

Rotating Maize Reduces the Risk and Rate of Nitrate Leaching

HR Pasley^{1*}, VA Nichols¹, MJ Castellano¹, MJ Helmers², ME Baum¹, EJ Kladvko³, SV Archontoulis¹

¹ Department of Agronomy, Iowa State University, Ames, Iowa

² Department of Agricultural and Biosystems Engineering, Iowa State University, Ames, Iowa

³ Department of Agronomy, Purdue University, West Lafayette, Indiana

*Correspondence:

Dr. Heather R. Pasley

Heather.pasley@csiro.au

Abstract

There is a strong link between nitrate ($\text{NO}_3\text{-N}$) leaching from fertilized annual crops and the rate of nitrogen (N) fertilizer input. However, this leaching-fertilizer relationship is poorly understood and the degree to which soil type, weather, and cropping system influence it is largely unknown. We calibrated the APSIM process-based cropping system model using 56 site-years of data sourced from eight field studies across six states in the U.S. Midwest that monitored $\text{NO}_3\text{-N}$ leaching from artificial subsurface drainage in two cropping systems: continuous maize and two-year rotation of maize followed by unfertilized soybean (maize-soybean rotation). We then ran a factorial simulation experiment and fit statistical models to the leaching-fertilizer response. A bi-linear model provided the best fit to the relationship between N fertilizer rate (kg/ha) and $\text{NO}_3\text{-N}$ leaching load (kg/ha) (from one year of continuous maize or summed over the two-year maize-soybean rotation). We found that the cropping system dictated the slopes and breakpoint (the point at which the leaching rate changes) of the model, but the site and year determined the intercept i.e. the magnitude of the leaching. In both cropping systems, the rate of $\text{NO}_3\text{-N}$ leaching increased at an N fertilizer rate higher than the N rate needed to optimize the leaching load per kg grain produced. Above the model breakpoint, the rate of $\text{NO}_3\text{-N}$ leaching per kg N fertilizer input was 300% greater than the rate below the breakpoint in the two-year maize-soybean rotation and 650% greater in continuous maize. Moreover, the model breakpoint occurred at only 16% above the average agronomic optimum N rate (AONR) in continuous maize, but 66% above the AONR in the maize-soybean rotation. Rotating maize with soybean, therefore, allows for a greater environmental buffer than continuous maize with regard to the impact of overfertilization on $\text{NO}_3\text{-N}$ leaching.

Key Words

Nitrate leaching, modeling, APSIM, crop rotation, yield-scaled leaching

1. Introduction

Globally, nitrogen (N) fertilization has been a crucial variable in increasing crop yields. This increase, however, comes at a high environmental cost, considering that around 15% of the N fertilizer applied to maize (*Zea mays* L.) leaches into the groundwater as nitrate ($\text{NO}_3\text{-N}$) [1-5].

There is inherent risk to the environment in applying N fertilizer at any rate as the ecosystem processes that determine the fate of N fertilizer are both complex and dynamic. Within a given field and season, however, there is a threshold N rate or “breakpoint” above which the risk

increases substantially [1,6-13]. The factors that determine the N rate at which that breakpoint occurs are not well understood as the $\text{NO}_3\text{-N}$ leached each year is not only derived from that year's N inputs, but also from soil organic matter as well as residual N fertilizer from previous applications and returned crop residue from previous seasons [14,15].

Nevertheless, it has been widely conjectured that the breakpoint occurs at or around the yield-optimizing N rate (known as the Agronomic Optimal N Rate or AONR) [1,6,12,16]. The ecosystem dynamics that determine the AONR are similar to those of the breakpoint in that they are the result of complex interactions with the weather, soil, and residue from the previously planted crop [17,18]. Therefore, as both the breakpoint and the AONR are products of complex systems, their relationship may not be as simple as previously conceived.

The aim of this study was to determine how the response of $\text{NO}_3\text{-N}$ leaching to increases in the N fertilizer rate relates to the AONR. This relationship can both define the environmental costs of over-fertilization and provide targeted guidance to improve management strategies.

The soil N storage capacity provides a proportional buffer between the AONR and the leaching breakpoint [19,20]. Any analysis of the leaching breakpoint-AONR relationship, therefore, needs to be conducted across a wide range of environments and management variables as differences in soil characteristics, climate, tillage, and cropping system may be influencing both the AONR and the leaching breakpoint and, thus, the size of the corresponding buffer [6]. Understanding how and why the breakpoint may differ across sites is also important when considering adjustments to N fertilizer recommendations to mitigate $\text{NO}_3\text{-N}$ leaching at a regional scale [21,22].

As $\text{NO}_3\text{-N}$ leaching originates from residual soil N, mineralized organic matter, and additional applications of N fertilizer, it is likely that the leaching breakpoint in a given season is influenced by the previously planted crop and management, that is the cropping system. The two cropping systems that account for about 72% of cultivated land in the top 12 contiguous maize-producing states in the U.S. (comprising the U.S. Midwest) are continuous maize and a two-year rotation of maize with soybean in which, typically, no N fertilizer is applied to the soybean phase [23,24]. Maize rotated with soybean yields 15% higher than continuous maize despite receiving 30% less N fertilizer input, raising questions about how the fate of N inputs differs between the two systems [25]. Moreover, the organic N inputs, both from previous crop residue and soybean N fixation may have an additive effect on how much N is lost from N fertilizer. Previous studies that have looked at system effects on leaching are often short-term (three to four years) and have noted that extreme variability in seasonal precipitation appears to influence the fate of $\text{NO}_3\text{-N}$ in the soil profile, particularly during a soybean year, potentially masking any differences in the relationship between fertilizer N rate and $\text{NO}_3\text{-N}$ leaching load in these two cropping systems [11,13,26-30]. There is need to expand these findings to understand if and how cropping system selection alters the impact N fertilizer rate on $\text{NO}_3\text{-N}$ leaching and if this relationship is consistent across multiple environments (climate and soil).

Cropping system models can successfully simulate $\text{NO}_3\text{-N}$ leaching [31]. These process-based models can also investigate the factors driving $\text{NO}_3\text{-N}$ leaching independent of each other without losing the complexity of the full cropping system. Specifically, Agricultural Production Systems sIMulator (APSIM) ([32]; www.apsim.info) has been found to accurately predict both AONR [18] and $\text{NO}_3\text{-N}$ leaching [15] in the U.S. Midwest. In a 2014 analysis, Laan et al. [31] called upon researchers to use such process-based models in conjunction with data from multi-

year field studies to improve our understanding of the processes driving NO₃-N leaching. Roberts et al. [33] found that using a statistical model in combination with a process-based model significantly improved the accuracy and scope of the process-based model's predictions. Therefore, in this study, by fitting a statistical model to the outputs from APSIM, we distilled APSIM's complexity into select parameters that we then used to answer two targeted questions: (1) is the leaching breakpoint related to the AONR? (2) does this relationship differ with site location and/or cropping system and, if so, what are the practical implications for farmers?

