-based catalysts show high activity for  $\text{NO}_3^-$  reduction at pH ranges closer to those of typical drinking water (e.g., 5.0–5.5).<sup>[62]</sup> Ru is also significantly less expensive than Pd ( $\$14.95 \text{ g}_{\text{Ru}}^{-1}$  versus  $\$42.91 \text{ g}_{\text{Pd}}^{-1}$ ; MetalsDaily and APMEX, Inc.; August 2023). A clear mechanistic understanding of how Pd and Ru surfaces alter the binding energies of species involved in competing  $\text{NO}_3^-$  reduction pathways is notably absent in this arena. Having a detailed understanding of key catalyst structure-function relationships can help

explain why certain catalyst compositions select for different end-products, and how reaction thermodynamics can be exploited to obtain desired ratios of N<sub>2</sub> to NH<sub>4</sub><sup>+</sup>; to-date, these relationships remain poorly understood from a fundamental standpoint. Although mixed ratios of N<sub>2</sub> and NH<sub>4</sub><sup>+</sup> might not be desirable for most nitrate reduction applications, the ability to tune binding energies to obtain specific product mixtures would have much wider applications in industrial-scale catalysis.

The goal of this work is to explore structure-selectivity correlations in ruthenium-palladium nanoparticles (Ru<sub>x</sub>Pd<sub>100-x</sub>NPs) as a function of catalyst composition, which can itself be dictated by synthetic design. We synthesized a series of Ru<sub>x</sub>Pd<sub>100-x</sub>NPs with target ratios that span the entire composition range, corresponding to  $x = 10, 30, 50, 70,$  and  $90$ , by adapting previously published methods [63] that utilize a convenient and scalable microwave (MW)-assisted synthetic approach.[64,65] We then characterized the resulting solid-solution nanoparticles using a variety of techniques, including powder X-ray diffraction (PXRD), transmission electron microscopy (TEM), scanning transmission electron microscopy with energy dispersive X-ray spectroscopy (STEM-EDS), X-ray photoelectron spectroscopy (XPS), and inductively coupled plasma optical emission spectrometry (ICP-OES). This structural information is combined with experimental model catalytic studies, which are able to provide kinetic information of NO<sub>3</sub><sup>-</sup>, NO<sub>2</sub><sup>-</sup>, and NO reduction as a function of Ru-Pd composition. Finally, we employed computational chemistry in the form of density functional theory (DFT) to elucidate the observed reactivity and to relate it to catalyst behavior as pertains to the most prevalent NO<sub>3</sub><sup>-</sup> reduction pathways that result in selective formation of N<sub>2</sub> and NH<sub>4</sub><sup>+</sup>.

All the Ru<sub>x</sub>Pd<sub>100-x</sub>NP catalysts studied in this work were found to display pseudo-first-order reaction rates for NO<sub>3</sub><sup>-</sup>, NO<sub>2</sub><sup>-</sup>, and NO reduction by H<sub>2</sub>, and the apparent hydrogenation activity was found to be heavily dependent on both the catalyst composition and the starting reagents employed. Furthermore, we found that selectivity towards NH<sub>4</sub><sup>+</sup> is indeed tunable as a function of catalyst composition, driven by competitive binding between N-species and H<sub>2</sub> at various Ru-Pd surface ensembles, which are present in different frequencies based upon the specific Ru:Pd ratios. Through our combined experimental and theoretical work, we were able to explore the hypothesis, that: “intrinsic variations in reactivity and adsorption of various reaction species on different Ru<sub>x</sub>Pd<sub>100-x</sub>NP surfaces are responsible for the observed difference in end-product selectivity, as well as the difference in activity.”

## 2. Experimental Results

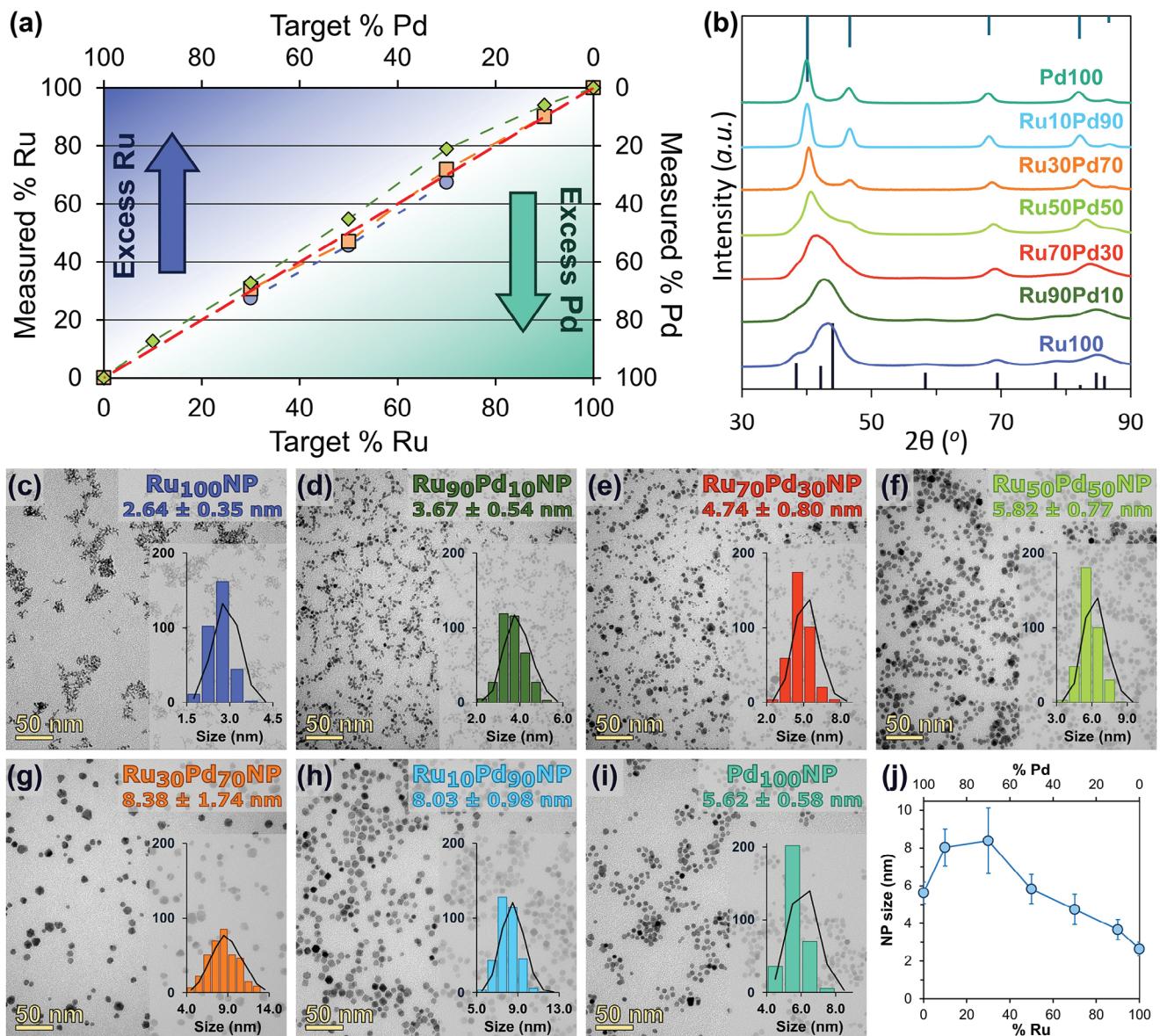
### 2.1. Microwave Synthesis of RuPdNPs

Ruthenium-palladium nanoparticles with random, solid-solution structures and finely tunable Ru:Pd compositions (Ru<sub>x</sub>Pd<sub>100-x</sub>NPs) were prepared by the co-reduction of common Ru<sup>3+</sup> and Pd<sup>2+</sup> salts using a microwave-assisted polyol method. The target value of  $x$  was determined based on the molar ratio of metal precursors employed in a given reaction, such that the total molar amount of metal (Ru + Pd) was 0.10 mmol in all cases (Table S1, Supporting Information). In general, a solution containing a specified molar ratio of RuCl<sub>3</sub> and K<sub>2</sub>PdCl<sub>4</sub> was

dissolved in ethylene glycol (EG) and injected directly into a hot solution of EG containing excess poly(vinylpyrrolidone) (PVP) that was pre-heated to 185 °C, at a controlled rate, within a CEM MARS-5 microwave reactor. After heating for 30 min, the resulting PVP-capped RuPdNPs were isolated by precipitation (see the Experimental Methods section in the Supporting Information for details). ICP-OES revealed actual resulting compositions with  $x = 83.9, 67.4, 45.6, 27.5,$  and  $11.1$ , compared to the notional (target) compositions of  $x = 90, 70, 50, 30,$  and  $10$ , as shown in Figure 1a; Table S2 (Supporting Information). The uniformly close agreement between target and actual compositions across all Ru:Pd ratios indicates that Ru<sup>3+</sup> and Pd<sup>2+</sup> were successfully co-reduced under the chosen reaction conditions, without the co-formation of secondary, monometallic species. The lattice structures of the Ru<sub>x</sub>Pd<sub>100-x</sub>NPs were investigated using powder X-ray diffraction (PXRD). In their monometallic forms, Ru exhibits hexagonal close packing (*hcp*), while Pd resides in the face-centered cubic (*fcc*) lattice. Ru can be crystallized in a relatively stable, kinetic *fcc* setting in small nanoparticles,[66–68] and as such, it is known to readily adopt an *fcc* structure when alloyed with another *fcc* metal.[63,69,70] Interestingly, in this case, Figure 1b shows that the diffraction patterns of the samples revealed characteristics of both *hcp* and *fcc* lattice structures (Figure 1b). The extent to which one structure dominated over the other was dependent on the particle composition, such that Ru-rich particles displayed greater *hcp* character, while Pd-rich particles showed more *fcc* character, in agreement with previous reports.[63,71,72]

Transmission electron microscopy (TEM; Figure 1c–i) reveals quasi-spherical particles for all Ru<sub>x</sub>Pd<sub>100-x</sub>NPs, contrary to previous reports of Ru-rich RuPdNPs which indicated elongated, non-spherical particles.[63] This difference in particle morphology could be due to capping agent effects from the higher PVP-to-metal ratio used in this work versus that of Kusada et al., and/or the presence of hotspots generated by the interaction of the microwave irradiation, in which early nanoparticle seeds could facilitate isotropic growth. In this work, as the atomic percent of Pd in the alloy was increased, the particle shape became more cuboctahedral, which is consistent with the equilibrium structure of *fcc* nanocrystals, such as pure PdNPs (Figure 1i).[73,74] As shown by Figure 1j, the size of the particles increases from monometallic RuNPs (2.6 nm) as the amount of Pd increased. The size reached a maximum at Ru<sub>30</sub>Pd<sub>70</sub>NPs (8.4 nm) and then decreased to Ru<sub>10</sub>Pd<sub>90</sub>NPs (8.0 nm) and again to monometallic PdNPs (5.6 nm). All compositions of Ru<sub>x</sub>Pd<sub>100-x</sub>NPs displayed narrow size dispersity ( $\sigma < \pm 20\%$ ).

