



**University of
Zurich**^{UZH}

**Zurich Open Repository and
Archive**

University of Zurich
University Library
Strickhofstrasse 39
CH-8057 Zurich
www.zora.uzh.ch

Year: 2018

Soil microbes promote complementarity effects among co-existing trees through soil nitrogen partitioning

Luo, Shan ; Schmid, Bernhard ; De Deyn, Gerlinde B ; Yu, Shixiao

Abstract: Plant resource partitioning is a mechanism promoting species coexistence and ecosystem functioning. Yet, we still have limited understanding of how soil microbes, especially plant symbiotic microbes, influence resource partitioning. We hypothesized that soil-borne microbes, in particular mycorrhizal fungi, facilitate differential performance of tree species depending on different nitrogen sources and that this leads to a positive plant diversity–community productivity relationship. We conducted two complementing glasshouse experiments. In a “monoculture experiment,” we supplied nitrogen as ammonium, nitrate or glycine and tested the growth response of three tree species associated with different root symbionts: one associated with ectomycorrhizal fungi, one associated with arbuscular mycorrhizal fungi, and the third associated with both arbuscular mycorrhizal fungi and N-fixing bacteria. In an “intermixed experiment,” we grew the tree species at three richness levels (one, two or three species) in soil supplied with a mix of the three nitrogen forms or no added nitrogen, and with or without soil microbes. The monoculture experiment showed that in the presence of soil microbes, the ectomycorrhizal plant species grew best when supplied with glycine and the two arbuscular mycorrhizal plant species grew best with either nitrate or ammonium addition. When the different forms of nitrogen were mixed in the intermixed experiment, plant mixtures produced more biomass than plant monocultures in the presence of soil microbes, with positive complementarity effects indicating microbe-mediated plant resource partitioning. Our results suggest that co-existing tree species can partition soil nitrogen when grown with their particular mycorrhizal symbionts or other soil microbes, resulting in positive biodiversity effects in complex resource environments.

DOI: <https://doi.org/10.1111/1365-2435.13109>

Posted at the Zurich Open Repository and Archive, University of Zurich

ZORA URL: <https://doi.org/10.5167/uzh-168396>

Journal Article

Accepted Version

Originally published at:

Luo, Shan; Schmid, Bernhard; De Deyn, Gerlinde B; Yu, Shixiao (2018). Soil microbes promote complementarity effects among co-existing trees through soil nitrogen partitioning. *Functional Ecology*, 32(7):1879-1889.

DOI: <https://doi.org/10.1111/1365-2435.13109>

1 **Soil microbes promote complementarity effects among co-existing trees through**
2 **soil nitrogen partitioning**

3

4 Shan Luo¹, Bernhard Schmid², Gerlinde B. De Deyn³, Shixiao Yu¹*

5

6 ¹Department of Ecology, School of Life Sciences/State Key Laboratory of Biocontrol,
7 Sun Yat-sen University, Guangzhou 510275, China

8 ²Department of Evolutionary Biology and Environmental Studies, University of Zürich,
9 Winterthurerstrasse 190, CH-8057 Zürich, Switzerland

10 ³Department of Environmental Sciences, Wageningen University, P.O. Box 47, 6700
11 AA Wageningen, The Netherlands

12

13 * Corresponding author: Prof. Dr. Shixiao YU

14 Email: lssysx@mail.sysu.edu.cn

15 Address: Department of Ecology, School of Life Sciences/State Key Laboratory of
16 Biocontrol, Sun Yat-sen University, Guangzhou 510275, China

17

18 Running title: Microbe-mediated plant N partitioning

19 **Abstract**

20 **1.** Plant resource partitioning is a mechanism promoting species coexistence and
21 ecosystem functioning. Yet, we still have limited understanding of how soil microbes,
22 especially plant symbiotic microbes, influence resource partitioning. We hypothesized
23 that soil borne microbes, in particular mycorrhizal fungi, facilitate differential
24 performance of tree species depending on different nitrogen sources and that this leads
25 to a positive plant diversity–community productivity relationship.

26 **2.** We conducted two complementing glasshouse experiments. In a “monoculture
27 experiment”, we supplied nitrogen as ammonium, nitrate or glycine and tested the
28 growth response of three tree species associated with different root symbionts: one
29 associated with ectomycorrhizal fungi, one associated with arbuscular mycorrhizal
30 fungi and the third associated with both arbuscular mycorrhizal fungi and N-fixing
31 bacteria. In an “intermixed experiment”, we grew the tree species at three richness
32 levels (one, two or three species) in soil supplied with a mix of the three nitrogen forms
33 or no added nitrogen, and with or without soil microbes.

34 **3.** The monoculture experiment showed that in the presence of soil microbes, the
35 ectomycorrhizal plant species grew best when supplied with glycine and the two
36 arbuscular mycorrhizal plant species grew best with either nitrate or ammonium
37 addition. When the different forms of nitrogen were mixed in the intermixed experiment,
38 plant mixtures produced more biomass than plant monocultures in the presence of soil
39 microbes, with positive complementarity effects indicating microbe-mediated plant
40 resource partitioning.

41 **4.** Our results suggest that co-existing tree species can partition soil nitrogen when
42 grown with their particular mycorrhizal symbionts or other soil microbes, resulting in
43 positive biodiversity effects in complex resource environments.

44 **Key-words:** Biodiversity, complementarity, ecosystem functioning, mycorrhizal fungi,
45 nitrogen partitioning, plant–soil interactions, soil microbes

46 **Introduction**

47 Plant niche partitioning is considered to promote species coexistence and
48 complementary resource use, providing a mechanistic explanation for positive
49 biodiversity effects on ecosystem primary productivity in biodiversity–ecosystem
50 functioning experiments (Tilman 1982; Tilman, Lehman, & Thomson 1997; Loreau &
51 Hector, 2001; Levine & HilleRisLambers, 2009). Functionally dissimilar species may
52 occupy complementary ecological niches, for example by specializing on different
53 chemical forms of the same elemental nutrient (McKane et al., 2002; Ceulemans et al.,
54 2017). Such specialized plant nutrient uptake may be enabled by different associations
55 with symbiotic microbes, in particular mycorrhizal fungi (Phillips, Brzostek, &
56 Midgley 2013). Symbiotic microbes can expand the niche of plant species through
57 increasing plant nutrient uptake rates or altering the plant preference for specific
58 nutrient forms (van der Heijden et al., 1998; Márquez, Redman, Rodriguez, &
59 Roossinck 2007; Wu et al., 2013; Afkhami, McIntyre, & Strauss 2014). For instance,
60 mycorrhiza can promote the growth of host plants in soils with low levels of plant-
61 available phosphorus (Schweiger & Jakobsen 1999). In addition, experimental
62 grassland studies found that diverse arbuscular mycorrhizal fungal communities
63 increased plant community productivity (van der Heijden et al., 1998; Maherali &
64 Klironomos, 2007; Wagg, Barendregt, Jansa, & van der Heijden 2015). These results
65 further suggest that mycorrhiza can have positive effects on plant community as a whole
66 by expanding community niche, as predicted by a theoretical study of Poisot, Mouquet
67 and Gravel (2013). However, we still have a limited understanding as to whether the

68 combination of different types of mycorrhizal fungi and other soil microbes mediates
69 resource partitioning among plant species and whether this can lead to greater plant
70 complementarity in diverse communities.