2. Methods and Materials

2.1 Model Description

2.1.1 The APSIM Model

The APSIM model is an open source field-scale cropping systems modeling platform capable of simulating soil-crop-atmospheric interactions [32]. The platform is customizable to different environments and management strategies and can be run sequentially to capture multi-year legacy effects [34]. It comprises of integrated crop and soil modules that accurately simulate water, temperature, carbon (C), and N dynamics [35-37]. In this project, we used a modified APSIM version 7.9 with waterlogging capabilities [38-39].

2.1.2 Model Setup and Calibration

We calibrated the APSIM model using 56 site-years of NO₃-N leaching data sourced from eight artificially subsurface-drained field experiments located in the U.S. Midwest (Figure 1; study details can be found in Supplementary Table 1). These experiments measured leaching as well as yield and/or drainage responses to changes in cropping system, tillage, fertilizer N rate, and subsurface drain spacing. As there is a wide variety of approaches to measuring drainage flow and NO₃-N leaching, we selected only studies that used artificial subsurface drain outlets to gather their data rather than lysimeters or suction cups in order to maintain consistency in the data across sites. The other criteria in study selection were that they were multi-year experiments and applied inorganic (not manure) N fertilizer with a single late winter/spring application. All experiments were rain-fed systems. The research design details and data used to model these experiments were primarily sourced from their corresponding publications [11, 40-47]. Input to the model included field-specific soil and weather data (daily air temperature, precipitation, and solar radiation) extracted from publications or public soil-weather sources ([48-50], <https://power.larc.nasa.gov/>). The soil profiles and weather summaries are provided in the supplementary materials (Supplementary Tables 2-3). The model was calibrated using the management scheme of the original study (i.e. cropping system, N source/timing/rate, tillage, subsurface drain depth/spacing etc.).

2.1.3 Model Simulation Experiment

Two cropping systems were considered: continuous maize and maize rotated with unfertilized soybean. For the two-year rotation, the model was run twice: one set of simulations with maize followed by soybean and the other with soybean followed by maize such that we could compare the outputs from rotated maize (maize in the two-year maize-soybean rotation) with continuous maize each year.

For each site and cropping system combination, the calibrated model was run for five years (1994-1999) with standard management per location in the corresponding rotation to stabilize

fast decomposing N and C pools and estimate initial conditions [51-53]. The model was then run for 20 years (2000-2019) at seven different N fertilizer rates (0, 56, 112, 168, 224, 250, and 300 kg N/ha). Maize and soybean cultivars were sourced from the APSIM database with different maturity groups per location (Supplementary Tables 4-5; [54]). The USDA-NASS 50% planting dates for each respective state were used for each season and simulation [55]. Tillage, N fertilizer source/timing, subsurface drain depth/spacing and soils parameters were kept consistent with the original calibrated study.

2.2 Statistical Analysis

2.2.1 Metrics for APSIM Model Calibration

In the calibration process, to test the model's accuracy in capturing grain yield, drainage, flow-weighted $\text{NO}_3\text{-N}$, and $\text{NO}_3\text{-N}$ load (kg/ha), we compared the simulated data from the calibrated models with the corresponding observed data. The statistical analysis of the model's performance was conducted using R version 3.6.2 [56]. Goodness of fit analysis was measured by calculating r^2 , root mean square error (RMSE) and modelling efficiency (ME) [57].

2.2.2 Analysis of APSIM Model Outputs

All processing of the APSIM simulation outputs and statistical analyses were done using the tidyverse meta package [58] in R version 3.6.2 [56]. All other packages are cited below. APSIM runs on a calendar year, but to facilitate comparisons between the rotations we wanted drainage and $\text{NO}_3\text{-N}$ leaching to be summed from crop sowing to crop sowing. In order to accommodate this timeline, drainage and $\text{NO}_3\text{-N}$ leaching were categorized into two groups: season/post-harvest (sowing to the end of year, December 31) and pre-season (January 1 to the day before sowing). The season/post-harvest data of one calendar year were then combined with the pre-season data of the next year to capture the full effect of the sown crop on drainage and leaching both during the season and following harvest.

For all systems, the leaching from maize sowing to maize sowing was compared. For the cropping system consisting of maize rotated with soybean, the data from the following soybean year were added onto that of the maize year to include in the analysis any residual N from the fertilizer applied in the maize year that leached out during the soybean year. A diagrammatic representation of this method is presented in Supplementary Figure 1.

Leaching

Data were grouped into experimental units, which were defined as a site-year-rotation combination, resulting in a total of 294 units. We fit three candidate non-linear models to N leaching as a function of N fertilizer rate using the nlraa package [59]. The three models investigated were (1) bi-linear, (2) exponential, and (3) exponential-linear [60]. We chose these three models because they offer meaningful parameters and/or are commonly found in literature. Final model selection was based on Akaike's Information Criteria (AIC) [61], examination of residual plots, and knowledge of biophysical constraints to these systems.

To assess the impact of rotation on the statistical model parameters, we fit a non-linear mixed effect model using the nlme package [62]. We used a fixed effect for rotation (continuous, rotated) and random effects for site and experimental unit nested within site. Estimates and

contrasts for the effect of rotation on the parameters were assessed using the emmeans package [63]. Contributions of random effects were assessed using interclass correlation.

Yields

Several models have been proposed to determine the AONR for maize. We chose to use a bi-linear model for its simplicity and applicability to a range of scenarios [64]. While this model may not capture the nuances of real-world relationships, in our simulations it provided a more robust estimation of the yield breakpoint, which is representative of the AONR. We fit a non-linear mixed effect model as described above.

A bi-linear model outputs lower AONR values than the more conventionally used quadratic plateau model [64]. As such, we have included quadratic plateau model-derived AONR values for all site-years in the supplementary materials for comparative purposes (Supplementary Table 6).