Scanning transmission electron microscopy with energy dispersive X-ray spectroscopy (STEM-EDS) analyses were performed to confirm random alloying of Ru and Pd within the particles as opposed to a mix of RuNPs and PdNPs or a segregated structure; representative results of the Ru<sub>50</sub>Pd<sub>50</sub>NPs are shown in Figure 2. Figure 2b–d show the maps of Ru, Pd, and their elemental overlays, respectively. Taken together, these maps give direct evidence of a solid-solution Ru-Pd structure, despite their classical immiscibility in bulk at room temperature.[75] EDS line scan analyses (Figure 2e,f) further corroborate solid-solution alloys within individual particles. STEM-EDS maps of various compositions (Figure S1–S4, Supporting Information) confirm that, on the nanoscale, Ru and Pd can be alloyed across the entire

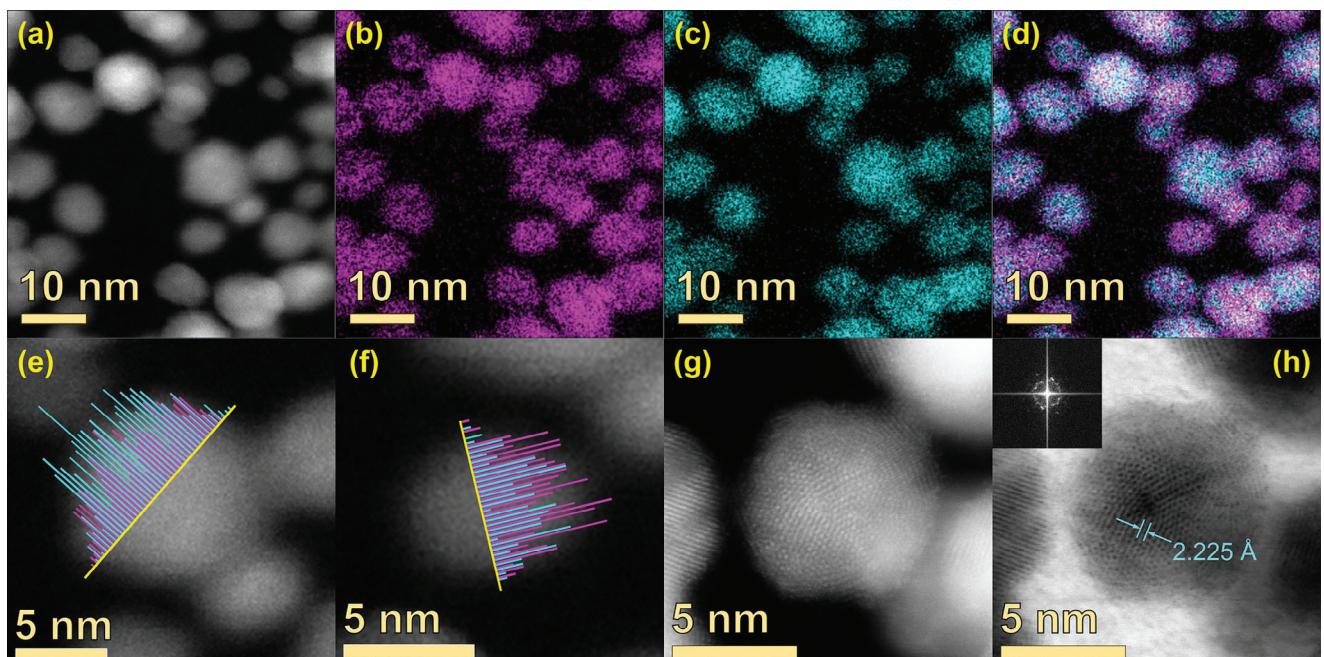


**Figure 1.** a) Comparison of measured nanoparticle compositions by different techniques versus the target composition (dashed red line): ICP-OES (blue circles), EDS (green diamonds), and XPS (orange squares). b) Powder X-ray diffraction patterns for Ru<sub>x</sub>Pd<sub>100-x</sub>NPs. The vertical lines represent expected reflections for hcp Ru (black; from the bottom) and fcc Pd (teal; from the top). Representative low-resolution TEM images for c) RuNPs, d) Ru<sub>90</sub>Pd<sub>10</sub>NPs, e) Ru<sub>70</sub>Pd<sub>30</sub>NPs, f) Ru<sub>50</sub>Pd<sub>50</sub>NPs, g) Ru<sub>30</sub>Pd<sub>70</sub>NPs, h) Ru<sub>10</sub>Pd<sub>90</sub>NPs, and i) PdNPs. The insets show size distributions with histograms of measured sizes. Scale bars shown are 50 nm. j) Relationship between particle composition and particle size for Ru<sub>x</sub>Pd<sub>100-x</sub>NPs.

composition range. The composition results from EDS analysis yield values of  $x = 94.1, 78.9, 54.7, 32.8$ , and  $12.6$ , in relatively close agreement with values calculated from ICP (Figure 1a; Table S2, Supporting Information). STEM analyses of the monometallic RuNPs and PdNPs show that both monometallic samples consisted of single crystalline particles (Figure S5 and S6, Supporting Information, respectively). Analysis of the RuNPs yields a lattice d-spacing of  $2.061\text{ \AA}$ , closely corresponding to the  $\langle 101 \rangle$  planes of pure hcp Ru (expected d-spacing of  $2.055\text{ \AA}$ ). Analysis of the monometallic PdNPs yields a d-spacing of  $2.247\text{ \AA}$ , which agrees with the expected d-spacing of the  $\langle 111 \rangle$  planes of pure fcc Pd ( $2.246\text{ \AA}$ ). The STEM images of the alloys show that

the alloyed particles were polycrystalline with small domains (see Figure 2g,h, Figure S7–S10, Supporting Information). For particles with  $x \leq 50$ , the measured d-spacing corresponds to the fcc  $\langle 111 \rangle$  planes, with the measured d-spacing decreasing with increasing Ru content ( $x$ ). The measured d-spacing of Ru<sub>70</sub>Pd<sub>30</sub>NPs and Ru<sub>90</sub>Pd<sub>10</sub>NPs corresponds well to what was expected for hcp  $\langle 002 \rangle$  planes, and the measured d-spacings again decrease with increasing Ru content ( $x$ ). Our results are in close agreement with previous reports of alloyed RuPdNPs that have demonstrated that individual particles consist of small hcp and fcc domains.<sup>[63]</sup>

The surface and near subsurface compositions of the Ru<sub>x</sub>Pd<sub>100-x</sub>NPs were assessed by X-ray photoelectron



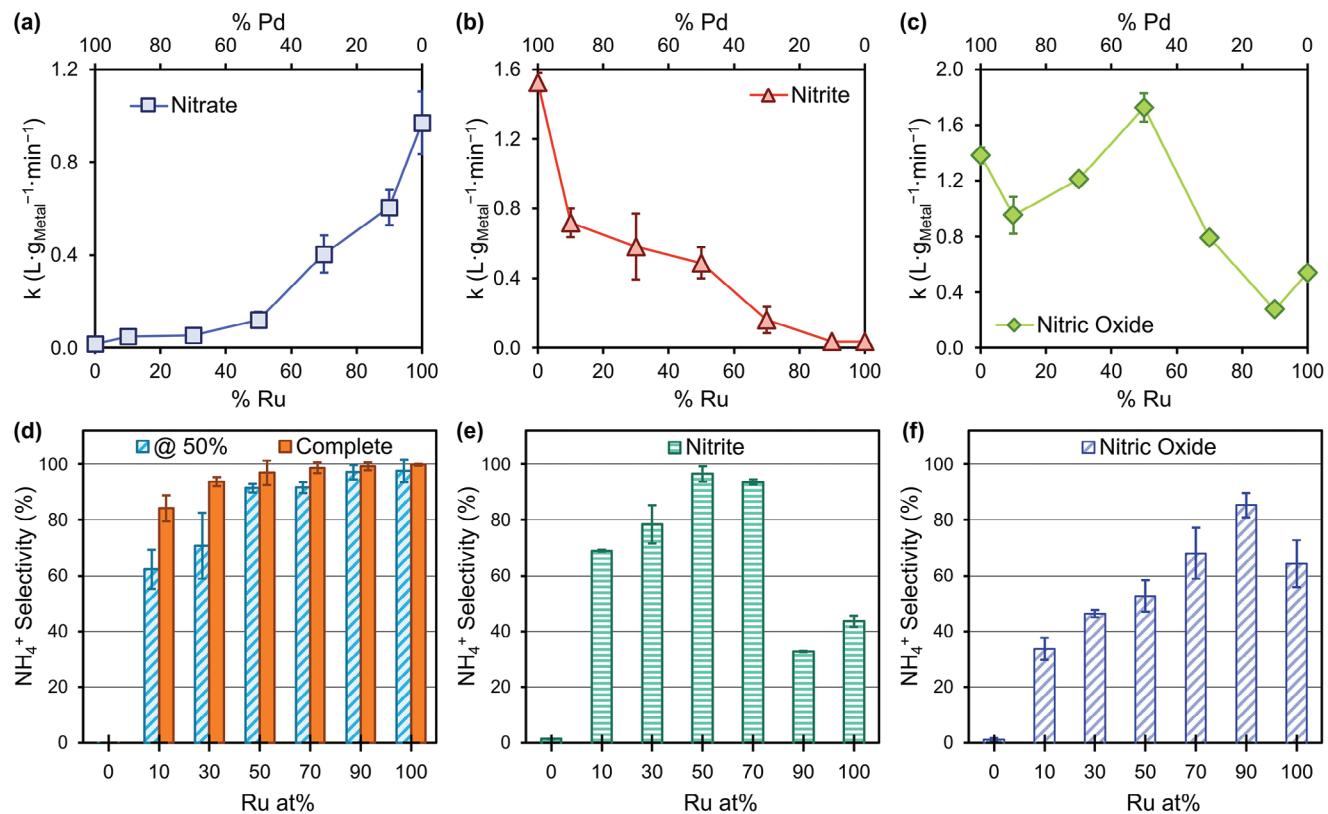
**Figure 2.** a) HAADF-STEM image, b) Ru-L EDS map, c) Pd-L EDS map, and d) Ru-Pd overlay of several  $\text{Ru}_{50}\text{Pd}_{50}$ NPs. Scale bars shown are 10 nm. e,f) Compositional EDS line profiles for Ru (purple) and Pd (cyan) for single  $\text{Ru}_{50}\text{Pd}_{50}$ NPs. Scale bars shown are 5 nm. g) HAADF-STEM image and h) corresponding BF-STEM image of a single  $\text{Ru}_{50}\text{Pd}_{50}$ NP. Scale bars shown are 5 nm. Shown in h is the measured lattice d-spacing of 2.225 Å, which corresponds to the fcc  $\langle 111 \rangle$  planes of the  $\text{Ru}_{50}\text{Pd}_{50}$ NPs. The inset in h shows the forward Fourier transform (FFT) calculated from the NP.