71 Arbuscular mycorrhizal (AM) and ectomycorrhizal (EM) fungi are the two most
72 common types of mycorrhiza; they form symbioses with about 80% of all terrestrial
73 plant species and can help plant nutrient acquisition, especially nitrogen (N) and
74 phosphorus (P) (Lambers, Raven, Shaver, & Smith 2008; Smith & Read, 2008;
75 Brundrett, 2009; van der Heijden, Martin, Selosse, & Sanders 2015). However, these
76 two types of mycorrhiza are functionally distinct with respect to N-acquisition (Taylor
77 & Alexander, 2005; Smith & Read, 2008). EM fungi are capable of producing
78 extracellular enzymes that degrade complex organic N and take up dissolved organic N
79 such as amino acids (Chalot & Brun, 1998; Endo, Norisada, Kogawara, Hogetsu, &
80 Kojima 2009; Courty et al., 2010; Li et al., 2016). Unlike EM fungi that have access to
81 organic N, AM fungi have no known saprotrophic capacity and function as an extension
82 of roots to acquire inorganic N (Smith & Read, 2008; Smith & Smith, 2011; but see
83 Hodge, Campbell, & Fitter 2001; Leigh, Hodge, & Fitter 2009). It has been shown that
84 AM fungi can mobilize soil nitrate and transfer it to plant root cells (Azcón, Gomez, &
85 Tobar 1996; Azcón & Tobar, 1998; Liu et al., 2017). Thus, it is possible that tree species
86 associated with either AM and EM fungi form complementary N-uptake niches that
87 may promote their combined resource uptake, growth and productivity in forests.

88 Here we test the hypothesis that plant nutrient partitioning arises from differential
89 associations of plant species with microbial symbionts that are able to access different

90 forms of soil nutrients (Bever et al., 2010; Reynolds, Packer, Bever, & Clay 2003). The
91 few attempts to test this hypothesis have focused on grassland ecosystems dominated
92 by a single mycorrhizal type (i.e. AM fungi) and have failed to support it (Reynolds,
93 Hartley, Vogelsang, Bever, & Schultz 2005; Vogelsang, Reynolds, & Bever 2006).
94 Additional tests with plants in other systems such as forests, where some trees are
95 associated with EM fungi and others with AM fungi, and provided with mixed forms
96 of soil nutrients, are needed to evaluate whether and how mycorrhiza and other root-
97 associated soil microbes contribute to plant resource partitioning.

98 To test our hypothesis, we set up two complementing glasshouse experiments. In
99 a “monoculture experiment”, we supplied N as ammonium, nitrate or glycine and
100 measured the corresponding growth response of three tree species associated with
101 different symbiotic partners (AM fungi, EM fungi or dual AM fungi and N-fixing
102 bacteria). We predicted that species associated with EM fungi grow best on glycine,
103 species with AM fungi grow best on nitrate, and species associated with both AM fungi
104 and N-fixing bacteria grow best on ammonium, as nitrate can inhibit symbiotic nitrogen
105 fixation (Voisin, Salon, Munier-Jolain, & Ney 2002). In an “intermixed experiment”,
106 we grew the tree species at three richness levels (one, two or three species) in soil
107 supplied with a mixture of the three N forms or no added N, and with or without soil
108 microbes. Our goal was to assess whether a positive plant species richness–community
109 productivity relationship can be promoted by microbial inoculation of an environment
110 with multiple forms of soil N. Finally, we tested whether plant community productivity
111 can be predicted by their “community niche”, which represents the potential for

112 complementary resource use in a community (Salles, Poly, Schmid, & Roux 2009).

113 Our experiments address two topics. First, our test of microbe-mediated plant N
114 partitioning contributes to our understanding of the mechanisms underlying plant
115 resource partitioning. Although evidence for plant partitioning of chemical N forms is
116 accumulating in arctic tundra, alpine meadows and temperate grassland, these studies
117 did not test whether N was directly taken up by plant roots or via symbiotic microbes
118 (McKane et al., 2002; Miller & Bowman, 2002; Weigelt, Bol, & Bardgett 2005; Miller,
119 Bowman, & Suding 2007). Second, our experiments step further to look at the
120 ecosystem consequence of microbe-mediated plant N partitioning. Previous studies of
121 plant N partitioning used short-term stable isotope methods, which raises the question
122 whether the observed differences in N uptake are representative of long-term use of N
123 that facilitates plant growth and community productivity.

124

125 **Materials and Methods**

126 *Study site and species*

127 We conducted the glasshouse experiments at Heishiding Nature Reserve
128 (111°53'E, 23°27'N; 150–927 m altitude) in Guangdong Province, China. The reserve
129 has a subtropical moist monsoon climate. Mean annual temperature is 19.6°C, with a
130 high mean temperature of 28.4°C in July and a mean low temperature of 10.6°C in
131 January.

132 We chose three commonly co-occurring tree species with different mycorrhizal
133 associations: one EM species, *Lithocarpus litseifolius* (Hance) Chun (Fagaceae), and

134 two AM species, *Cryptocarya concinna* Hance (Lauraceae) and *Ormosia glaberrima* Y.
135 C. Wu (Fabaceae). *Ormosia glaberrima* is also a N-fixing species. We collected seeds
136 of these focal species during the autumn and winter of 2013. We surface-sterilized the
137 seeds (1 min 70% ethanol, 3 min 2.625% NaOCl, 1 min 70% ethanol, and 1 min distilled
138 water), and stored them at 4°C until March 2014. From March to April, all seeds were
139 germinated in plastic boxes filled with sterilized sand.

140

141 ***Background soil characteristics***

142 In June 2014, we collected loamy clay soil from the forest field. The soil had high
143 capacities for holding water and nutrients. The soil was mixed with sand (1:2 v:v),
144 sterilized by gamma radiation (25 KGy) and used as background soil. We used a
145 high proportion of sand to reduce the nutrient availability of the background soil, which
146 promoted plants to use the supplied N sources. In the background soil (pH = 4.47),
147 the ammonium concentration was 4.99 mg kg⁻¹, the nitrate concentration was 29.7
148 mg kg⁻¹, the concentration of available P was 2.17 mg kg⁻¹, and the percentage
149 organic matter was 0.55%.

150

151 ***Monoculture experiment testing for the growth response of tree species to different*** 152 ***forms of N***

153 The monoculture experiment had a factorial design that each of the three species
154 grew with unsterilized or sterilized soil inoculum and was supplied with each one of
155 the different forms of N (ammonium, nitrate or glycine addition). Each treatment was

156 replicated 15 times yielding a total of 270 pots (3 species \times 2 soil-inoculum treatments
157 \times 3 N treatments \times 15 replicates). We filled pots of 6 cm diameter and 11 cm height
158 with 200 ml of the background soil, with each pot receiving either unsterilized or
159 sterilized soil inoculum. One newly germinated seedling was transplanted into each pot.
160 The transplanted seedlings were similar in size to help reduce any seedling size-
161 dependent effects that could be introduced prior to the beginning of the experiment.

162 To prepare unsterilized or sterilized soil inoculum, we collected living soil
163 beneath adults of all focal species at the same forest field site, with three replicated
164 adults for each species. Living soil samples of the same species were then mixed
165 thoroughly and divided into two halves. The first half was sterilized by gamma
166 radiation and the second half was used as unsterilized soil inoculum. For the
167 unsterilized soil-inoculum treatment, we added 10 ml of living soil onto the
168 background soil in each pot, covered by another 10 ml of sterilized soil to prevent
169 cross-infection among pots. For the sterilized soil-inoculum treatment, we added 20 ml
170 of sterilized soil into each pot as a layer on top of the background soil. Together with
171 the original 200 ml of background soil, each pot contained 220 ml (277.4 ± 2.7 g dry
172 weight) of soil.