Relationship between leaching breakpoint and AONR

The difference between the leaching and yield breakpoints was defined as the “buffer,” with positive values indicating the leaching breakpoint occurred at a higher N fertilizer rate than the yield breakpoint. The conditional breakpoint parameter estimates for the leaching and yield experimental units were calculated on a per-experimental-unit basis. The buffer for each rotation was compared using a linear model with rotation as a fixed effect.

3. Results

3.1 Model Calibration and Simulation

The APSIM model simulated yield, drainage, flow-weighted $\text{NO}_3\text{-N}$, and $\text{NO}_3\text{-N}$ leaching load well with ME values falling primarily between 0.7 and 0.95. In two studies, however, model accuracy evaluation metrics reflected modeling limitations caused by a lack of reported data: leaching and flow-weighted $\text{NO}_3\text{-N}$ data from MI [43] were reported as 3 year averages, resulting in an ME of 0.46 and yield data from IN2 [47] were reported as 6 year averages, resulting in a high $\text{RMSE}_{\text{Yield}}$ (11.22 kg/ha) (Supplementary Figures 2-5; Supplementary Table 7). Most importantly, the model captured treatment differences such as till vs no-till, narrow vs wide subsurface drain spacing, and cropping systems (Supplementary Figures 6-8). There was a wide range of AONR values found for both continuous (66-170 kg N/ha) and rotated maize (24-119 kg N/ha); the average AONR, however, was 111 kg N/ha and 70 kg N/ha for continuous and rotated maize, respectively (Figure 2b; Supplementary Figure 9; Supplementary Table 6).

3.2 Leaching Model

Of the 294 simulated site-years, 277 had all three tested non-linear models (bi-linear, linear-exponential, and exponential) converge for the $\text{NO}_3\text{-N}$ leaching load’s response to fertilizer N rate. For 92% of those site-years, the bi-linear model fit was best and for 8%, the linear-exponential fit was best. The exponential fit was not found to be the best fit for any of the site-years. The rest of our analysis, therefore, will be based on the bi-linear model fit (Figure 2c).

When no fertilizer was applied, significantly more $\text{NO}_3\text{-N}$ was leached from the maize-soybean rotation than from continuous maize (13 kg N/ha and 6 kg N/ha, respectively). Both systems’

leaching loads varied significantly by site (28% of variation) and, to a lesser extent, by year (16%) (Figure 2c).

The parameters defining the slope below and above the breakpoint as well as the breakpoint itself differed significantly with cropping system. Location and year, however, did not have a significant influence on these parameters. Below the breakpoint, continuous maize lost 0.08 kg NO₃-N per kg N applied while the maize-soybean rotation lost 0.1 kg NO₃-N. There was then a breakpoint at the fertilizer N rate 129 kg N/ha (SE: 0.6) for continuous maize and at 116 kg N/ha (SE: 1.9) for the maize-soybean rotation at which the rate of leaching changed. Above the breakpoint, continuous maize lost an average of 0.6 kg NO₃-N per kg N applied (95% CI: 0.54-0.63), but the maize-soybean rotation lost only 0.4 kg NO₃-N per kg N applied (95% CI: 0.37-0.43).

3.3 Relationship between AONR and Leaching Breakpoint

There was a greater margin for error in rotated maize than in continuous maize around overestimating a given field's AONR without drastically increasing the rate of NO₃-N leaching. The leaching breakpoint in the model occurred at an N rate 46 kg N/ha (95% CI: 44-49 kg N/ha) above the AONR in the maize-soybean rotation, but only 17 kg N/ha (95% CI: 14-20 kg N/ha) above the AONR in continuous maize (Figure 2a). The N rate that minimized the yield-scaled leaching load (i.e. the NO₃-N leaching load per unit maize grain yield) was lower than the leaching breakpoint in both cropping systems (Supplementary Figure 10).

Within the maize-soybean rotation, at N rates above the breakpoint, leaching loads during soybean following maize were higher than those in maize following maize, pointing to the residual effects of applying excessive amounts of N during maize (Figure 3). However, the leaching per year from the maize-soybean rotation system (be it the maize or soybean year) was always lower than that of continuous maize.

4. Discussion

4.1 N fertilizer-Leaching Relationship

Research suggests the N leaching-fertilizer rate breakpoint occurs around the AONR, however, we found that the leaching breakpoint occurs at N rates 16% above the AONR in maize following maize and 66% above the AONR in the maize-soybean rotation. There is, therefore, less risk in a maize-soybean rotation that farmers will have to under-fertilize while trying to minimize environmental N losses than in a continuous maize system. A meta-analysis also found that the leaching breakpoint occurs at least 15% above the AONR [1]. Our results expand upon this previous finding as it made no distinction between continuous and rotated maize: we found that system selection impacts where that breakpoint occurs. Moreover, in our analysis, we found that the breakpoint was not a function of site or year, only cropping system, whereas AONR was strongly influenced by both site and year.

Our results suggest that reductions in N fertilizer inputs have little risk of reducing productivity per unit N loss. We found that the N rate that optimized the leaching load per unit grain produced was lower than the leaching breakpoint in both systems. Because the N leaching breakpoint occurred well above the AONR and yield-scale N leaching was minimized at a N rate below the N leaching breakpoint, there should be opportunities to reduce N leaching without losing

productivity. Basing our analysis on long-term simulated experiments rather than on data from shorter-term experiments strengthens our confidence in these conclusions.

4.2 Rotation Effect

Above the breakpoint, leaching per unit N applied in a single year of continuous maize increased at a rate 1.5 times that of a full two-year cycle of a maize-soybean rotation. There was also a larger buffer between the AONR and breakpoint in the maize-soybean rotation than in continuous maize.

In APSIM, the size of the leachable N pool in each soil layer is calculated as the balance of mineralized soil N (either sourcing from soil organic matter or N fertilizer) that is not immobilized, denitrified, or taken up by the plant. The actual fate of this N is driven by the drainage of water through the soil profile. The model showed that at low N rates, the mineralized soil N pool was smaller under continuous maize than under the rotated system. When the fertilizer input exceeded the breakpoint, the mineralized N pool under continuous maize increased to the same size as that under rotation. This increase has also been found in field experiments where increasing the applied N rate above a certain threshold N rate (between 100 and 200 kg N/ha) increased the residual soil N pool under continuous maize systems more than under rotated systems [65,66]. This system-dependent effect of fertilizer on the mineralized N pool size is further evidence of a greater leaching risk under continuous maize in response to increases in N fertilizer rates.