spectroscopy (XPS), where the X-ray penetration depth was somewhat selective to the NP exteriors. These data are in good agreement with both the target nominal compositions and the corresponding compositions measured by ICP-OES and EDS (Figure 1a; Table S2, Supporting Information). In direct agreement with elemental mapping studies, Ru surface enrichment was not observed, despite the difference in reduction potentials of the two metals:  $E_{\text{Ru}} = 0.60$  V versus  $E_{\text{Pd}} = 0.95$  V.<sup>[76]</sup> The lack of an observed surface Ru bias is likely due to the presence of PVP overlayers; previous studies have demonstrated that adsorbates such as H<sup>[77]</sup> and N<sup>[6]</sup> can stabilize Pd atoms at NP surfaces. XPS was also used to examine the distribution of average oxidation states of surface atoms, which were exposed to air during synthesis and storage and were therefore assumed to be oxidized to a greater extent than when under reducing catalytic conditions (Figure S11–S13, Supporting Information). The subsurface Ru and Pd atoms are expected to be zero-valent since they are fully coordinated with other metal atoms. The Ru 3p peaks were chosen for analysis due to the superimposition and resulting convolution of the Ru 3d region with the C 1s region (Figure S12, Supporting Information). For all Ru-containing samples, the Ru atoms at the particle surface were present as  $\text{Ru}^0$ , with the Ru 3p<sub>3/2</sub> peak appearing at 461.7 eV (Table S3, Supporting Information; cf. 461.20–461.70 eV).<sup>[78,79]</sup> XPS analysis revealed a combination of  $\text{Pd}^0$  and  $\text{Pd}^{2+}$  for all Pd-containing catalysts, with Pd 3d<sub>5/2</sub> peaks at 335.3 eV and 337.0 eV for  $\text{Pd}^0$  and  $\text{Pd}^{2+}$ , respectively. The Pd 3d<sub>3/2</sub> peaks appear approximately 5.3 eV higher than the Pd 3d<sub>5/2</sub> peaks (Table S4, Supporting Information). Generally, ≈25% of the Pd surface atoms were present in the  $\text{Pd}^{2+}$  oxidation state for all compositions (Table S5, Supporting Information); this was in direct opposition to prior

work with Pd-based alloys where the fraction of  $\text{Pd}^{2+}$  decreased as the amount of Pd in the alloy decreased.<sup>[80–82]</sup> However, it was anticipated that surface oxidation should not impact catalyst performance, since the catalysts require pre-reduction under flowing  $\text{H}_2$  gas, so all exposed metal surfaces should be reduced to their metallic, zero-valent state.<sup>[83]</sup>

## 2.2. Catalytic Nitrate Reduction

The as-synthesized  $\text{Ru}_x\text{Pd}_{100-x}$ NPs, RuNPs, and PdNPs were supported on amorphous silica ( $\alpha\text{-SiO}_2$ ) via direct deposition from aqueous-organic suspensions, with notional targeted metal loadings of 4–5 wt% (Ru + Pd). ICP-OES analysis of the supported catalysts confirmed successful loading of all RuPdNP compositions on  $\alpha\text{-SiO}_2$ , with actual loadings between 3.8–4.7 wt% (Table S6, Supporting Information). Analyses of the supported particles by TEM demonstrated no change in particle morphology during the deposition process (Figure S14, Supporting Information). Further, STEM-EDS analysis of  $\text{SiO}_2$ -supported  $\text{Ru}_{50}\text{Pd}_{50}$ NPs after catalysis (Figure S15, Supporting Information) confirmed the stability of the NPs on the  $\text{SiO}_2$  support, and showed that the atomic structure of the NPs was stable throughout both deposition and catalysis. Amorphous  $\text{SiO}_2$  was chosen here as the catalyst support because it is relatively inert and thus should not play a role in the measured catalytic activity or selectivity, allowing for examination of the intrinsic structure-function relationship of RuPdNPs.<sup>[84,85]</sup> Catalytic reduction experiments using  $\text{NO}_3^-$  were performed in semi-batch reactors using  $\text{H}_2$  gas as the reducing agent. For  $\text{NO}_3^-$  reduction, all catalysts displayed pseudo-first-order kinetics (Figure S16, Supporting



**Figure 3.** First-order rate constants for a) aqueous NO<sub>3</sub><sup>-</sup> reduction, b) aqueous NO<sub>2</sub><sup>-</sup> reduction, and c) aqueous <sup>15</sup>NO reduction normalized to total metal (i.e., Ru + Pd) as wt% of catalyst in units of liters per gram metal per min. NH<sub>4</sub><sup>+</sup> selectivity (as a fraction of mol NH<sub>4</sub><sup>+</sup> produced per mol N lost) at d) 50% conversion (blue striped bars) and complete conversion of initial NO<sub>3</sub><sup>-</sup> (orange solid bars), e) complete conversion during NO<sub>2</sub><sup>-</sup>, and f) complete conversion during <sup>15</sup>NO reduction.

Information), from which apparent first-order rate constants were calculated (Table S7, Supporting Information). These rate constants were then normalized to the total amount of metal in each catalyst as shown in Figure 3a–c. These rate constants were further normalized to turnover frequencies (TOFs; Figure S17a, Supporting Information), and the trends in activity between the mass-normalized activity and the TOFs were in good agreement.

The catalytic activity of Ru<sub>x</sub>Pd<sub>100-x</sub>NPs in NO<sub>3</sub><sup>-</sup> reduction was found to have a direct correlation with the value of  $x$ : as the amount of Ru in the catalyst decreased, the NO<sub>3</sub><sup>-</sup> activity also decreased. The monometallic RuNP-SiO<sub>2</sub> catalyst displayed the highest metal-normalized activity of 0.97 L g<sub>M</sub><sup>-1</sup> min<sup>-1</sup>, several times lower than reported by Huo et al. (cf. 4.13 L g<sub>M</sub><sup>-1</sup> min<sup>-1</sup>).<sup>[62]</sup> This difference may reflect the different reaction conditions used here versus those used by Huo and coworkers, or intrinsic reactivity differences that stem from the different preparation methods of the RuNPs in that work and ours. The monometallic PdNP-SiO<sub>2</sub> catalyst displayed very little degradation of NO<sub>3</sub><sup>-</sup>, with a calculated activity of 0.01 L g<sub>M</sub><sup>-1</sup> min<sup>-1</sup>; this was in good agreement with other studies using Pd-based catalysts for NO<sub>3</sub><sup>-</sup> reduction.<sup>[52,57,86,87]</sup> The decreasing activity of the alloyed Ru<sub>x</sub>Pd<sub>100-x</sub>NP-SiO<sub>2</sub> catalysts with decreasing amount of Ru present was due to the limited ability of Pd to abstract the first oxygen from NO<sub>3</sub><sup>-</sup> to initiate the reaction. The estimated Damköhler coefficient (Da) was <<1 for the least active catalyst

(Ru<sub>10</sub>Pd<sub>90</sub>NP-SiO<sub>2</sub>), indicating that external mass transfer limitations of NO<sub>3</sub><sup>-</sup> and H<sub>2</sub> on the reactions were negligible; similarly, the Weisz-Prater parameter (C<sub>WP</sub>) was <<1, indicating that internal mass transfer within the SiO<sub>2</sub> pores did not limit the overall reaction rates (see Supporting Information for details on the calculations of both parameters).

The end-product selectivity of the Ru<sub>x</sub>Pd<sub>100-x</sub>NP-SiO<sub>2</sub> catalysts was assessed at both 50% loss and complete loss of NO<sub>3</sub><sup>-</sup> (Figure 3d and Table 1). In both instances, the selectivity for NH<sub>4</sub><sup>+</sup> decreased as the amount of Ru in the catalyst ( $x$ ) was

**Table 1.** Selectivity of NH<sub>4</sub><sup>+</sup> versus N<sub>2</sub> (as % NH<sub>4</sub><sup>+</sup>) during NO<sub>3</sub><sup>-</sup>, NO<sub>2</sub><sup>-</sup>, and NO hydrogenation over Ru<sub>x</sub>Pd<sub>100-x</sub>NP-SiO<sub>2</sub> catalysts.