173 One month after the planting of seedlings, each species was treated with
174 ammonium (NH_4^+ , as $(\text{NH}_4)_2\text{SO}_4$), nitrate (NO_3^- , as KNO_3), or amino acid glycine
175 ($\text{NH}_2\text{CH}_2\text{COOH}$). Nitrogen treatments were applied in solution forms (20 ml per pot)
176 every month during the experiment at 100 mg N l^{-1} , a relatively low concentration (7.21
177 mg kg^{-1}) designed to ensure full uptake of the supplied N by plants, which according to

178 our observation was not stressful for plant growth. We intermittently stopped watering
179 the plants four days before each N treatment without causing any drought symptoms to
180 reduce soil moisture and prevent the outflow of added N. In total, 28 mg N was added
181 per pot during the experimental period. We placed all pots in 15 blocks, within which
182 pots were arranged in random positions.

183

184 ***Intermixed experiment testing for the response of communities to mixed N forms***

185 Using the same set of species as in the monoculture experiment, we assembled
186 plant communities into pots at three species richness levels (one, two or three species)
187 in the intermixed experiment. Specifically, each of the three species was planted as a
188 monoculture (richness level of one). All three possible two-species mixtures and the
189 three-species mixture were planted with the species in equal portions and the same total
190 density as the monocultures, namely six individuals per pot (substitutive design). Each
191 community composition grew in three different soil conditions: 1) unsterilized or 2)
192 sterilized soil supplied with a mix of the three N forms used separately in the
193 monoculture experiment or 3) unsterilized soil without N addition. Each experimental
194 unit had 10 replicates, in total there were 210 pots (7 assemblages \times 3 soil conditions \times
195 10 replicates). We filled pots of 12 cm diameter and 13 cm height with 1000 ml of the
196 background soil as used in the monoculture experiment, with each pot receiving either
197 unsterilized or sterilized soil inoculum. Six newly germinated seedlings were
198 transplanted into each pot. In total, there were 1260 tree seedlings (210 pots \times 6
199 seedlings).

200 To prepare the unsterilized and sterilized soil inoculum, we followed the same
201 procedure as in the monoculture experiment, with the exception that we mixed all living
202 soil samples. For the unsterilized soil-inoculum treatment, we added 50 ml of living
203 soil onto the background soil in each pot, covered by another 50 ml of sterilized soil.
204 For the sterilized soil-inoculum treatment, we added 100 ml of sterilized soil into each
205 pot as a layer on top of the background soil. Together with the original 1000 ml of
206 background soil, each pot contained 1100 ml (1363.2 ± 10.8 g dry weight) of soil. For
207 the N-addition treatment, a mixture of all three N forms was applied monthly as a
208 solution (50 ml per pot) at 200 mg N l^{-1} . The concentration of added N (7.34 mg kg^{-1})
209 in the intermixed experiment was similar to that of the monoculture experiment (7.21
210 mg kg^{-1}). In total, 140 mg N was added per pot for the N-addition treatment during the
211 experiment. The control pots received the same amount of water as was supplied to the
212 pots with N addition. There were 10 blocks, within which the pots with the different
213 treatment combinations were arranged in random positions.

214 For both experiments, plants were watered twice a week between nutrient
215 applications and allowed to grow for 14 months. Then we harvested the plants and
216 determined the dry weight of shoots and roots for every individual.

217

218 *Community niche*

219 The niche of a given community (community niche, CN) was calculated based on
220 the performance of each species on each N form in the presence of soil microbes,
221 according to the formula:

$$CN = \sum_{i=1}^3 \max_{j=1}^n (P_{ij}) \text{ (eqn 1)}$$

where P_{ij} is the best performance (here we used biomass) of species j on N form i , and n is the number of species in the considered community (Salles et al., 2009). The calculated community niche corresponds to the sum of the best performances of each species across the different N forms.

Mycorrhizal root colonization

To determine the effectiveness of mycorrhizal fungi inoculation using live soil inoculum and soil sterilization, we measured mycorrhizal colonization of roots sampled from both unsterilized and sterilized soil-inoculum treatments. As the percentage of soil inoculum that distinguished the unsterilized and sterilized soil treatments in each pot remained the same for the monoculture and intermixed experiments, we only sampled roots in the monoculture experiment. We acknowledge that this sampling method may neglect the potential effects of plant competition on mycorrhizal root colonization in the intermixed experiment, but at least it reflects the direct effect of soil-inoculum treatments on mycorrhizal root colonization in both experiments.

For each species, we collected five root samples for each soil-inoculum treatment. To evaluate the level of AM colonization, representative root samples were stained using the ink and vinegar staining technique (Vierheilig, Coughlan, Wyss, & Piché 1998). We assessed colonization of 10 randomly chosen 2-cm root sections per sample using the gridline intersection method (Giovannetti & Mosse, 1980). For each of the 2-

243 cm root sections, we checked 20 intersections. In total, we checked 200 intersections
244 for each sample (20×10). We scored vesicles, arbuscules and AM hyphae as positive
245 AM encounters. For each sample, the percentage of infection was calculated as the ratio
246 of intersections that contained AM fungal structures to the total number of intersections
247 observed.

248 The level of EM colonization was determined on representative root samples and
249 based on counts of all root tips present in the counting dish. We also had 10 randomly
250 chosen 2-cm root sections per sample. The presence of Hartig net and fungal mantle
251 distinguished EM from non-EM root tips (Gehring & Whitham, 1991). Therefore, all
252 root tips (non-EM and EM tips) were counted until 100 EM or 100 non-EM was reached
253 (Teste, Karst, Jones, Simard, & Durall 2006). Percentage of EM colonization was
254 calculated as the number of active EM root tips divided by the total number of root tips.

255

256 *Statistical analyses*

257 *Plant species-specific growth response to different N forms in the monoculture* 258 *experiment*

259 We tested whether different tree species grew differently on the three N forms and
260 whether this depended on soil inoculum. We used a general linear model with the total
261 of above- and belowground biomass (log-transformed) of plant individuals as
262 dependent variable, and plant species identity, N treatments (ammonium, nitrate or
263 glycine addition), soil-inoculum treatments (unsterilized or sterilized inoculum) and
264 their interactions as explanatory variables.

265 Further, we tested whether each tree species grew best on a different form of N,
266 as predicted in our hypothesis: *L. litseifolius* with EM fungi was expected to grow best
267 on organic N, *C. concinna* with AM fungi on nitrate, and *O. glaberrima* with AM fungi
268 and N-fixing bacteria on ammonium. We modified the above model by using the
269 contrast “home” within the species identity \times N treatments interaction, calling those
270 species-by-N treatment combinations “home” that corresponded to the predicted N
271 forms with which each tree species was expected to grow best (see Table 1 and Joshi et
272 al., 2001 for how to apply “home” contrasts to interactions). The corresponding three-
273 way interaction with soil-inoculum treatments was then used to test if the “home” effect
274 was stronger in unsterilized soil, which would support the hypothesis that soil microbes
275 are responsible for the different responses of tree species to different N forms.