In the model, rotated maize took up more N than the continuous maize. This gap in uptake efficiency increased as the N rate increased. Varvel and Peterson [65, 67] found that in a maize-soybean rotation, maize took up around 50% of the N applied to the crops regardless of the N rate. Meanwhile, when grown continuously, maize took up 50% of N applied at low N rates but only 30% at higher N rates [65]. The smaller buffer we found between the AONR and the leaching breakpoint under continuous maize, therefore, stems from both N saturated soil and low N uptake efficiency.

When evaluating the fate of N in a two-crop rotation, it is important to consider how much leaching occurs throughout the two-year cycle, as much of the $\text{NO}_3\text{-N}$ that was not leached during the maize year may be lost during the soybean year [30]. Our simulations found that within the two-crop rotation, soybeans experienced more drainage and leaching than rotated maize (Figure 3). Higher drainage levels correspond to more leaching, but that is likely not the only reason for the greater leaching loads during the soybean phase. While soybean took up the same amount of N as rotated maize, much of that N sources from N fixation, leaving the residual N fertilizer from the maize phase of the rotation to leach from the soil profile (68). Carryover N from the previous maize year and high drainage levels during soybean year can increase the leaching load in soybean years almost to the level of continuous maize [4,7].

At high N rates, like those above the breakpoint, continuous maize accumulated a greater pool of mineralized soil N (leachable N) relative to the rotated cropping system. High drainage loads in the soybean years of the maize-soybean rotation resulted in the loss of more N than in rotated maize years, but still less than in the continuous maize system.

4.3 Model Robustness

We found that the cropping system dictated the shape of the leaching response and breakpoint while the site and year (which encompasses annual differences in plant growth, temperature, precipitation, and soil N pool size/composition) explained the variability in the magnitude of the leaching. Therefore, while our leaching model can be applied across multiple sites and years to determine the degree to which increasing the N fertilizer rate impacts the leaching load, use of an average site-year prediction should be avoided in calculations of total leaching load from multiple sites/years at a given N rate.

The large range of baseline yields predicted may be indicative of minor limitations in the accuracy of APSIM model. Puntel et al. [18] found that while, in general, APSIM simulates yields well across the range of the N rates used in this analysis (average $RRMSE_{Yield}=14.9\%$ and 11.6% in continuous and rotated maize, respectively), its accuracy is lower when 0 kg N/ha is applied to continuous maize ($RRMSE_{Yield}=30.8\%$) than at non-zero N rates ($RRMSE_{Yield}<16\%$ at N rates ranging from $67\text{--}268\text{ kg N/ha}$).

The significant effect of site and year on the baseline leaching load is further evidence of the important role soil and weather have in determining the fate of residual N over multiple seasons. Residual N can be leached from the soil profile or taken up by the plant, resulting in high leaching loads even in seasons when no N fertilizer is applied. These legacy effects may explain why previous studies have found inconsistent and/or inconclusive differences in leaching response to N fertilizer between cropping systems depending on the site, system, and year [13,26-30].

Establishing and maintaining long-term field drainage experiments with as many different N-rates and cropping systems as we simulated ($7\text{ N rates} \times 3\text{ phases} \times 4\text{ replication} = 84\text{ plots} \times 20\text{ years}$) is practically impossible as it is expensive and laborious. Moreover, artificially drained experiments tend to be constrained by field size more than other field experiments. For instance, the drainage experimental site with the most plots in the US Midwest has 72 plots, but the plots are small ($\sim 0.05\text{ ha}$) [11]. Using cropping system and statistical models to expand upon field experiments is the only way to delineate the complex relationships between N rate, AONR, and leaching.

4.4 Implications of Findings

The smaller buffer between the AONR and the breakpoint in continuous maize suggests that the risk of negatively impacting water quality via over-fertilization in continuous maize is much greater than in a maize-soybean rotation.

A larger buffer between the AONR and breakpoint can reduce the risk that fertilizer over-application leads to groundwater contamination. After conducting a state-wide survey, Sellars et al. [69] reported that 67% of fields in Illinois received N fertilizer rates that exceeded the AONR. Combining their findings with our own demonstrates the potential impact replacing continuous maize with a maize-soybean rotation can have on $\text{NO}_3\text{-N}$ leaching in the U.S. Midwest. If the 67% statistic observed in Illinois is assumed consistent across continuous maize cropping systems in the U.S. Midwest, changing from continuous maize to rotated maize could greatly reduce the leaching from around five million hectares of cropland across the region.

This cropping system approach to reducing N leaching is an inexpensive but effective N leaching-mitigating strategy [70]. Nevertheless, there is room for improvement as there has been very little research on what might be a more desirable rotation crop in simple U.S. Midwestern two-crop rotations from a leaching perspective.

5. Conclusion

The cropping system selection plays a significant role in defining the degree to which fertilizer N rate impacts $\text{NO}_3\text{-N}$ leaching as well as the rate above which leaching increases significantly. Our findings were robust across both environment and climate variability and can provide sound guidance for how farmers can improve downstream water quality.

Acknowledgments

This work was supported in part by the Iowa Nutrient Reduction Center, the Foundation for Food and Agricultural Research (#534264), Iowa Crop Improvement Association, Iowa State University Plant Science Institute faculty scholar program, NSF (#1830478), USDA-NIFA Hatch project (IOW10480), and the National Science Foundation under grant No. DGE-1828942. We thank the APSIM Initiative for making the software publicly available and for ensuring software quality. We also thank Drs. Fernando Miguez and Phillip Dixon for their guidance and statistical expertise and Gerasimos G. Danalatos for his assistance via his GIS expertise.

Conflict of Interest Statement

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

References

- [1] Zhou M and Butterbach-Bahl K 2014 Assessment of nitrate leaching loss on a yield-scaled basis from maize and wheat cropping systems *Plant soil* **347** 977-91 doi: 10.1007/s11104-013-1876-9
- [2] Dinnes D L, Karlen K L, Jaynes D B, Kasper T C, Hatfield J L, Colvin T S and Cambardella C A 2002 Nitrogen management strategies to reduce nitrate leaching in tile-drained midwestern soils *Agron. J.* **94** 153-71
- [3] Moriasi D N, Gowda P H, Arnold J G, Mulla D J, Ale S and Steiner J L 2013 Modeling the impact of nitrogen fertilizer application and tile drain configuration on nitrate leaching using SWAT *Agric. Water Manag.* **130** 36-43
- [4] King K W, Williams M R and Fausey N R 2016 Effect of crop type and season on nutrient leaching to tile drainage under a corn-soybean rotation *J. Soil Water Conserve.* **71(1)** 56-68, doi: 10.2489/jswc.71.1.56
- [5] Bailey A, Meyer L, Pettingell N, Macie M and Korstad J 2020 Agricultural Practices Contributing to Aquatic Dead Zones *Ecological and Practical Applications for Sustainable Agriculture* 373-393 (Singapore: Springer)