Ru:Pd	NO <sub>3</sub> <sup>-</sup> reduction <sup>a)</sup>	NO <sub>3</sub> <sup>-</sup> reduction <sup>b)</sup>	NO <sub>2</sub> <sup>-</sup> reduction <sup>b)</sup>	NO reduction <sup>b)</sup>
100:0	97.6 (4.2)	99.8 (0.3)	43.7 (2.0)	64.3 (8.5)
90:10	97.1 (2.6)	99.2 (1.4)	32.7 (0.2)	85.3 (4.3)
70:30	91.6 (1.9)	98.7 (1.9)	93.6 (0.9)	68.0 (9.2)
50:50	91.4 (1.5)	96.9 (4.4)	96.5 (2.8)	52.7 (5.6)
30:70	70.8 (11.8)	93.7 (1.6)	78.4 (6.8)	46.3 (1.3)
10:90	62.3 (7.0)	84.1 (4.7)	68.9 (0.3)	33.7 (3.9)
0:100	—	—	1.4 (0.1)	1.2 (0.4)

<sup>a)</sup> Measured at 50% loss of NO<sub>3</sub><sup>-</sup>; <sup>b)</sup> Measured at 100% loss of N-parent species. Values in parentheses represent standard deviation of measurements.

decreased. The  $\text{NH}_4^+$  selectivity of the catalysts increased as  $\text{NO}_3^-$  conversion approached complete conversion. This was likely due to the decreasing concentration of  $\text{NO}_3^-$  in solution, which was anticipated to decrease the N:H ratio on the catalyst surface, thus making the formation of N-H bonds more likely.<sup>[54,88–90]</sup> Supported nanoparticle catalysts with intermediate or higher amounts of Ru displayed > 90% selectivity towards  $\text{NH}_4^+$  even at 50% reduction of  $\text{NO}_3^-$ ; in contrast, more Pd-rich catalysts displayed significantly less  $\text{NH}_4^+$  selectivity (70.8% and 62.3% for  $x = 30$  and 10, respectively).

### 2.3. Nitrite and NO as Reduction Intermediates

The current mechanistic understanding of  $\text{NO}_3^-$  reduction by noble metal catalysts indicates that  $\text{NO}_3^-$  is first reduced to  $\text{NO}_2^-$ , which is further reduced to NO, then transformed to N, which can ultimately lead to either  $\text{NH}_4^+$  formation by reaction with nearby H atoms or  $\text{N}_2$  formation by direct reductive elimination.<sup>[51,62,86,91,92]</sup> In this study, neither  $\text{NO}_2^-$  nor NO was detected during  $\text{NO}_3^-$  reduction, indicating that the reduction of both species was faster than their formation under the measured conditions or they were strongly sorbed. To better understand the impact of both species on the selectivity of  $\text{NO}_3^-$  reduction, reactions were performed with either  $\text{NO}_2^-$  or NO as the initial N-species.

When reactions were initiated using  $\text{NO}_2^-$ , all catalysts again displayed pseudo-first-order kinetics from which apparent first-order rate constants were calculated before being normalized to the total metal content (Figure S18 and Table S7, Supporting Information) and to TOFs (Figure S17b, Supporting Information). The activity of  $\text{Ru}_x\text{Pd}_{100-x}$ NPs for  $\text{NO}_2^-$  reduction decreased as the amount of Ru in the catalyst ( $x$ ) increased (Figure 3b). This trend was in direct opposition to that seen when  $\text{NO}_3^-$  was the initial reactant species. The monometallic PdNP-SiO<sub>2</sub> catalyst displayed the highest activity of  $1.53 \text{ L g}_M^{-1} \text{ min}^{-1}$  (cf.  $1.3 \text{ L g}_M^{-1} \text{ min}^{-1}$  from Hörold et al.,<sup>[52]</sup>  $1.35 \text{ L g}_M^{-1} \text{ min}^{-1}$  from Troutman et al.,<sup>[82]</sup> and  $2.0 \text{ L g}_M^{-1} \text{ min}^{-1}$  from Seraj et al.<sup>[80]</sup>). Only  $\approx 15\%$  of the  $\text{NO}_2^-$  was degraded after three hours when using the pure RuNP-SiO<sub>2</sub> or  $\text{Ru}_{10}\text{Pd}_{90}$ NP-SiO<sub>2</sub> catalysts ( $k = 0.036 \text{ L g}_M^{-1} \text{ min}^{-1}$  for both). This trend agrees with previous reports, where the  $\text{NO}_2^-$  reduction activity on Ru-catalysts was much lower than Pd-catalysts.<sup>[62]</sup> Selectivity behavior of the  $\text{Ru}_x\text{Pd}_{100-x}$ NP-SiO<sub>2</sub> catalysts during  $\text{NO}_2^-$  reduction was completely different than the selectivity during  $\text{NO}_3^-$  reduction (Figure 3e and Table 1). In agreement with previous reports, the PdNP-SiO<sub>2</sub> catalyst displayed almost complete selectivity for  $\text{N}_2$  gas (98.6%  $\text{N}_2$ ), while the RuNP-SiO<sub>2</sub> catalyst yielded a mix of  $\text{NH}_4^+$  and  $\text{N}_2$  (43.7%  $\text{NH}_4^+$  versus 56.3%  $\text{N}_2$ ).<sup>[62,80,82]</sup> All alloy catalysts except  $\text{Ru}_{10}\text{Pd}_{90}$ NP-SiO<sub>2</sub> displayed higher selectivity towards  $\text{NH}_4^+$  than the pure RuNP-SiO<sub>2</sub> catalyst. The  $\text{Ru}_{50}\text{Pd}_{50}$ NP-SiO<sub>2</sub> displayed near-complete selectivity towards  $\text{NH}_4^+$ , closely followed by  $\text{Ru}_{70}\text{Pd}_{30}$ NP-SiO<sub>2</sub> (96.5%  $\text{NH}_4^+$  and 93.6%  $\text{NH}_4^+$ , respectively).

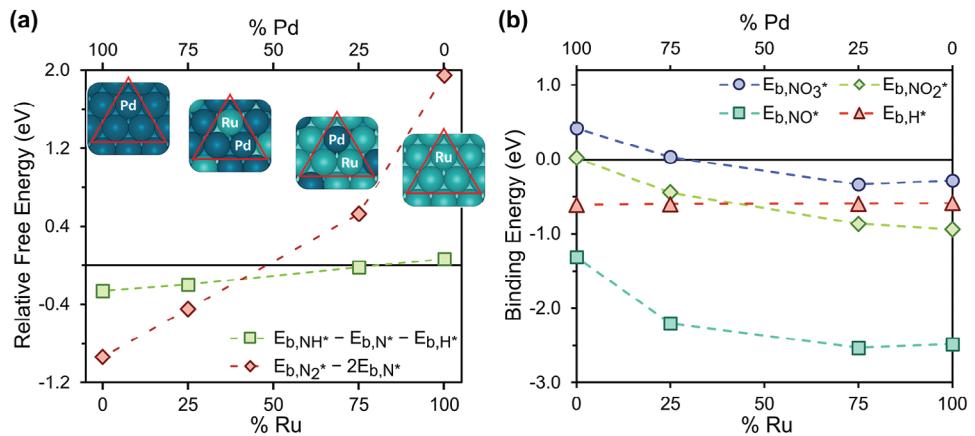
When reactions were initiated using  $^{15}\text{NO}$ , all catalysts displayed pseudo-first-order kinetics and demonstrated near-complete ( $\geq 90\%$ ) reduction of all  $^{15}\text{NO}$  (Figure S19 and Table S7, Supporting Information). The  $^{15}\text{NO}$  reduction activity of the  $\text{Ru}_x\text{Pd}_{100-x}$ NP-SiO<sub>2</sub> displayed a volcano-shaped dependence on

composition (Figure 3c; Figure S17c, Supporting Information), in contrast with both  $\text{NO}_3^-$  and  $\text{NO}_2^-$  reduction activity. The pure PdNP-SiO<sub>2</sub> catalyst outperformed the pure RuNP-SiO<sub>2</sub> ( $1.34 \text{ L g}_M^{-1} \text{ min}^{-1}$  versus  $0.55 \text{ L g}_M^{-1} \text{ min}^{-1}$ , respectively). Upon incorporation of a small amount of secondary metal (e.g.,  $\text{Ru}_{90}\text{Pd}_{10}$ NPs and  $\text{Ru}_{10}\text{Pd}_{90}$ NPs), the activity was markedly decreased. The  $\text{Ru}_{50}\text{Pd}_{50}$ NP-SiO<sub>2</sub> catalyst displayed the highest  $^{15}\text{NO}$  reduction activity of  $1.80 \text{ L g}_M^{-1} \text{ min}^{-1}$ . The selectivity behavior of the  $\text{Ru}_x\text{Pd}_{100-x}$ NP-SiO<sub>2</sub> catalysts is similar to that during  $\text{NO}_3^-$  reduction: generally, as the amount of Ru in the alloy (i.e., the value of  $x$ ) increases the catalyst becomes more selective for  $\text{NH}_4^+$  formation (Figure 3f and Table 1). The pure PdNP-SiO<sub>2</sub> catalyst displayed very little  $\text{NH}_4^+$  formation during  $^{15}\text{NO}$  reduction (1.2%), in agreement with previous studies looking at NO reduction using a Pd catalyst, which found that Pd catalysts display high (> 90%)  $\text{N}_2$  selectivity.<sup>[93]</sup> When a small amount (e.g., 10%) of Ru was added to the catalyst, the  $\text{NH}_4^+$  selectivity increased to 33.7%. The selectivity steadily increased to reach that of the  $\text{Ru}_{90}\text{Pd}_{10}$ NP-SiO<sub>2</sub> catalyst (85.3%  $\text{NH}_4^+$ ). However, the pure RuNP-SiO<sub>2</sub> notably did not follow this trend; instead,  $^{15}\text{NO}$  reduction over the RuNP-SiO<sub>2</sub> resulted in a mixture of both  $^{15}\text{NH}_4^+$  and  $^{15}\text{N}_2$  gas (64.3%  $^{15}\text{NH}_4^+$  and 35.7%  $^{15}\text{N}_2$ ).