276

277 *Effects of soil inoculum on plant species richness–community productivity relationships*
278 *and plant biodiversity effects*

279 We split the data of the intermixed experiment into two parts, namely with or
280 without N addition, and analyzed them separately. For the first part with N addition, we
281 tested how community composition (the three monocultures, three 2-species mixtures
282 and one 3-species mixture) — partitioned into a linear contrast of plant species richness
283 and remaining community composition (for detailed explanation of this approach see
284 Schmid, Baruffol, Wang, & Niklaus 2017) — and soil-inoculum treatments
285 (unsterilized or sterilized inoculum) affected community productivity (the total of plant
286 above- and belowground biomass) with an overall general linear model. Furthermore,

287 we analyzed the community productivity with N addition but with unsterilized or
288 sterilized soil inoculum separately using general linear models. For the second part
289 without N addition, similarly, we analyzed the responses of community productivity to
290 community composition using a general linear model. In all analyses we included block
291 as a fixed-effects term at the beginning of the model (Schmid et al., 2017). Because we
292 treated community composition as a fixed-effects term our inferences about
293 biodiversity effects are restricted to the particular species composition investigated.

294 Moreover, we tested whether soil-inoculum treatments (unsterilized or sterilized
295 inoculum) affected plant biodiversity effects when supplied with N. We used the
296 additive partitioning method of Loreau and Hector (2001) to partition net biodiversity
297 effects (NEs) into complementarity (CEs) and selection effects (SEs). We used similar
298 general linear models and fitted the following terms in sequential order: block, soil-
299 inoculum treatments and community composition. We did not include plant species
300 richness in this analysis, because it would only have contrasted the three-species
301 mixture with the mean of the three two-species mixtures. However, we also tested
302 whether the overall mean of the NE, CE and SE were different from zero, listing the
303 significance of the intercept in the corresponding analysis of variance (ANOVA) table
304 (Table 2).

305

306 *Effects of community niche on community productivity*

307 In unsterilized soil supplied with a mixture of N forms, where plant species
308 richness was expected to significantly affect community productivity, we tested

309 whether community niche (CN) predicted community productivity. To do this, we
310 replaced plant species richness by CN in the general linear model.

311 All statistical analyses were done in R 3.4.2 (R Development Core Team, 2017).

312 **Results**

313 ***Mycorrhizal root colonization***

314 Plants grown in sterilized soil were relatively free of root colonization by
315 mycorrhiza at the end of the experiment (see Table S1), with a mean colonization of
316 8.1% compared with 66.0% in plants grown in unsterilized soil ($P < 0.001$). For plants
317 grown in unsterilized soil, the percentage of mycorrhizal colonization was $73.7\% \pm 4.8\%$
318 (mean \pm one standard error) for *C. concinna* (AM fungi), $77.3\% \pm 6.7\%$ for *O.*
319 *glaberrima* (AM fungi) and $47.0\% \pm 6.0\%$ for *L. litseifolius* (EM fungi).

320

321 ***Biomass of tree species in response to different N forms***

322 The three tree species differed in biomass on different N forms when inoculated
323 with unsterilized soil in the predicted way (significant $ID \times N \times S$ interaction; Table 1;
324 Fig. 1). Specifically, each species achieved the highest biomass on a different form of
325 N (home contrast of interaction $ID \times N$ in Table 1) and this effect was particularly strong
326 when microbes were present in the soil (home \times S contrast of interaction $ID \times N \times S$ in
327 Table 1; Fig. 1a). The residual interaction terms after removing the home contrasts were
328 very small and insignificant ($ID \times N$ and $ID \times N \times S$ terms following the home and
329 home \times S contrasts, respectively, in Table 1), suggesting that there were no further
330 differences among the three species with respect to the different N forms on which each

331 species grew best. In contrast, when inoculated with sterilized soil, all tree species grew
332 equally well on the different forms of N (Fig. 1b).

333

334 *Effects of soil inoculum on plant biodiversity effects and plant species richness–*
335 *community productivity relationships*

336 In soils supplied with a mix of N forms, the overall means of the NE and CE were
337 significantly positive and the overall mean of the SE was significantly negative
338 (intercept in Table 2). Soil-inoculum treatments had no significant effects on the NE
339 (Fig. 2 a-b), but significantly affected the CE and SE in an opposite way (soil-inoculum
340 treatments in Table 2). When supplied with a mix of N forms, the CE was significantly
341 larger with than without soil microbes (Fig. 2 c-d), whereas the presence of soil
342 microbes caused a more negative SE (Fig. 2 e-f).

343 In soils supplied with a mix of N forms, plant community productivity linearly
344 increased with increasing plant species richness (SR in Table 3, model (A)). Moreover,
345 the presence of soil microbes significantly increased community productivity across
346 plant species richness as compared to that without soil microbes (S in Table 3, model
347 (A)), but it did not change the overall pattern of the plant species richness–community
348 productivity relationship ($S \times SR$ interaction in Table 3, model (A)). Nevertheless, when
349 analyzed separately, there were non-significant and significantly positive relationships
350 between species richness and productivity in soils without or with microbes,
351 respectively (SR in Table 3, model (B) & model (C)). For communities without N
352 addition, their productivity was not significantly related to plant species richness (SR

353 in Table S2).

354

355 *Community niche predicts community productivity*

356 Our intermixed experiment was designed in such a way that species richness
357 should correlate as highly as possible with community niche, which was indeed the case
358 ($r = 0.93$, $P < 0.001$, see Fig. S1). In addition to that, plant species richness significantly
359 affected community productivity in unsterilized soil supplied with a mix of N forms
360 (SR in Table 3, model (C)), the surrogate parameter CN (community niche) had a
361 similarly positive effect on community productivity (Table 4).

362

363 **Discussion**

364 In this study, we selected three tree species associated with different microbial
365 symbionts (one species with EM fungi, one species with AM fungi and the third species
366 with AM fungi and N-fixing bacteria) to examine microbe-mediated N partitioning and
367 its effects on plant diversity–community productivity relationships. We found that in
368 the presence of microbes each tree species grew best on a different form of N. This
369 suggests that these tree species have the potential for N partitioning. Our study provides
370 the first empirical evidence for microbe-mediated resource partitioning among woody
371 species (Reynolds et al., 2003), but given the small number of species tested, we cannot
372 say how general the phenomenon could be. Previous studies focusing on grassland plant
373 species did not find evidence for it (Reynolds et al., 2005; Vogelsang et al. 2006).
374 Consistent with the observed microbe-mediated plant resource partitioning, our mixed-

species plant communities had larger biodiversity complementarity effects on soils supplied with a mix of N forms in the presence of soil microbes as compared to when soil microbes were absent. Under this condition, community niche, representing the potential of the community to extract N from the soil, was also positively related to community productivity. Our results highlight the potential of microbe-mediated N partitioning among plant species to influence plant diversity–community productivity relationships.

Exploitation of different N forms by different tree species

In sterilized soil, the three tree species grew equally well on the three chemical forms of N, i.e. nitrate, ammonium and organic glycine. In the presence of soil microbes, however, the EM host species *L. litseifolius* grew best with glycine, the AM host species *C. concinna* grew best with nitrate and the AM host and N-fixing species *O. glaberrima* did not show a clear growth distinction but compared with the other two species made best use of ammonium (see Fig. 1). As these differential plant growth responses to N treatments existed in unsterilized soil but not in sterilized soil, it is reasonable to speculate that the differential performance of plant species is related to the direct effect of different N forms rather than the potential different soil pH as a side effect of N addition. We note that our experiment cannot distinguish whether the added N forms were taken up directly by plants or as other N forms after microbial transformation (Harrison, Bol, & Bardgett 2007; Harrison, Bol, & Bardgett 2008), but it does provide evidence that addition of different N forms can cause differential performance of tree

397 species associated with different mycorrhizal types or other soil microbes. This suggest
398 that the three species had the potential to partition N use if the element was available in
399 the different chemical forms, and that this ability depended on the presence of soil
400 microbes, especially mycorrhizal symbionts, which achieved high root colonization in
401 the experiment. Although our plants interacted with entire soil microbial communities,
402 and although the degree of mycorrhizal colonization is not necessarily strongly
403 correlated with mycorrhizal function (Aerts, 2003), it seems most parsimonious to
404 assign the observed effects mainly to mycorrhiza, with other soil biota contributing to
405 a smaller extent.