- [6] Lord E I and Mitchell R D J 1998 Effect of nitrogen inputs to cereals on nitrate leaching from sandy soils. *Soil Use Manag.* **14** 78-83
- [7] Pittelkow C M, Clover M W, Hoefl R G, Nafziger E D, Warren J J, Gonizini L C and Greer K D 2017 Tile Drainage Nitrate Losses and Maize Yield Response to Fall and Spring Nitrogen Management *J. Environ. Qual.* **46** 1057-67, doi: 10.2134/jeq2017.03.0109
- [8] Perego A, Basile A, Bonfante A, de Mascellis R, Terribile F, Brenna S and Acutis M 2012 Nitrate leaching under maize cropping systems in Po Valley (Italy) *Agric. Eco. Env.* **147** 57-65, doi: 10.1016/j.agee.2011.06.014
- [9] Zhang Y, Want H, Liu S, Lei Q, Liu J, He J, Zhai L, Ren T and Liu H 2015 Identifying critical nitrogen application rate for maize yield and nitrate leaching in a Haplic Luvisol soil using the DNDC model *Sci. Total Environ.* **514** 388-98, doi: 10.1016/j.scitotenv.215.02.022
- [10] Simmelsgaard S E and Djurhuus J 1998 An empirical model for estimating nitrate leaching as affected by crop type and the long-term N fertilizer rate *Soil Use Manag.* **14(1)** 37-43
- [11] Lawlor P A, Helmers M J, Baker J L, Melvin S W and Lemke D W 2008 Nitrogen application rate effect on nitrate-nitrogen concentration and loss in subsurface drainage for a maize-soybean rotation *Trans. ASABE* **51(1)** 83-94
- [12] Delin S and Stenberg M 2013 Effect of nitrogen fertilization on nitrate leaching in relation to grain yield response on loamy sand in Sweden *Euro. J. Agron.* **52** 291-6
- [13] Christianson L E and Harmel R D 2015 4R Water Quality impacts: an assessment and synthesis of forty years of drainage nitrogen losses *J. Environ. Qual.* **44** 1852-60, doi: 10.2134/jeq2015.03.0170
- [14] Jaynes D B 2015 Maize yield and nitrate loss in subsurface drainage affected by timing of anhydrous ammonia application *Soil Sci. Soc. Am. J.* **79** 1131-41, doi: 10.2136/sssaj2015.01.0033.
- [15] Martinez-Feria R, Nichols V, Basso B and Archontoulis, S 2019 Can multi-strategy management stabilize nitrate leaching under increasing rainfall? *Environ. Res. Lett.* **14(12)** 124079
- [16] Poffenbarger H J, Barker D W, Helmers M J, Miguez F E, Olk D C, Sawyer J E, et al. 2017 Maximum soil organic carbon storage in midwest U.S. cropping systems when crops are optimally nitrogen-fertilized *PlosOne*, doi: 10.1371/journal.pone.0172293

- [17] Morris T F, Murrell T S, Beegle D B, Camberato J J, Ferguson R B, Grove J, et al. 2017 Strength and limitations of nitrogen rate recommendations for maize and opportunities for improvement *Agron. J.* **110**(1) 1-37, doi: 10.2134/agronj2017.02.0112
- [18] Puntel L A, Sawyer J E, Barker D W, Thorburn P J, Castellano M J and Moore K J 2018 A systems modeling approach to forecast maize economic optimum nitrogen rate *Front. Plant Sci.* **9** 436
- [19] Johnson G V and Raun W R 1995 Nitrate leaching in continuous winter wheat: use of a soil-plant buffering concept to account for fertilizer nitrogen *J. Prod. Agric.* **8**(4) 443-90, doi: 10.2134/jpa1995.0486
- [20] Thorburn P J, Biggs J S, Webster A J and Biggs IM 2010 An improved way to determine nitrogen fertilizer requirements of sugarcane crops to meet global environmental challenges *Plant soil* **339** 51-67, doi: 10.1007/s11104-010-0406-2
- [21] Raymond P A, David M B and Saiers J E 2012 The impact of fertilization and hydrology on nitrate fluxes from Mississippi watersheds *Curr. Opin. Env. Sust.* **4** 212-8, doi: 10.1015/jcosust.2012.04.001
- [22] Xin J, Yang L, Chen F, Duan Y, Wei G, Zheng X and Li M 2019 The missing nitrogen pieces. a critical review of the distribution, transformation, and budget of nitrogen in the valdose zone-groundwater system *Water Res.* **165** 114977, doi: 10.1016/j.watres.2019.114977
- [23] Green T R, Kipka H, David O and McMaster G S 2017 Where is the USA maize belt, a how is it changing? *Sci. Total. Environ.*, doi: 10.1016/j.scitotenv.2017.09.325
- [24] NASS 2020 Acreage USDA
- [25] Gentry L E, Below F E, David M B, and Bergerou J A 2001 Source of the soybean N credit in maize production. *Plant soil* **236** 175-184
- [26] Owens L B, Malone R W, Shipitalo M J, Edwards W M and Bonta J V 2000 Lysimeter study of nitrate leaching from a corn-soybean rotation *J. Environ. Qual.* **29** 467-74
- [27] Hernandez-Ramirez G, Brouder S M, Ruark M D and Turco R F 2011 Nitrate, phosphate, and ammonium loads at subsurface drains: agroecosystems and nitrogen management *J. Environ. Qual.* **40** 1229-40, doi: 10.2134/jeq2010.0195
- [28] Christianson L E and Harmel R D 2015 The MANAGE drain load database: Review and compilation of more than fifty years of North American drainage nutrient studies *Agric. Water Manag.* **159** 277-89, doi: 10.1016/j.agwat.2015.06.021
- [29] Daigh A L M, Zhou X, Helmers M J, Pederson C H, Horton R, Jarchow M and Liebman M 2015 Subsurface drainage nitrate and total reactive phosphorus losses in bioenergy-