### 3. Computational Results

As mentioned above, selectivity can be controlled either by changing the reaction conditions (i.e., pH,  $\text{H}_2$  gas flow rate) or by changing the intrinsic reactivity via composition/structure tuning (resulting in the control of relative barrier heights). It is well documented that both pH and  $\text{H}_2$  flow rate play a major role in both  $\text{NO}_3^-$  removal activity and selectivity towards  $\text{NH}_4^+$  versus  $\text{N}_2$  when using Pd-based catalysts.<sup>[52,89,92]</sup> Ru-catalysts have also shown pH-dependent  $\text{NO}_3^-$  reduction activity, although the catalytic selectivity remained high (> 90%) at all tested pH values (pH 3–10).<sup>[62]</sup> Additionally, all  $\text{NO}_3^-$  reduction experiments in this work were well buffered (pH =  $6.4 \pm 0.2$ ) and were performed with the same  $\text{H}_2$  flow rate. Therefore, neither the pH nor the  $\text{H}_2$  flow were expected to cause the observed selectivity differences of the alloy catalysts.

We hypothesize that the observed difference in end-product selectivity, as well as the differences in activity, are due to variations in intrinsic reactivity and adsorption of various reaction species on the different  $\text{Ru}_x\text{Pd}_{100-x}$ NP surfaces. This infers that, when reactivity dominates, the observed selectivity is primarily determined by the thermodynamics of the competing pathways for N-species reduction. By comparison, when surface coverage effects are dominant, the product selectivity is controlled by the ratio of adsorbed N versus H species that are co-adsorbed on the nanoparticle surfaces, which determines whether N-H or N-N bond formation is more likely (regardless of potential differences in the relative energies of the competing pathways). To evaluate our hypothesis, we used DFT to calculate the adsorption free energies of key reaction species (e.g.,  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , NO, N, H, NH, and  $\text{N}_2$ ) on various  $\text{Ru}_x\text{Pd}_{100-x}$  surfaces and to calculate the energetics related to  $\text{NH}_4^+$  and  $\text{N}_2$  formation, shown in Figure 4. We were thus able to examine how the thermodynamics of the competing pathways for  $\text{NH}_4^+$  formation and  $\text{N}_2$  formation, as



**Figure 4.** a) Energy requirement to move from the  $\text{N}^*$ ,  $\text{H}^*$  stage to the  $\text{NH}^*$  stage (green squares) or from the  $2\text{N}^*$  stage to the  $\text{N}_2^*$  stage (red diamonds) on a  $\text{Ru}_{100}$ ,  $\text{Ru}_{75}\text{Pd}_{25}$ ,  $\text{Ru}_{25}\text{Pd}_{75}$ , and  $\text{Pd}_{100}$  surface. The more favorable pathway is the one with a lower relative free energy for each composition. The insets show the four possible representative triatomic ensembles on the alloy surface:  $\text{Ru}_3$ ,  $\text{Ru}_2\text{Pd}_1$ ,  $\text{Ru}_1\text{Pd}_2$ , and  $\text{Pd}_3$ . b) Comparison of the binding energies of nitrate ( $\text{NO}_3^*$ ; blue circles), nitrite ( $\text{NO}_2^*$ ; green diamonds), nitric oxide ( $\text{NO}^*$ ; teal squares), and hydrogen ( $\text{H}^*$ ; red triangles) on various  $\text{Ru}_x\text{Pd}_{100-x}$  surfaces.

well as surface coverage effects, impact selectivity during  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and NO reduction over RuPdNP surfaces.

### 3.1. Thermodynamics of Competing $\text{NH}_4^+$ versus $\text{N}_2$ Formation Pathways on RuPd Surfaces

Previous theoretical studies have demonstrated that the hydrogenative reduction of  $\text{NO}_3^-$  follows a single pathway to the formation of adsorbed NO (i.e.,  $\text{NO}^*$ , where “\*” represents an adsorbed state), before splitting into the pathways leading to  $\text{NH}_4^+$  or  $\text{N}_2$ .<sup>[54,62]</sup> Here, using DFT calculations, we calculated binding energies of the  $\text{NO}^*$ ,  $\text{N}^*$ ,  $\text{H}^*$ ,  $\text{NH}^*$ , and  $\text{N}_2^*$  species on  $\text{Ru}_x\text{Pd}_{100-x}$ NP catalysts to better understand the end-product tunability of reaction selectivity during  $\text{NO}_3^-$  reduction (Scheme 1). The relative free energies of the different reaction steps were calculated on four different  $\text{Ru}_x\text{Pd}_{100-x}$ NP surfaces ( $\text{Ru}_{100}$ ,  $\text{Ru}_{75}\text{Pd}_{25}$ ,  $\text{Ru}_{25}\text{Pd}_{75}$ , and  $\text{Pd}_{100}$ ). For alloy surfaces, the overall adsorbate binding energy was calculated by taking a weighted average of the adsorbate binding energies at each of the possible triatomic ensembles (Table S8, Supporting Information). The results indicate that the most important step in determining selectivity is the progression from  $\text{N}^*$  to either  $\text{NH}^*$  or  $\text{N}_2^*$  (i.e., forming an N–H bond from  $\text{N}^*$  and  $\text{H}^*$  or forming an N–N bond from two adsorbed  $\text{N}^*$ ). Thus, comparing the N–H bond formation energy with the N–N bond formation energy may be a good descriptor of catalyst selectivity, where the lower formation energy (i.e., a more negative value meaning more exothermic or a less positive value meaning less endothermic) will be more favorable. These results are shown in Figure 4a, where the difference in binding energies of States 2 and 1 are plotted as a function of catalyst composition (for example  $[\text{E}_{\text{b},\text{NH}^*} - \text{E}_{\text{b},\text{N}^*} - \text{E}_{\text{b},\text{H}^*}]$  as shown by the green squares and  $[\text{E}_{\text{b},\text{N}_2^*} - 2\text{E}_{\text{b},\text{N}^*}]$  as shown by the red diamonds); the values used are presented in Table S9 (Supporting Information). DFT calculations show that  $\text{N}^*$  strongly binds to the pure Ru catalyst surface, and so both N–H and N–N formation are endothermic. On monometallic  $\text{Ru}_{100}$  catalysts, N–H formation requires 0.06 eV, which is much less than the

1.95 eV required for N–N formation. Thus, the formation of N–H, and eventually  $\text{NH}_4^+$ , is thermodynamically favored on a  $\text{Ru}_{100}$  catalyst. As the fraction of Ru in the catalyst decreases, the N binding energy becomes weaker (more positive), and the difference between the N–H bond formation energy and the N–N bond formation energy also decreases. Correspondingly, a  $\text{Ru}_{75}\text{Pd}_{25}$  surface is still predicted to favor N–H formation over N–N (−0.02 eV versus 0.53 eV, respectively). However, in the other extreme, the energy for N–N formation at a  $\text{Ru}_{25}\text{Pd}_{75}$  nanoparticle surface is predicted to become slightly more favorable than N–H formation (−0.45 eV versus −0.19 eV), indicating a change in overall selectivity as the amount of Pd in the catalyst increases. On a pure  $\text{Pd}_{100}$  surface,  $\text{N}^*$  binds weakly compared with  $\text{NH}^*$  and  $\text{N}_2^*$  so both N–H and N–N formation are exothermic. N–N formation is much more exothermic than N–H formation (−0.94 eV versus −0.26 eV, respectively) meaning  $\text{N}_2$  is expected to be the primary product using a pure Pd catalyst.

### 3.2. Surface Coverage Effects from the Adsorption of H, $\text{NO}_3^-$ , $\text{NO}_2^-$ , and NO

It is important to note that the calculations in Section 3.1 do not take into consideration surface coverage effects, and therefore only represent the thermodynamically favored end-product formation. We postulate that the selectivity of  $\text{Ru}_x\text{Pd}_{100-x}$ NP catalysts during  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and NO reduction may be impacted by, if not controlled by, the competitive adsorption of  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , NO, and  $\text{H}_2$ . For example, it has been proposed that increasing initial  $\text{NO}_2^-$  concentrations during reduction over pure Ru and Pd catalysts leads to decreasing activity because the  $\text{NO}_2^-$  outcompetes the  $\text{H}_2$  for adsorption onto the catalyst surface, which limits the overall reaction.<sup>[62,94]</sup> To understand how  $\text{NO}_3^-$  and  $\text{NO}_2^-$  adsorption impact selectivity during reduction starting with each of these species, we calculated the binding energies of aqueous nitrate, nitrite, NO, and H on a pure  $\text{Ru}_{100}$  surface, a  $\text{Ru}_{75}\text{Pd}_{25}$  surface, a  $\text{Ru}_{25}\text{Pd}_{75}$  surface, and a pure  $\text{Pd}_{100}$  surface, shown in Figure 4b; Table S10 (Supporting Information),

following reactions 1–4. These binding energies can influence the surface coverage of N- and H-species (and thus the N:H ratio on the surface), which in turn influences the catalyst activity and selectivity.