406 These results support our hypothesis of microbe-mediated complementary N-
407 uptake niches among plant species, adding to the evidence that distinct root mycorrhizal
408 symbionts are important for plant nutrient acquisition (Phillips, Brzostek, & Midgley
409 2013; Steidinger, Turner, Corrales, & Dalling 2015). Some studies with tropical tree
410 species found that partitioning of soil P can occur between or even within plant
411 functional groups (i.e. AM fungi) that host different mycorrhizal types (Steidinger et
412 al., 2015; Nasto et al., 2017). In contrast, another study on more fertile soil only
413 provided weak evidence for N partitioning among tropical tree species with different
414 mycorrhizal types (Andersen, Mayor, & Turner 2017). It is possible that the N
415 partitioning we observed is more common on soils with low overall N availability than
416 on N-rich soils.

417

418 *Microbe mediated N partitioning among tree species underpinning the positive*

419 *complementarity effect*

420 In the intermixed experiment, we found that soil microbes had a significantly
421 positive effect on the CE when supplied with a mix of N forms, whereas these
422 conditions did not lead to an increased NE because the negative SE counteracted the
423 positive CE. This may explain why soil microbes did not significantly affect the plant
424 species richness–community productivity relationship. However, our hypothesis that
425 microbe-mediated resource partitioning among plant species promote complementary
426 N use in mixed species communities is evident in terms of the complementarity effect
427 CE. It should be noted that the increased CE could not be attributed to the N-fixing
428 species (i.e. *O. glaberrima*) and the associated N-fixing bacteria, because the plant
429 species combinations with this species did not show a high CE (see Fig. 2 c).

430 Moreover, we found similarly positive effects on community productivity of tree
431 species richness and community niche in the presence of soil microbes with added N.
432 Therefore, we conclude that under these conditions the positive complementarity
433 effects are likely driven by community niche related to N extraction. This interpretation
434 is consistent with the suggestion that microbe-mediated resource partitioning between
435 tree species allows mixed-species plant communities to better exploit N when this
436 resource is available in different forms. Our study suggests that even though plant
437 species harbouring different mycorrhiza or other soil microbes have the potential to
438 partition N use, the combination of different species will have no growth benefit if N is
439 not available in multiple forms (Table S2). This corresponds to results of other studies,
440 which also found weaker biodiversity effects when available biotope space was

441 experimentally reduced by changing soil depth or complexity of resource environments
442 (Dimitrakopoulos & Schmid, 2004; Jousset, Schmid, Scheu, & Eisenhauer 2011).

443 One previous study found that plants associated with one mycorrhizal type can
444 promote the growth of their neighbors with different root symbionts (Teste, Veneklaas,
445 Dixon, & Lambers 2014). We took that idea a step further and show that tree species
446 with different mycorrhizal types potentially partition N via soil microbes, which can
447 promote complementarity effects in communities. Our results do not exclude the
448 possibility that different plants in a community may further trade nutrients through
449 common hyphal networks and this may further affect community productivity (Simard
450 & Durall, 2004; Walder et al., 2012; Klein, Siegwolf, & Körner 2016). However, such
451 exchange requires common hyphal networks (Simard & Durall, 2004) and thus should
452 not occur between species associating with different types of mycorrhizal fungi. Thus,
453 potential hyphal links should not alter our main conclusion that N partitioning among
454 tree species promoted complementarity effects of mixed-species communities.

455

456 ***Implications for species coexistence and ecosystem functioning***

457 If and where the microbe-mediated soil nutrient partitioning does occur in forest,
458 it may not only increase resource extraction by mixed-species plant communities but
459 also facilitate plant species coexistence in complex resource environments. Although
460 plant–mycorrhiza interactions have been repeatedly shown to affect the coexistence of
461 different plant species and biodiversity–productivity relationships (van der Heijden et
462 al., 1998; Klironomos, Mccune, Hart, & Neville 2000; Maherli & Klironomos, 2007;

463 van der Heijden, Bardgett, & van Straalen 2008; Wagg et al., 2015; Bennett et al., 2017;
464 Jiang et al., 2017; Luo, De Deyn, Jiang, & Yu 2017; Teste et al., 2017), our case study
465 reveals the role of soil microbes in plant-N partitioning. Our species pool contained
466 only three tree species, therefore we cannot extrapolate the results to a broader set of
467 mature canopy trees in natural forest. However, we at least provide a proof of principle
468 for the mediation of N partitioning by soil microbes, which furthermore may promote
469 plant species coexistence by minimizing interspecific competition.

470 Our study also provides a proof of principle that microbe-mediated plant N
471 partitioning can enhance the CE component of the net biodiversity effect (NE), even
472 though at the expense of a decreased SE component. Thus, our study establishes a
473 mechanistic link between complementary resource use and the positive plant
474 diversity–community productivity relationship (Turnbull, Levine, Loreau, & Hector
475 2013). Disadvantages of our study are the relatively short duration of 14 months and
476 the use of tree seedlings rather than larger plant individuals. The complementarity
477 among plant species is likely to increase over time, and the effects of soil microbes on
478 plants may have a time lag (Cardinale et al., 2007; Fargione et al., 2007; Eisenhauer,
479 Reich, & Scheu 2012; Reich et al., 2012). It is therefore conceivable that the observed
480 microbe-mediated N partitioning among tree species, as well as the positive
481 complementarity effect, would be even larger in longer-term experiments.

482 In conclusion, the complementing results of our two experiments provide
483 evidence that plant resource partitioning mediated by soil microbes may be an
484 underlying mechanism for efficient resource extraction by species-rich plant

485 communities in environments with complex forms of elemental nutrients. Our study
486 expands previous concepts of complementary resource use by showing the intimate
487 connections between plant–microbe associations, plant niche partitioning and
488 ecosystem functioning. Future studies applying our proofs of concepts to natural forest
489 ecosystems with higher species diversity, longer time scales and *in situ* experimentation
490 may provide further mechanistic insight into species coexistence and biodiversity–
491 ecosystem functioning relationships.

492

493 **Author's contributions**

494 SL designed the experiment with advices from SY, and conducted the experiment. SL
495 and BS analyzed the data. SL wrote the first draft of the manuscript, and all authors
496 discussed the results and contributed substantially to revisions.

497

498 **Acknowledgements**

499 We are grateful to Xubing Liu for discussions about the experimental design, and to
500 Bin Jiang, Zishan Li, Saisai Tian and Weinan Ye for their assistance in the field. The
501 helpful comments of two reviewers are greatly appreciated. This research was funded
502 by the National Natural Science Foundation of China (grant no. 31230013 to SY) and
503 the Zhang-Hongda Science Foundation in Sun Yat-sen University. BS was supported
504 by the University of Zurich Research Priority Program on Global Change and
505 Biodiversity (URPP GCB).