- based prairies and maize systems *J. Environ. Qual.* **44** 1638-1646, doi: 10.2134/jeq2015.02.0080
- [30] Iqbal J, Necpalova M, Archontoulis S V, Anex R P, Bourguignon M, Herzmann D, et al. 2017 Extreme weather-year sequences have nonadditive effects on environmental nitrogen losses *Glob. Change Biol.*, doi: 10.1111/gcb.13866
- [31] Laan M V D, Annandale J G, Bristow K L, Stirzaker R J, du Preez C C and Thorburn P J 2014 Modelling nitrogen leaching: are we getting the right answer for the right reason? *Agric. Water Manag.* **133** 74-80, doi: 10.1016/j.agwat.2013.10.017
- [32] Holzworth D P, Huth N I, deVoil P G, Zurcher E J, Herrmann N I, McLean G, et al. 2014 APSIM—Evolution towards a new generation of agricultural systems simulation *Environ. Modell. Soft.* **62**, doi: 10.1016/j.ensoft.2014.07.009
- [33] Roberts M J, Braun N O, Sinclair T R, Lobell D B and Schlenker W 2017 Comparing and combining process-based crop models and statistical models with some implication for climate change *Environ. Res. Lett.* **12**, doi: 10.1088/1748-9326/aa7f33
- [34] Basso B, Martinez-Feria R A, Dumont B 2019 Modeling crop rotations: capturing short- and long-term feedbacks for sustainability and soil health *Burleigh Dodds Series in Agricultural Science*
- [35] Gaydon D S, Wang E, Poulton P L, Ahmad B, Ahmed F, Akhter S, et al. 2017 Evaluation of the APSIM model in cropping systems of Asia *Field Crop. Res.* **204** 52-75
- [36] Smith C J, Macdonald B C T, Xing H, Denmead O T, Wang E, McLachlan G, Toumi S, Turner D and Chen D 2019 Measurement and APSIM modelling of soil C and N dynamics. *Soil Res.* **58(1)**, doi: 10.1071/SR19021.
- [37] Archontoulis S V, Castellano M J, Licht M A, Nichols V, Baum M, Huber I, et al. 2020 Predicting crop yields and soil-plant nitrogen dynamics in the US maize belt *Crop Sci.* **60(2)** 721-38.
- [38] Pasley H R, Huber I, Castellano M J and Archontoulis S V 2020 Modeling flood-induced stress in soybeans *Front. Plant Sci.* **11(62)**
- [39] Ebrahimi-Mollabashi E, Huth N I, Holz D P, Ordonez R A, Hatfield J L, Huber I, Castellano, M J and Archontoulis S V 2019 Enhancing APSIM to simulate excessive moisture effects on root growth *Field Crop. Res.* **236**, doi: 10.1016/j.fcr.2019.03.014
- [40] Randall G W and Iragavarapu T K 1995 Impact of long-term tillage systems for continuous maize on nitrate leaching to tile drainage *J. Environ. Qual.* **24** 360-66

- [41] Huggins D R, Randall G W and Russelle M P 2001 Subsurface drain losses of water and nitrate following conversion of perennials to row crops *Agron. J.* **93**(3) 477-86
- [42] Masarik K and Norman J 2014 Long-term drainage and nitrate leaching below well-drained continuous maize agroecosystems and a prairie *J. Environ. Protect.* **5** 240-54, doi: 10.4236/jep.2014.54028
- [43] Gold A J and Loudon T L 1989 Tillage effects on surface runoff water quality from artificially drained cropland *ASAE.* **32**(4) 1329-34
- [44] Gentry L E, David M B, Smith-Starks K M and Kovacic D A 2000 Nitrogen fertilizer and herbicide transport from tile drained fields *J. Environ. Qual.* **29** 232-40
- [45] Kladvko E J, Frankenberger J R, Jaynes D B, Meek D W, Jenkinson B J and Fausey N R 2004 Nitrate leaching to subsurface drains as affected by drain spacing and changes in crop production system *J. Environ. Qual.* **33** 1803-13
- [46] Kladvko E J, Willoughby G L and Santini J B 2005 Maize growth and yield response to subsurface drain spacing on Clermont silt loam soil *Agron. J.* **97** 1419-28, doi: 10.2134/agonj2005.0090
- [47] Hofmann B S, Brouder S M and Turco R F 2004 Tile spacing impacts on Zea mays L. yield and drainage water nitrate load *J. Ecol. Engin.* **23** 251-67, doi: 10.1016/j.ecoleng.2004.09.008
- [48] Archontoulis S V, Miguez F E and Moore K J 2014 A methodology and an optimization tool to calibrate phenology of short-day species included in the APSIM PLANT model: Application to soybean *Environ. Modell. Softw.* **62** 465-477 doi: 10.1016/j.envsoft.2014.04.009
- [49] Soil Survey Staff, Natural Resources Conservation Service, United States Department of Agriculture. Web soil survey. USDA NRCS. <http://websoilsurvey.nrcs.usda.gov/app/> (accessed 1 Mar. 2019).
- [50] Thornton P E, Thornton M M, Mayer B W, Wei Y, Devarakonda R, Vose R S and Cook R B 2018 Daymet: daily surface weather data on a 1-km grid for North America, version 3, ORNL DAAC, Oak Ridge, Tennessee, USA, doi: 10.3334/ORNLDAAC/1328
- [51] Basche A D, Archontoulis S V, Kaspar T C, Jaynes D B, Parkin T B and Miguez F E. 2016 Simulating long-term impacts of cover crops and climate change on crop production and environmental outcomes in the midwestern United States *Agric. Ecosyst. Environ.* **218** 95-106