### 3.2.1. Hydrogen Binding

Previous computational studies of nitrite reduction over metallic catalysts have demonstrated that N binding on a metal surface is often a key descriptor in catalyst activity and selectivity, and that H binding energy, while important, is not a key determinant in reactivity or selectivity for catalytic  $\text{NO}_2^-$  reduction.<sup>[6]</sup> Our work here confirms this trend. Shown in Figure 4b by the red triangles, the overall difference in H binding between a pure Ru fcc (111) surface and a pure Pd fcc (111) surface is relatively small ( $-0.59$  eV versus  $-0.61$  eV, respectively). Additionally, it should be noted that the H binding using several pure Ru hcp slabs is also similar (approximately  $-0.6$  eV). Further, our calculations show that alloying Ru and Pd does not significantly impact the H binding energy (Table S10, Supporting Information), which is reasonable, since both Ru and Pd have a high proficiency for dissociative hydrogen activation.<sup>[95,96]</sup> When compared to the binding energy trends of nitrogen-containing species, the relative consistency of H binding indicates that while H binding to the catalyst surface is important, dissociative  $\text{H}_2$  binding alone is not a key determining factor for overall selectivity. This is further evidenced by the experimental setup: in all cases the experimental system is sparged with hydrogen gas so that the catalyst surface is anticipated to be covered with a layer of adsorbed H\* atoms when the  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , or NO is introduced.

### 3.2.2. Comparison of Nitrate, Nitrite, and NO Binding Energies to H

Figure 4b shows that  $\text{NO}_3^-$  (blue circles) binds weaker to all catalysts than H (red triangles), indicating that  $\text{NO}_3^-$  should not outcompete H for available surfaces sites and nitrate reduction is not limited by a lack of H\* atoms on the catalyst. On Pd-rich surfaces (pure Pd<sub>100</sub> and Ru<sub>25</sub>Pd<sub>75</sub>) the nitrate binding energy is positive (0.42 eV and 0.03 eV, respectively) indicating a weak binding, so nitrate is unable to bind to the catalyst surface or binds very weakly, resulting in negligible nitrate reduction activity for Pd-rich catalysts on the extreme end of the composition range and low activity for intermediate Pd-rich catalysts (Ru<sub>30</sub>Pd<sub>70</sub>NPs). On Ru-rich surfaces (pure Ru<sub>100</sub> and Ru<sub>75</sub>Pd<sub>25</sub>), nitrate binding energy is favorable (negative) but weaker than H binding ( $-0.28$  eV on Ru<sub>100</sub> and  $-0.33$  eV on Ru<sub>75</sub>Pd<sub>25</sub>). Thus,  $\text{NO}_3^-$  can bind to surface sites to initiate reduction, but a relatively high H\* coverage is expected which contributes to the good performance of Ru-rich catalysts for reducing nitrate. Hence, Ru-rich catalysts have

higher activity for  $\text{NO}_3^-$  reduction than catalysts with a moderate or low amount of Ru.

As shown in Figure 4b,  $\text{NO}_2^-$  (green diamonds) binds more strongly to a pure Ru<sub>100</sub> surface ( $-0.94$  eV) and a Ru<sub>75</sub>Pd<sub>25</sub> surface ( $-0.86$  eV) than H (red triangles). This strong binding can poison the catalyst surface with nitrite species leading to both decreased activity and  $\text{NH}_4^+$  selectivity due to a lack of H\* atoms (resulting in a high N:H ratio). On the other hand, nitrite binds weaker to a pure Pd<sub>100</sub> surface (0.02 eV) and a Ru<sub>25</sub>Pd<sub>75</sub> surface ( $-0.44$  eV) than H, resulting in more H\* atoms on the surface to carry out nitrite reduction. Interestingly, Figure 4b also reveals that NO (teal squares) binds much more strongly than H. NO binding is weakest on a pure Pd<sub>100</sub> surface ( $-1.32$  eV) and increases once Ru is introduced. NO binding is strongest on the Ru<sub>75</sub>Pd<sub>25</sub> surface ( $-2.53$  eV), and only slightly weaker on a pure Ru<sub>100</sub> surface ( $-2.48$  eV). This suggests that NO should strongly outcompete H for available surface sites, effectively poisoning the catalysts and leading to low NO reduction activity, similar to what is predicted for  $\text{NO}_2^-$ . However, this is not observed experimentally. We believe this is because NO was introduced as a gas, which introduces gas-to-liquid phase mass transfer limitations that experimentally limit the adsorption of NO to the catalysts, thereby lowering the N:H ratio on the catalyst surface. This limitation of NO adsorption results in higher apparent NO reduction activity.

## 4. Discussion

### 4.1. Selectivity During $\text{NO}_3^-$ and NO Reduction is Primarily Controlled by Thermodynamics

Our experimental results for  $\text{NO}_3^-$  reduction, where  $\text{NH}_4^+$  is predominantly formed by Ru-rich catalysts, and  $\text{N}_2$  is formed in significant amounts only when the Ru composition is  $< 50\%$  (Figure 3d) are in good agreement with the thermodynamic behavior of  $\text{NH}_4^+$  and  $\text{N}_2$  production as calculated by DFT (Figure 4a). At 50% loss of nitrate, all Ru-rich catalysts (i.e.,  $x \geq 50$ ) displayed over 90% selectivity for  $\text{NH}_4^+$  production. Based on our DFT results, the formation of N–H from adsorbed N and adsorbed H is much more favorable than N–N formation on these catalysts. Additionally, the surface coverage effects contribute to the  $\text{NH}_4^+$  selectivity. The relatively weak binding of  $\text{NO}_3^-$  as compared to H will mean there is a relatively low N:H ratio on the surface making N–H formation statistically more likely to occur. Thus, both thermodynamics and surface coverage result in high  $\text{NH}_4^+$  selectivity during  $\text{NO}_3^-$  reduction on Ru-rich catalysts. As we shift to more Pd-rich catalysts (i.e.,  $x < 50$ ), the selectivity for  $\text{NH}_4^+$  at 50% loss of  $\text{NO}_3^-$  decreases and  $\text{N}_2$  is produced in appreciable amounts (29.2%  $\text{N}_2$  for Ru<sub>30</sub>Pd<sub>70</sub>NPs and 37.7%  $\text{N}_2$  for Ru<sub>10</sub>Pd<sub>90</sub>NPs), though the majority of the end-product is still  $\text{NH}_4^+$ . The relatively weak  $\text{NO}_3^-$  binding (0.03 eV) on a Pd-rich surface indicates that though  $\text{NO}_3^-$  will adsorb to these surfaces, it is difficult to do so. This weak binding will result in a low N:H ratio on the catalyst surface, and the abundance of H atoms and lack of N atoms make N–H formation more likely than N–N formation, even though N–N is thermodynamically favored. Note that since a pure Pd catalyst cannot reduce  $\text{NO}_3^-$  on its own, we cannot compare selectivity experiments with our theoretical calculations for  $\text{NO}_3^-$  reduction. When we look at

selectivity at complete conversion of nitrate, we see a similar trend as at 50% loss of nitrate. All of the Ru-rich catalysts again show near-complete selectivity for  $\text{NH}_4^+$  production (> 95%), and even the  $\text{Ru}_{30}\text{Pd}_{70}$ NPs show 93.7% selectivity for  $\text{NH}_4^+$ . As we approach complete conversion of  $\text{NO}_3^-$  using  $\text{Ru}_{30}\text{Pd}_{70}$ NPs, the surface coverage effects become even more important. Previous work by Zhao, et al. demonstrated that as hydrogenation approaches full conversion, the ratio of adsorbed N:H becomes even lower than at initial stages of reaction.<sup>[55]</sup> This decreasing N:H ratio further emphasizes the surface coverage effects, making N–H formation even more likely to occur, especially since both N–N and N–H formation are exothermic on Pd-rich surfaces. Thus, the  $\text{NH}_4^+$  selectivity of all the catalysts increases as  $\text{NO}_3^-$  reduction approaches complete conversion, and more so for the Pd-rich catalysts than the Ru-rich catalysts.

Generally, the selectivity behavior of  $\text{Ru}_x\text{Pd}_{100-x}$ NPs during NO reduction (Figure 3f) follows that of  $\text{NO}_3^-$  reduction (Figure 3d). Despite the very strong binding energies of NO, the adsorption of NO to the catalyst surface is limited by mass transfer from the gas phase to the aqueous phase, and so thermodynamics typically dominate the selectivity. Throughout the entire range of alloy compositions, as the amount of Ru increases, the  $\text{NH}_4^+$  selectivity increases following the trend of N–N formation being more favored than N–H formation on Pd-rich surfaces but switching to N–H formation being more favored on Ru-rich surfaces. However, drawing firm conclusions about selectivity behavior is difficult because of the problems that arise from the NO feedstock being introduced as a gas, versus an aqueous species like  $\text{NO}_3^-$ .

#### 4.2. Selectivity During $\text{NO}_2^-$ Reduction is Controlled by both Thermodynamics and Surface Coverage Effects

While the selectivity behavior of  $\text{Ru}_x\text{Pd}_{100-x}$ NPs during  $\text{NO}_3^-$  is well explained by  $\text{NH}^*$  and  $\text{N}_2^*$  formation, there is a deviation between the selectivity during  $\text{NO}_2^-$  reduction (Figure 3e) and the thermodynamics of the formation pathways. This deviation stems from the surface coverage effects demonstrated by the binding energies of  $\text{NO}_2^-$  and H. As mentioned above, the activities of the  $\text{Ru}_x\text{Pd}_{100-x}$ NPs during  $\text{NO}_2^-$  are also controlled by  $\text{NO}_2^-$  and H binding. For pure Ru and very Ru-rich surfaces,  $\text{NO}_2^-$  binds more strongly than H. This strong  $\text{NO}_2^-$  binding poisons the catalyst surface, slowing activity. Additionally, because the  $\text{NO}_2^-$  outcompetes the H for available surface sites, there is a very high ratio of N:H on the catalyst surface for Ru-rich catalysts. Thus, even though the thermodynamics would favor N–H formation, the lack of available adsorbed H means that N–N formation is more likely on the Ru-rich surfaces. When we look at the behavior of the  $\text{Ru}_{90}\text{Pd}_{10}$ NPs, this is exactly what we see: 32.7% selectivity for  $\text{NH}_4^+$ . However, as we shift to more intermediate compositions (i.e.,  $x = 50$  and 70), the binding energy of  $\text{NO}_2^-$  becomes weaker while the binding energy of H stays relatively constant. Thus,  $\text{NO}_2^-$  does not as strongly outcompete H for available surface sites, and so the effect of surface coverage decreases and the effect of thermodynamics increase. Thus, the  $\text{NH}_4^+$  selectivity increases moving from  $\text{Ru}_{90}\text{Pd}_{10}$ NPs to  $\text{Ru}_{70}\text{Pd}_{30}$ NPs to  $\text{Ru}_{50}\text{Pd}_{50}$ NPs. As we move to intermediate Pd-rich and very Pd-rich catalysts, the binding energy of  $\text{NO}_2^-$  continues to weaken;

on a  $\text{Ru}_{25}\text{Pd}_{75}$  surface, the calculated binding energy of  $\text{NO}_2^-$  is only slightly weaker than the binding energy of H, so we would expect a slightly higher H surface coverage than  $\text{NO}_2^-$  surface coverage. Additionally, the thermodynamically favored pathway switches, and N–N formation becomes more favored than N–H formation. Consequently, the selectivity of  $\text{Ru}_{30}\text{Pd}_{70}$ NPs is a product of both effects, which is evidenced by the experimental results: 78.4%  $\text{NH}_4^+$  and 21.6%  $\text{N}_2$ . As we continue to become more Pd-rich, the binding of  $\text{NO}_2^-$  becomes even weaker, but N–N formation also becomes much more favorable, and thermodynamics become more dominant resulting in a lower selectivity for  $\text{NH}_4^+$  production.