506

507 **Data Accessibility**

508 The data supporting the results are archived in Dryad:
509 <https://doi.org/10.5061/dryad.96r47g0> (Luo, Schmid, De Deyn, & Yu 2018)

510

511 **References**

512 Aerts, R. (2003) The role of various types of mycorrhizal fungi in nutrient cycling and
513 plant competition. In M.G.A. van der Heijden & I.R. Sanders (Eds.),
514 *Mycorrhizal Ecology* (pp. 117-133). Berlin Heidelberg: Springer.

515 Afkhami, M.E., McIntyre, P.J., & Strauss, S.Y. (2014). Mutualist-mediated effects on
516 species' range limits across large geographic scales. *Ecology Letters*, 17, 1265-
517 1273. <https://doi.org/10.1111/ele.12332>

518 Andersen, K.M., Mayor, J.R., & Turner, B.L. (2017). Plasticity in N uptake among
519 sympatric species with contrasting nutrient acquisition strategies in a tropical
520 forest. *Ecology*, 98, 1388-1398. <https://doi.org/10.1002/ecy.1793>

521 Azcón, R., Gomez, M., & Tobar, R. (1996). Physiological and nutritional responses by
522 *Lactuca sativa* L. to nitrogen sources and mycorrhizal fungi under drought
523 conditions. *Biology and Fertility of Soils*, 22, 156-161.

524 Azcón, R., & Tobar, R. M. (1998). Activity of nitrate reductase and glutamine
525 synthetase in shoot and root of mycorrhizal *Allium cepa*: effect of drought
526 stress. *Plant Science*, 133, 1-8. [https://doi.org/10.1016/S0168-9452\(96\)04533-](https://doi.org/10.1016/S0168-9452(96)04533-5)

527 5

528 Bennett, J.A., Maherali, H., Reinhart, K.O., Lekberg, Y., Hart, M.M., & Klironomos, J.

529 (2017). Plant-soil feedbacks and mycorrhizal type influence temperate forest
 530 population dynamics. *Science*, 355, 181-184.
 531 <https://doi.org/10.1126/science.aai8212>

532 Bever, J.D., Dickie, I.A., Facelli, E., Facelli, J.M., Klironomos, J., Moora, M., ... Zobel,
 533 M. (2010). Rooting theories of plant community ecology in microbial
 534 interactions. *Trends in Ecology and Evolution*, 25, 468-478.
 535 <https://doi.org/10.1016/j.tree.2010.05.004>

536 Brundrett, M.C. (2009). Mycorrhizal associations and other means of nutrition of
 537 vascular plants: understanding the global diversity of host plants by resolving
 538 conflicting information and developing reliable means of diagnosis. *Plant and*
 539 *Soil*, 320, 37-77.

540 Cardinale, B.J., Wright, J.P., Cadotte, M.W., Carroll, I.T., Hector, A., Srivastava,
 541 D.S., ... Weis, J.J. (2007). Impacts of plant diversity on biomass production
 542 increase through time because of species complementarity. *Proceedings of the*
 543 *National Academy of Sciences of the United States of America*, 104, 18123-
 544 18128. <https://doi.org/10.1073/pnas.0709069104>

545 Ceulemans, T., Bode, S., Bollyn, J., Harpole, S., Coorevits, K., Peeters, G., ... Honnay,
 546 O. (2017). Phosphorus resource partitioning shapes phosphorus acquisition and
 547 plant species abundance in grasslands. *Nature Plants*, 3, 16224.
 548 <https://doi.org/10.1038/nplants.2016.224>

549 Chalot, M., & Brun, A. (1998). Physiology of organic nitrogen acquisition by
 550 ectomycorrhizal fungi and ectomycorrhizas. *FEMS Microbiology Reviews*, 22,

551 21-44. <https://doi.org/10.1111/j.1574-6976.1998.tb00359.x>

552 Courty, P.-E., Buée, M., Diedhiou, A.G., Frey-Klett, P., Le Tacon, F., Rineau, F., ...

553 Garbaye, J. (2010). The role of ectomycorrhizal communities in forest

554 ecosystem processes: New perspectives and emerging concepts. *Soil Biology*

555 *and Biochemistry*, 42, 679-698. <https://doi.org/10.1016/j.soilbio.2009.12.006>

556 Dimitrakopoulos, P.G., & Schmid, B. (2004). Biodiversity effects increase linearly with

557 biotope space. *Ecology Letters*, 7, 574-583. [https://doi.org/10.1111/j.1461-](https://doi.org/10.1111/j.1461-0248.2004.00607.x)

558 [0248.2004.00607.x](https://doi.org/10.1111/j.1461-0248.2004.00607.x)

559 Eisenhauer, N., Reich, P.B., & Scheu, S. (2012). Increasing plant diversity effects on

560 productivity with time due to delayed soil biota effects on plants. *Basic and*

561 *Applied Ecology*, 13, 571-578. <https://doi.org/10.1016/j.baae.2012.09.002>

562 Endo, I., Norisada, M., Kogawara, S., Hogetsu, T., & Kojima, K. (2009). Nitrogen-form

563 preference of *Pinus densiflora* seedlings is affected by ectomycorrhizal

564 association. *Journal of Plant Nutrition and Soil Science*, 172, 623-625.

565 <https://doi.org/10.1002/jpln.200800088>

566 Fargione, J., Tilman, D., Dybzinski, R., Lambers, J.H.R., Clark, C., Harpole, W.S., ...

567 Loreau, M. (2007). From selection to complementarity: shifts in the causes of

568 biodiversity–productivity relationships in a long-term biodiversity experiment.

569 *Proceedings of the Royal Society of London B: Biological Sciences*, 274, 871-

570 876. <https://doi.org/10.1098/rspb.2006.0351>

571 Gehring, C.A., & Whitham, T.G. (1991). Herbivore-driven mycorrhizal mutualism in

572 insect-susceptible pinyon pine. *Nature*, 353, 556-557.

573 <https://doi.org/10.1038/353556a0>

574 Giovannetti, M., & Mosse, B. (1980). An evaluation of techniques for measuring
575 vesicular arbuscular mycorrhizal infection in roots. *New Phytologist*, 84, 489-
576 500. <https://doi.org/10.1111/j.1469-8137.1980.tb04556.x>

577 Harrison, K.A., Bol, R., & Bardgett, R.D. (2007). Preferences for different nitrogen
578 forms by coexisting plant species and soil microbes. *Ecology*, 88, 989-999.
579 <https://doi.org/10.1890/06-1018>

580 Harrison, K.A., Bol, R., & Bardgett, R.D. (2008). Do plant species with different
581 growth strategies vary in their ability to compete with soil microbes for
582 chemical forms of nitrogen? *Soil Biology and Biochemistry*, 40, 228-237.
583 <https://doi.org/10.1016/j.soilbio.2007.08.004>

584 Hodge, A., Campbell, C.D., & Fitter, A.H. (2001). An arbuscular mycorrhizal fungus
585 accelerates decomposition and acquires nitrogen directly from organic material.
586 *Nature*, 413, 297-299. <http://doi.org/10.1038/35095041>

587 Jiang, J., Moore, J.A.M., Priyadarshi, A., & Classen, A.T. (2017). Plant-mycorrhizal
588 interactions mediate plant community coexistence by altering resource demand.
589 *Ecology*, 98, 187-197. <https://doi.org/10.1002/ecy.1630>