- [52] Dietzel R, Liebman M, Ewing R, Helmers M, Horton R, Jarchow M and Archontoulis S 2016 How efficiently do maize-and soybean-based cropping systems use water? A systems modeling analysis *Glob. Change Biol.* **22(2)** 666-81
- [53] Martinez-Feria R A, Dietzel R, Liebman M, Helmers M J and Archontoulis S V 2016 Rye cover crop effects on maize: A system-level analysis *Field Crop. Res.* **196** 145-59
- [54] Mourtzinis S and Conley S P 2017 Delineating soybean maturity groups across the United States *Agron. J.* **109(4)** 1397-1403, doi: 10.2134/agronj2016.10.0581
- [55] Baum M E, Archontoulis S V and Licht M A 2018 Planting date, hybrid maturity, and weather effects on maize yield and crop stage *Agron. J.* **111** 303-13, doi: 10.2134/agronj2018.04.0297
- [56] R Core Team 2019 R: A Language and environment for statistical computing, Vienna, Austria. Available at <https://www.R-project.org/>
- [57] Archontoulis S V and Miguez F E 2015 Nonlinear regression models and applications in agricultural research *Agron. J.* **107** 2, doi: 10.2134/agronj2012.0506
- [58] Wickham H, Averick M, Bryan J, Chang W, McGowan L D A, François R et al. 2019 Welcome to the Tidyverse *J. Open Source Softw.* **4(43)** 1686
- [59] Miguez F 2020 nlraa: Nonlinear Regression for Agricultural Applications. R package version 0.65. <https://CRAN.R-project.org/package=nlraa>
- [60] Miguez F, Archontoulis S and Dokoohaki H 2018 Nonlinear regression models and applications. *Applied statistics in agricultural, biological, and environmental sciences*, 401-447.
- [61] Bozdogan H 1987 Model selection and Akaike's information criterion (AIC): The general theory and its analytical extensions *Psychometrika* **52(3)** 345-370
- [62] Pinheiro J, Bates D, DebRoy S, Sarkar D, R Core Team (2019). *_nlme: Linear and Nonlinear Mixed Effects Models_*. R package version 3.1-142, <URL: <https://CRAN.R-project.org/package=nlme>>.
- [63] Lenth R 2019 emmeans: Estimated Marginal Means, aka Least-Squares Means. R package version 1.4.2. <https://CRAN.R-project.org/package=emmeans>
- [64] Cerrato M E and Blackmer AM 1990 Comparison of models for describing; corn yield response to nitrogen fertilizer *Agron. J.* **82(1)** 138-143
- [65] Varvel G E and Peterson T A 1990 Residual soil nitrogen as affected by continuous, two-year, and four-year crop rotation systems *Agron. J.* **82** 958-62

- [66] Zhu Y and Fox R H 2003 Corn-soybean rotation effects on nitrate leaching *Agron. J.* **95** 1028-1033
- [67] Varvel G E and Peterson T A 1992 Nitrogen fertilizer recovery by soybean in monoculture and rotation systems. *Agron. J.* **84** 215-8
- [68] Salvagiotti F, Cassman K G, Specht J E, Walters D T, Weiss A and Dobermann A 2008 Nitrogen uptake, fixation and response to fertilizer N in soybeans: A review *Field Crop. Res.* **108** 1-13
- [69] Sellars S C, Schnitkey G D and Gentry L F 2020 Do Illinois farmers follow university-based nitrogen recommendations? Agricultural & Applied Economics Association Annual Meeting, Kansas City, MO.
- [70] Roley S S, Tank J L, Tyndall J and Witter J D 2016 How cost-effective are cover crops, wetlands, and two stage ditches for nitrogen removal in the Mississippi River Basin? *Water Resour. Econ.* doi: 10.1016/j.wre.2016.06.003

Figures

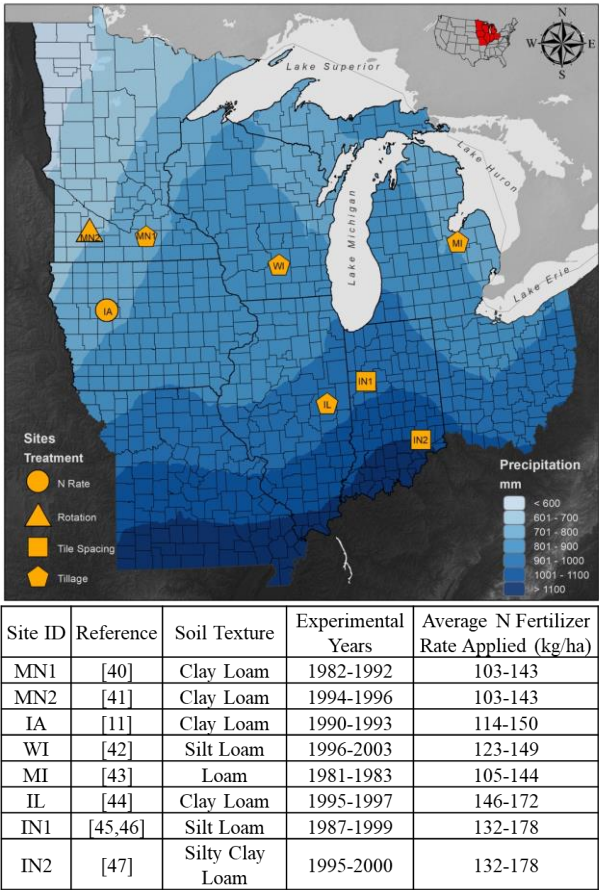


Figure 1: Location, 35-year average annual precipitation (1985-2019), general details of studies modelled in this analysis, and range of average N fertilizer rates applied to maize in each state. (Data source: USDA ERS 2019)