#### 5. Conclusion

We have shown that the end-product selectivity of aqueous  $\text{NO}_3^-$  reduction can be effectively and broadly controlled by varying the ratio of Ru:Pd in randomly alloyed  $\text{Ru}_x\text{Pd}_{100-x}$ NP catalysts, supported on amorphous  $\text{SiO}_2$ . Although Ru and Pd are considered to be classically immiscible at room temperature across all compositions in bulk, metastable nanoalloys of Ru and Pd were confirmed using a variety of complementary spectroscopic techniques. Experimentally, we demonstrated that the selectivity of  $\text{NO}_3^-$  and NO reduction can be tuned by controlling the composition of RuPdNPs; specifically, a Ru-rich composition favors the formation of  $\text{NH}_4^+$  and a Pd-rich composition favors  $\text{N}_2$  formation. These results are supported by extensive DFT calculations of the thermodynamics of the competing pathways for  $\text{NH}_4^+$  and  $\text{N}_2$ . However, the selectivity during  $\text{NO}_2^-$  reduction follows a slightly different trend where  $\text{NH}_4^+$  selectivity peaks at  $\approx 50\%$  Ru. DFT calculations indicate this behavior is because the surface coverage of reactants plays a larger role than the thermodynamics of competing pathways during  $\text{NO}_2^-$  reduction compared to  $\text{NO}_3^-$  and NO reduction. Specifically, Ru-rich surfaces are susceptible to poisoning by  $\text{NO}_2^-$  coverage resulting in reduced activity and higher  $\text{N}_2$  selectivity. Overall, in agreement with our original hypothesis, we prove that one must consider both the thermodynamics of competing pathways and the surface coverages of starting reactants when considering catalyst design for nitrate reduction. This provides an important lesson for those who pursue the future discovery of other new catalysts for industrially relevant reactions, where end-product selectivity is a critically important consideration.

#### Supporting Information

Supporting Information is available from the Wiley Online Library or from the author.

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## Conflict of Interest

The authors declare no conflict of interest.

## Data Availability Statement

The data that support the findings of this study are available from the corresponding author upon reasonable request.

## Keywords

density functional theory, microwave synthesis, nitrate reduction, noble metal catalysts, ruthenium-palladium alloys

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- [1] J. H. Clark, *Green Chem.* **1999**, *1*, 1.
- [2] G. A. Somorjai, R. M. Rioux, *Catal. Today.* **2005**, *100*, 201.
- [3] G. A. Somorjai, J. Y. Park, *Angew. Chem., Int. Ed.* **2008**, *47*, 9212.
- [4] I. Lee, M. A. Albiter, Q. Zhang, J. Ge, Y. Yin, F. Zaera, *Phys. Chem. Chem. Phys.* **2011**, *13*, 2449.
- [5] F. Zaera, *J. Phys. Chem. B.* **2002**, *106*, 4043.
- [6] H. Li, S. Guo, K. Shin, M. S. Wong, G. Henkelman, *ACS Catal.* **2019**, *9*, 7957.
- [7] H. Li, C. Yan, H. Guo, K. Shin, S. M. Humphrey, C. J. Werth, G. Henkelman, *ACS Catal.* **2020**, *10*, 7915.
- [8] H. Guo, J. A. Trindell, H. Li, D. Fernandez, S. M. Humphrey, G. Henkelman, R. M. Crooks, *J. Mater. Chem. A* **2020**, *8*, 8421.
- [9] S. M. Davis, F. Zaera, G. A. Somorjai, *J. Am. Chem. Soc.* **1982**, *104*, 7453.
- [10] P. Christopher, S. Linic, *AIChE Annu. Meet. Conf. Proc.* **2008**, 11264.
- [11] J. B. Ernst, S. Muratsugu, F. Wang, M. Tada, F. Glorius, *J. Am. Chem. Soc.* **2016**, *138*, 10718.
- [12] A. Fedorov, H. J. Liu, H. K. Lo, C. Copéret, *J. Am. Chem. Soc.* **2016**, *138*, 16502.
- [13] L. J. Durndell, C. M. A. Parlett, N. S. Hondow, M. A. Isaacs, K. Wilson, A. F. Lee, *Sci. Rep.* **2015**, *5*, 9425.
- [14] E. H. Boymans, P. T. Witte, D. Vogt, *Catal. Sci. Technol.* **2015**, *5*, 176.
- [15] M. J. Taylor, L. J. Durndell, M. A. Isaacs, C. M. A. Parlett, K. Wilson, A. F. Lee, G. Kyriakou, *Appl. Catal. B.* **2016**, *180*, 580.
- [16] B. Chen, F. Li, Z. Huang, G. Yuan, *Appl. Catal. B.* **2017**, *200*, 192.
- [17] A. Liu, J. Ming, R. O. Ankumah, *Sci. Total Environ.* **2005**, *346*, 112.
- [18] Y. Zhang, P. Shi, J. Song, Q. Li, *Appl. Sci.* **2018**, *9*, 18.
- [19] R. Biddau, R. Cidu, S. Da Pelo, A. Carletti, G. Ghiglieri, D. Pittalis, *Sci. Total Environ.* **2019**, *647*, 1121.
- [20] N. Kazakis, I. Matiatos, M. M. Ntona, M. Bannenberg, K. Kalaitzidou, E. Kaprara, M. Mitrakas, A. Ioannidou, G. Vargemezis, K. Voudouris, *Sci. Total Environ.* **2020**, *724*, 138211.
- [21] B. T. Nolan, J. D. Stoner, *Environ. Sci. Technol.* **2000**, *34*, 1156.
- [22] L. J. Puckett, *Environ. Sci. Technol.* **1995**, *29*, 408A.
- [23] W. R. Raun, G. V. Johnson, *Agron. J.* **1999**, *91*, 357.
- [24] L. Lassaletta, G. Billen, B. Grizzetti, J. Anglade, J. Garnier, *Environ. Res. Lett.* **2014**, *9*, 105011.
- [25] K. R. Burow, B. T. Nolan, M. G. Rupert, N. M. Dubrovsky, *Environ. Sci. Technol.* **2010**, *44*, 4988.
- [26] L. Pelletier, A. McFarlan, N. Maffei, *J. Power Sources.* **2005**, *145*, 262.
- [27] C. Zamfirescu, I. Dincer, *J. Power Sources.* **2008**, *185*, 459.
- [28] A. G. M. Da Silva, T. S. Rodrigues, S. J. Haigh, P. H. C. Camargo, *Chem. Commun.* **2017**, *53*, 7135.
- [29] P. H. Pfromm, *J. Renew. Sustain. Energy.* **2017**, *9*, 034702.
- [30] M. J. Palys, H. Wang, Q. Zhang, P. Daoutidis, *Curr. Opin. Chem. Eng.* **2021**, *31*, 100667.
- [31] J. W. Erisman, M. A. Sutton, J. Galloway, Z. Klimont, W. Winiwarter, *Nat. Geosci.* **2008**, *1*, 636.
- [32] S. L. Foster, S. I. P. Bakovic, R. D. Duda, S. Maheshwari, R. D. Milton, S. D. Minteer, M. J. Janik, J. N. Renner, L. F. Greenlee, *Nat. Catal.* **2018**, *1*, 490.
- [33] L. Fewtrell, *Environ. Health Perspect.* **2004**, *112*, 1371.
- [34] P. J. Weyer, J. R. Cerhan, B. C. Kross, G. R. Hallberg, J. Kantamneni, G. Breuer, M. P. Jones, W. Zheng, C. F. Lynch, *Epidemiology.* **2001**, *12*, 327.
- [35] S. M. A. Adelana, *Water Encyclopedia.* **2005**, *4*, 30.
- [36] A. R. Townsend, R. W. Howarth, F. A. Bazzaz, M. S. Booth, C. C. Cleveland, S. K. Collinge, A. P. Dobson, P. R. Epstein, E. A. Holland, D. R. Keeney, M. A. Mallin, C. A. Rogers, P. Wayne, A. H. Wolfe, *Front. Ecol. Environ.* **2003**, *1*, 240.
- [37] C. J. Werth, C. Yan, J. P. Troutman, *ACS ES&T Eng.* **2021**, *1*, 6.
- [38] M. Duca, M. T. M. Koper, *Energy Environ. Sci.* **2012**, *5*, 9726.
- [39] H. Xu, Y. Ma, J. Chen, W. X. Zhang, J. Yang, *Chem. Soc. Rev.* **2022**, *51*, 2710.
- [40] G. Yang, P. Zhou, J. Liang, H. Li, F. Wang, *Inorg. Chem. Front.* **2023**, *10*, 4610.
- [41] S. Garcia-Segura, M. Lanzarini-Lopes, K. Hristovski, P. Westerhoff, *Appl. Catal. B.* **2018**, *236*, 546.
- [42] Z. Wang, D. Richards, N. Singh, *Catal. Sci. Technol.* **2021**, *11*, 705.
- [43] M. Duca, M. C. Figueiredo, V. Climent, P. Rodriguez, J. M. Feliu, M. T. M. Koper, *J. Am. Chem. Soc.* **2011**, *133*, 10928.
- [44] S.-E. Bae, K. L. Stewart, A. A. Gewirth, *J. Am. Chem. Soc.* **2007**, *129*, 10171.
- [45] T. Chen, H. Li, H. Ma, M. T. M. Koper, *Langmuir.* **2015**, *31*, 3277.
- [46] Z. Wang, S. D. Young, B. R. Goldsmith, N. Singh, *J. Catal.* **2021**, *395*, 143.
- [47] H. Begum, M. N. Islam, S. Ben Aoun, J. A. Safwan, S. S. Shah, M. A. Aziz, M. A. Hasnat, *Environ. Sci. Pollut. Res.* **2023**, *30*, 34904.
- [48] M. N. Islam, M. Ahsan, K. Aoki, Y. Nagao, A. E. Alsafrani, H. M. Marwani, A. Almahri, M. M. Rahman, M. A. Hasnat, *J. Environ. Chem. Eng.* **2023**, *11*, 111149.
- [49] X. Wu, M. Nazemi, S. Gupta, A. Chismar, K. Hong, H. Jacobs, W. Zhang, K. Rigby, T. Hedtke, Q. Wang, E. Stavitski, M. S. Wong, C. Muhich, J. H. Kim, *ACS Catal.* **2023**, *13*, 6804.
- [50] S. Guo, H. Li, K. N. Heck, X. Luan, W. Guo, G. Henkelman, M. S. Wong, *Appl. Catal. B.* **2022**, *305*, 121048.
- [51] S. Guo, K. Heck, S. Kasiraju, H. Qian, Z. Zhao, L. C. Grabow, T. Miller, M. S. Wong, *ACS Catal.* **2018**, *8*, 503.
- [52] S. Hörold, K. D. Vorlop, T. Tacke, M. Sell, *Catal. Today.* **1993**, *17*, 21.
- [53] S. D. Ebbesen, B. L. Mojet, L. Lefferts, *J. Catal.* **2008**, *256*, 15.
- [54] H. Shin, S. Jung, S. Bae, W. Lee, H. Kim, *Environ. Sci. Technol.* **2014**, *48*, 12768.
- [55] Y. Zhao, N. Koteshwara Rao, L. Lefferts, *J. Catal.* **2016**, *337*, 102.
- [56] K. Vorlop, T. Tacke, *Chemie Ing. Tech.* **1989**, *61*, 836.
- [57] S. Hörold, T. Tacke, K. D. Vorlop, *Environ. Technol.* **1993**, *14*, 931.
- [58] WHO, *Guidelines for Drinking-Water Quality*, World Health Organization, , 2017.
- [59] E. J. Bouwer, P. B. Crowe, *J. Am. Water Work. Assoc.* **1988**, *80*, 82.