590 Joshi, J., Schmid, B., Caldeira, M.C., Dimitrakopoulos, P.G., Good, J., Harris, R., ...
591 Lawton, J.H. (2001). Local adaptation enhances performance of common plant
592 species. *Ecology Letters*, 4, 536-544. [https://doi.org/10.1046/j.1461-](https://doi.org/10.1046/j.1461-0248.2001.00262.x)
593 [0248.2001.00262.x](https://doi.org/10.1046/j.1461-0248.2001.00262.x)

594 Jousset, A., Schmid, B., Scheu, S., & Eisenhauer, N. (2011). Genotypic richness and

595 dissimilarity opposingly affect ecosystem functioning. *Ecology Letters*, 14,
 596 537-545. <https://doi.org/10.1111/j.1461-0248.2011.01613.x>

597 Klein, T., Siegwolf, R.T.W., & Körner, C. (2016). Belowground carbon trade among
 598 tall trees in a temperate forest. *Science*, 352, 342-344.
 599 <https://doi.org/10.1126/science.aad6188>

600 Klironomos, J.N., McCune, J., Hart, M., & Neville, J. (2000). The influence of
 601 arbuscular mycorrhizae on the relationship between plant diversity and
 602 productivity. *Ecology Letters*, 3, 137-141. [https://doi.org/10.1046/j.1461-](https://doi.org/10.1046/j.1461-0248.2000.00131.x)
 603 [0248.2000.00131.x](https://doi.org/10.1046/j.1461-0248.2000.00131.x)

604 Lambers, H., Raven, J.A., Shaver, G.R., & Smith, S.E. (2008). Plant nutrient-
 605 acquisition strategies change with soil age. *Trends in Ecology and Evolution*, 23,
 606 95-103. <https://doi.org/10.1016/j.tree.2007.10.008>

607 Leigh, J., Hodge, A., & Fitter, A.H. (2009). Arbuscular mycorrhizal fungi can transfer
 608 substantial amounts of nitrogen to their host plant from organic material. *New*
 609 *Phytologist*, 181, 199-207. <http://doi.org/10.1111/j.1469-8137.2008.02630.x>

610 Levine, J.M., & HilleRisLambers, J. (2009). The importance of niches for the
 611 maintenance of species diversity. *Nature*, 461, 254-257.
 612 <http://doi.org/10.1038/nature08251>

613 Li, C., Li, Q., Qiao, N., Xu, X., Li, Q., & Wang, H. (2016). Inorganic and organic
 614 nitrogen uptake by nine dominant subtropical tree species. *iForest -*
 615 *Biogeosciences and Forestry*, 9, 253-258. <http://doi.org/10.3832/ifor1502-008>

616 Liu, M., Li, C., Xu, X., Wanek, W., Jiang, N., Wang, H., & Yang, X. (2017). Organic

617 and inorganic nitrogen uptake by 21 dominant tree species in temperate and
 618 tropical forests. *Tree Physiology*, 37, 1515-1526.
 619 <https://doi.org/10.1093/treephys/tpx046>
 620 Loreau, M., & Hector, A. (2001). Partitioning selection and complementarity in
 621 biodiversity experiments. *Nature*, 412, 72-76. <https://doi.org/10.1038/35083573>
 622 Luo, S., De Deyn, G.B., Jiang, B., & Yu, S. (2017). Soil biota suppress positive plant
 623 diversity effects on productivity at high but not low soil fertility. *Journal of*
 624 *Ecology*, 105, 1766-1774. <https://doi.org/10.1111/1365-2745.12773>
 625 Luo, S., Schmid, B., De Deyn, G.B., & Yu, S. (2018). Data from: Soil microbes promote
 626 complementarity effects among co-existing trees through soil nitrogen
 627 partitioning. Dryad Digital Repository. <https://doi.org/10.5061/dryad.96r47g0>
 628 Maherali, H., & Klironomos, J.N. (2007). Influence of phylogeny on fungal community
 629 assembly and ecosystem functioning. *Science*, 316, 1746-1748.
 630 <https://doi.org/10.1126/science.1143082>
 631 Márquez, L.M., Redman, R.S., Rodriguez, R.J., & Roossinck, M.J. (2007). A virus in a
 632 fungus in a plant: three-way symbiosis required for thermal tolerance. *Science*,
 633 315, 513-515. <https://doi.org/10.1126/science.1136237>
 634 McKane, R.B., Johnson, L.C., Shaver, G.R., Nadelhoffer, K.J., Rastetter, E.B., Fry,
 635 B., ... Laundre, J.A. (2002). Resource-based niches provide a basis for plant
 636 species diversity and dominance in arctic tundra. *Nature*, 415, 68-71.
 637 <https://doi.org/10.1038/415068a>
 638 Miller, A.E., & Bowman, W.D. (2002). Variation in nitrogen-15 natural abundance and

639 nitrogen uptake traits among co-occurring alpine species: do species partition
640 by nitrogen form? *Oecologia*, *130*, 609-616.

641 Miller, A.E., Bowman, W.D., & Suding, K.N. (2007). Plant uptake of inorganic and
642 organic nitrogen: neighbor identity matters. *Ecology*, *88*, 1832-1840.
643 <https://doi.org/10.1890/06-0946.1>

644 Nasto, M.K., Osborne, B.B., Lekberg, Y., Asner, G.P., Balzotti, C.S., Porder, S., ...
645 Cleveland, C.C. (2017). Nutrient acquisition, soil phosphorus partitioning and
646 competition among trees in a lowland tropical rain forest. *New Phytologist*, *214*,
647 1506-1517. <https://doi.org/10.1111/nph.14494>

648 Phillips, R.P., Brzostek, E., & Midgley, M.G. (2013). The mycorrhizal-associated
649 nutrient economy: a new framework for predicting carbon-nutrient couplings in
650 temperate forests. *New Phytologist*, *199*, 41-51.
651 <https://doi.org/10.1111/nph.12221>

652 Poisot, T., Mouquet, N., & Gravel, D. (2013). Trophic complementarity drives the
653 biodiversity–ecosystem functioning relationship in food webs. *Ecology Letters*,
654 *16*, 853-861. <https://doi.org/10.1111/ele.12118>

655 R Core Team (2017). *R: A language and environment for statistical computing*. R
656 Foundation for Statistical Computing, Vienna, Austria.

657 Reich, P.B., Tilman, D., Isbell, F., Mueller, K., Hobbie, S.E., Flynn, D.F., & Eisenhauer,
658 N. (2012). Impacts of biodiversity loss escalate through time as redundancy
659 fades. *Science*, *336*, 589-592. <https://doi.org/10.1126/science.1217909>