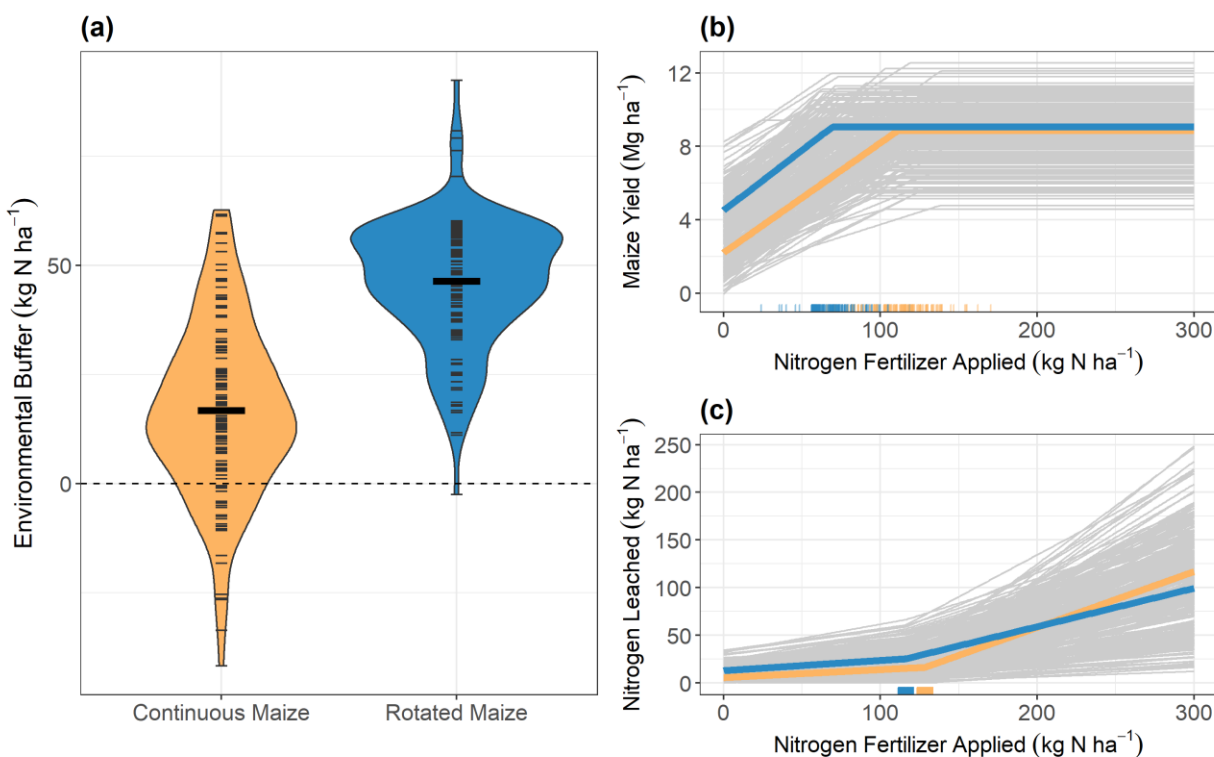


Figure 2: (a) Violin plot of the difference/buffer between the leaching breakpoint and the agronomically-optimum nitrogen rate (AONR) in continuous maize (*Zea mays*) and rotated maize (i.e. maize rotated with soybean (*Glycine max*) in a two-year cycle). The long thick horizontal line in the middle of the violin is the median. The shape and finer lines show the distribution of the data. (b) Maize yield and (c) NO₃-N leaching response to N fertilizer. Gray lines in (b) and (c) are the bilinear model predictions for each site-year, colored lines are statistical model predictions at the rotation level. Colored bars along the x-axis indicate the statistical model predicted breakpoints for continuous (orange) and rotated (blue) maize for each site.

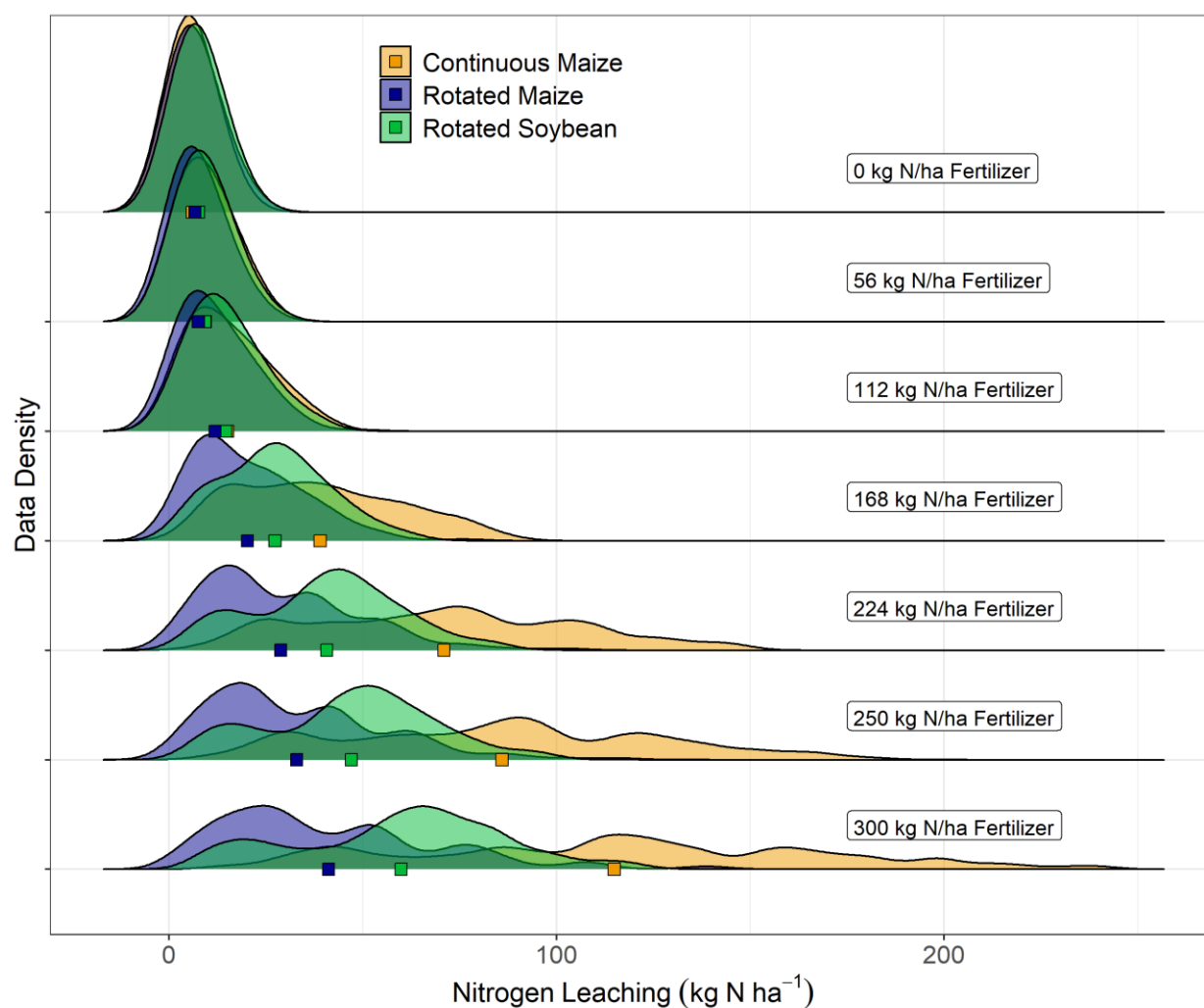


Figure 3: Distribution of $\text{NO}_3\text{-N}$ leaching loads under continuous maize (yellow), rotated maize (maize rotated with soybean in a two-year cycle) (blue), and rotated soybean (soybean rotated with maize in a two-year cycle) (green) at different fertilizer N rates with mean values indicated by squares on the x-axes. The height of the peak at a given leaching load (x axis) corresponds to the prevalence of that load across site-years when 0, 56, 112, 168, 224, 250, or 300 kg N/ha (from top to bottom) is applied. For context, the average AONR for continuous maize was 111 kg N/ha and for rotated maize, 70 kg N/ha.