- [60] R. G. Perkins, E. I. Slavin, T. M. C. Andrade, C. Blenkinsopp, P. Pearson, T. Froggatt, G. Godwin, J. Parslow, S. Hurley, R. Luckwell, D. J. Wain, *J. Environ. Manage.* **2019**, *244*, 276.
- [61] L. Wei, D.-J. Liu, B. A. Rosales, J. W. Evans, J. Vela, *ACS Catal.* **2020**, *10*, 3618.
- [62] X. Huo, D. J. Van Hoomissen, J. Liu, S. Vyas, T. J. Strathmann, *Appl. Catal. B* **2017**, *211*, 188.
- [63] K. Kusada, H. Kobayashi, R. Ikeda, Y. Kubota, M. Takata, S. Toh, T. Yamamoto, S. Matsumura, N. Sumi, K. Sato, K. Nagaoka, H. Kitagawa, *J. Am. Chem. Soc.* **2014**, *136*, 1864.
- [64] N. Dahal, S. García, J. Zhou, S. M. Humphrey, *ACS Nano* **2012**, *6*, 9433.
- [65] S. García, L. Zhang, G. W. Piburn, G. Henkelman, S. M. Humphrey, *ACS Nano* **2014**, *8*, 11512.
- [66] K. Kusada, H. Kobayashi, T. Yamamoto, S. Matsumura, N. Sumi, K. Sato, K. Nagaoka, Y. Kubota, H. Kitagawa, *J. Am. Chem. Soc.* **2013**, *135*, 5493.
- [67] Y. Nanba, T. Ishimoto, M. Koyama, *J. Phys. Chem. C* **2017**, *121*, 27445.
- [68] H. Lin, H. Lin, J. X. Liu, H. Fan, W. X. Li, *J. Phys. Chem. C* **2020**, *124*, 11005.
- [69] Q. Zhang, K. Kusada, D. Wu, T. Yamamoto, T. Toriyama, S. Matsumura, S. Kawaguchi, Y. Kubota, H. Kitagawa, *Nat. Commun.* **2018**, *9*, 510.
- [70] B. Huang, H. Kobayashi, T. Yamamoto, S. Matsumura, Y. Nishida, K. Sato, K. Nagaoka, M. Haneda, S. Kawaguchi, Y. Kubota, H. Kitagawa, *Chem. Sci.* **2020**, *11*, 11413.
- [71] K. Sato, H. Tomonaga, T. Yamamoto, S. Matsumura, N. D. B. Zulkifli, T. Ishimoto, M. Koyama, K. Kusada, H. Kobayashi, H. Kitagawa, K. Nagaoka, *Sci. Rep.* **2016**, *6*, 28265.
- [72] C. Song, A. Tayal, O. Seo, J. Kim, Y. Chen, S. Hiroi, L. S. R. Kumara, K. Kusada, H. Kobayashi, H. Kitagawa, O. Sakata, *Nanoscale Adv.* **2019**, *1*, 546.
- [73] J. W. M. Frenken, P. Stoltze, *Phys. Rev. Lett.* **1999**, *82*, 3500.
- [74] E. Ringe, R. P. Van Duyne, L. D. Marks, *Nano Lett.* **2011**, *11*, 3399.
- [75] S. N. Tripathi, S. R. Bharadwaj, S. R. Dharwadkar, *J. Phase Equilib.* **1993**, *14*, 638.
- [76] D. C. Harris, *Quantitative Chemical Analysis*, W. H. Freeman And Company, New York, NY, **2007**.
- [77] H. Li, W. Chai, G. Henkelman, *J. Mater. Chem. A* **2019**, *7*, 23868.
- [78] C. D. Wagner, A. V. Naumkin, A. Kraut-Vass, J. W. Allison, C. J. Powell, J. R. Rumble, "NIST X-ray Photoelectron Spectroscopy Database, Version 4.1," DOI 10.18434/T4T88K can be found under <https://srdata.nist.gov/xps>, 2012.
- [79] D. J. Morgan, *Surf. Interface Anal.* **2015**, *47*, 1072.
- [80] S. Seraj, P. Kunal, H. Li, G. Henkelman, S. M. Humphrey, C. J. Werth, *ACS Catal.* **2017**, *7*, 3268.
- [81] P. Kunal, H. Li, B. L. Dewing, L. Zhang, K. Jarvis, G. Henkelman, S. M. Humphrey, *ACS Catal.* **2016**, *6*, 4882.
- [82] J. P. Troutman, H. Li, A. M. Haddix, B. A. Kienzle, G. Henkelman, S. M. Humphrey, C. J. Werth, *ACS Catal.* **2020**, *10*, 7979.
- [83] O. S. G. P. Soares, J. J. M. Órfão, J. Ruiz-Martínez, J. Silvestre-Albero, A. Sepúlveda-Escribano, M. F. R. Pereira, *Chem. Eng. J.* **2010**, *165*, 78.
- [84] R. L. Vander Wal, T. M. Ticich, V. E. Curtis, *Carbon* **2001**, *39*, 2277.
- [85] M. E. Strayer, J. M. Binz, M. Tanase, S. M. Kamali Shahri, R. Sharma, R. M. Rioux, T. E. Mallouk, *J. Am. Chem. Soc.* **2014**, *136*, 5687.
- [86] F. Epron, F. Gauthard, J. Barbier, *J. Catal.* **2002**, *206*, 363.
- [87] N. Krawczyk, S. Karski, I. Witońska, *React. Kinet. Mech. Catal.* **2011**, *103*, 311.
- [88] B. P. Chaplin, M. Reinhard, W. F. Schneider, C. Schüth, J. R. Shapley, T. J. Strathmann, C. J. Werth, *Environ. Sci. Technol.* **2012**, *46*, 3655.
- [89] U. Prütse, K. D. Vorlop, *J. Mol. Catal. A: Chem.* **2001**, *173*, 313.
- [90] A. Pintar, M. Šetinc, J. Levec, *J. Catal.* **1998**, *174*, 72.
- [91] F. A. Marchesini, S. Irusta, C. Querini, E. Miró, *Appl. Catal. A* **2008**, *348*, 60.
- [92] U. Prütse, M. Hähnlein, J. Daum, K. D. Vorlop, *Catal. Today* **2000**, *55*, 79.
- [93] R. Zhang, D. Shuai, K. A. Guy, J. R. Shapley, T. J. Strathmann, C. J. Werth, *ChemCatChem* **2013**, *5*, 313.
- [94] R. S. Postma, R. Brunet Espinosa, L. Lefferts, *ChemCatChem* **2018**, *10*, 3770.
- [95] F. Engelke, R. Vincent, T. S. King, M. Pruski, *J. Chem. Phys.* **1994**, *101*, 7262.
- [96] H. Zheng, H. Li, W. Song, Z. Zhao, G. Henkelman, *J. Phys. Chem. C* **2021**, *125*, 12028.