660 Reynolds, H.L., Hartley, A.E., Vogelsang, K.M., Bever, J.D., & Schultz, P. (2005).

661 Arbuscular mycorrhizal fungi do not enhance nitrogen acquisition and growth
 662 of old - field perennials under low nitrogen supply in glasshouse culture. *New*
 663 *Phytologist*, 167, 869-880. <https://doi.org/10.1111/j.1469-8137.2005.01455.x>
 664 Reynolds, H.L., Packer, A., Bever, J.D., & Clay, K. (2003). Grassroots ecology: plant-
 665 microbe-soil interactions as drivers of plant community structure and dynamics.
 666 *Ecology*, 84, 2281-2291. <https://doi.org/10.1890/02-0298>
 667 Salles, J.F., Poly, F., Schmid, B., & Roux, X.L. (2009). Community niche predicts the
 668 functioning of denitrifying bacterial assemblages. *Ecology*, 90, 3324-3332.
 669 <https://doi.org/10.1890/09-0188.1>
 670 Schmid, B., Baruffol, M., Wang, Z., & Niklaus, P.A. (2017). A guide to analyzing
 671 biodiversity experiments. *Journal of Plant Ecology*, 10, 91-110.
 672 <https://doi.org/10.1093/jpe/rtw107>
 673 Schweiger, P., & Jakobsen, I. (1999). Direct measurement of arbuscular mycorrhizal
 674 phosphorus uptake into field-grown winter wheat. *Agronomy Journal*, 91, 998-
 675 1002. <https://doi.org/10.2134/agronj1999.916998x>
 676 Simard, S.W., & Durall, D.M. (2004). Mycorrhizal networks: a review of their extent,
 677 function, and importance. *Canadian Journal of Botany*, 82, 1140-
 678 1165. <https://doi.org/10.1139/b04-116>
 679 Smith, S. E., & Read, D. J. (2008). Mycorrhizal symbiosis (3rd ed.). Cambridge, UK:
 680 Academic Press.
 681 Smith, S.E., & Smith, F.A. (2011). Roles of arbuscular mycorrhizas in plant nutrition
 682 and growth: new paradigms from cellular to ecosystem scales. *Annual Review*

683 *of Plant Biology*, 62, 227-250.

684 Steidinger, B.S., Turner, B.L., Corrales, A., & Dalling, J.W. (2015). Variability in
685 potential to exploit different soil organic phosphorus compounds among tropical
686 montane tree species. *Functional Ecology*, 29, 121-130.
687 <https://doi.org/10.1111/1365-2435.12325>

688 Taylor, A.F.S., & Alexander, I.A.N. (2005). The ectomycorrhizal symbiosis: life in the
689 real world. *Mycologist*, 19, 102-112.
690 <http://doi.org/10.1017/S0269915X05003034>

691 Teste, F.P., Kardol, P., Turner, B.L., Wardle, D.A., Zemunik, G., Renton, M., &
692 Laliberté, E. (2017). Plant-soil feedback and the maintenance of diversity in
693 Mediterranean-climate shrublands. *Science*, 355, 173-176.
694 <http://doi.org/10.1126/science.aai8291>

695 Teste, F.P., Karst, J., Jones, M.D., Simard, S.W., & Durall, D.M. (2006). Methods to
696 control ectomycorrhizal colonization: effectiveness of chemical and physical
697 barriers. *Mycorrhiza*, 17, 51-65.

698 Teste, F.P., Veneklaas, E.J., Dixon, K.W., & Lambers, H. (2014). Complementary plant
699 nutrient-acquisition strategies promote growth of neighbour species. *Functional*
700 *Ecology*, 28, 819-828. <http://doi.org/10.1111/1365-2435.12270>

701 Tilman, D. (1982) Resource competition and community structure. Princeton, New
702 Jersey: Princeton University Press.

703 Tilman, D., Lehman, C.L., & Thomson, K.T. (1997). Plant diversity and ecosystem
704 productivity: theoretical considerations. *Proceedings of the National Academy*

705 *of Sciences of the United States of America*, 94, 1857-1861.

706 Turnbull, L.A., Levine, J.M., Loreau, M., & Hector, A. (2013). Coexistence, niches and
707 biodiversity effects on ecosystem functioning. *Ecology Letters*, 16, 116-127.
708 <https://doi.org/10.1111/ele.12056>

709 van der Heijden, M.G., Bardgett, R.D., & van Straalen, N.M. (2008). The unseen
710 majority: soil microbes as drivers of plant diversity and productivity in
711 terrestrial ecosystems. *Ecology Letters*, 11, 296-310.
712 <https://doi.org/10.1111/j.1461-0248.2007.01139.x>

713 van der Heijden, M.G., Klironomos, J.N., Ursic, M., Moutoglis, P., Streitwolf-Engel,
714 R., Boller, T., ... Sanders, I.R. (1998). Mycorrhizal fungal diversity determines
715 plant biodiversity, ecosystem variability and productivity. *Nature*, 396, 69-72.
716 <https://doi.org/10.1038/23932>

717 van der Heijden, M.G., Martin, F.M., Selosse, M.A., & Sanders, I.R. (2015).
718 Mycorrhizal ecology and evolution: the past, the present, and the future. *New*
719 *Phytologist*, 205, 1406-1423. <https://doi.org/10.1111/nph.13288>

720 Vierheilig, H., Coughlan, A.P., Wyss, U., & Piché, Y. (1998). Ink and vinegar, a simple
721 staining technique for arbuscular-mycorrhizal fungi. *Applied and*
722 *Environmental Microbiology*, 64, 5004-5007.

723 Vogelsang, K.M., Reynolds, H.L., & Bever, J.D. (2006). Mycorrhizal fungal identity
724 and richness determine the diversity and productivity of a tallgrass prairie
725 system. *New Phytologist*, 172, 554-562. [http://doi.org/10.1111/j.1469-](http://doi.org/10.1111/j.1469-8137.2006.01854.x)
726 [8137.2006.01854.x](http://doi.org/10.1111/j.1469-8137.2006.01854.x)

727 Voisin, A.S., Salon, C., Munier-Jolain, N.G., & Ney, B. (2002). Quantitative effects of
 728 soil nitrate, growth potential and phenology on symbiotic nitrogen fixation of
 729 pea (*Pisum sativum* L.). *Plant and Soil*, 243, 31-42.

730 Wagg, C., Barendregt, C., Jansa, J., & van der Heijden, M.G.A. (2015).
 731 Complementarity in both plant and mycorrhizal fungal communities are not
 732 necessarily increased by diversity in the other. *Journal of Ecology*, 103, 1233-
 733 1244. <https://doi.org/10.1111/1365-2745.12452>

734 Walder, F., Niemann, H., Natarajan, M., Lehmann, M.F., Boller, T., & Wiemken, A.
 735 (2012). Mycorrhizal networks: common goods of plants shared under unequal
 736 terms of trade. *Plant Physiology*, 159, 789-797.
 737 <https://doi.org/10.1104/pp.112.195727>

738 Weigelt, A., Bol, R., & Bardgett, R.D. (2005). Preferential uptake of soil nitrogen forms
 739 by grassland plant species. *Oecologia*, 142, 627-635.

740 Wu, J., Ma, H., Xu, X., Qiao, N., Guo, S., Liu, F., ... Zhou, L. (2013). Mycorrhizas alter
 741 nitrogen acquisition by the terrestrial orchid *Cymbidium goeringii*. *Annals of*
 742 *Botany*, 111, 1181-1187. <https://doi.org/10.1093/aob/mct062>

743

744

745 **Supporting Information**

746 Details of Supporting Information are provided below.

747 **Table S1.** Root mycorrhizal colonization of each tree species.

748 **Table S2.** Relationship between plant species richness and community productivity in

749 unsterilized soil without N addition

750 **Fig. S1.** Relationship between plant species richness and community niche.

751 **Appendix S1.** Supporting analyses

752 **Table 1.** Summary of general linear model results showing the effects of plant species identity, N treatments (ammonium, nitrate or glycine
753 addition), soil-inoculum treatments (sterilized vs. unsterilized inoculum) and their interactions on the total above- and belowground biomass of
754 plant individuals. The preferred N form of each species was considered as its “home” N form.

Source of variation		Plant biomass			
		df	%SS	<i>F</i>	<i>P</i>
Plant species identity (ID)		2	13.02	20.76	<0.001
N treatments (N)		2	0.65	1.03	0.360
Soil inoculum treatments (S)		1	0.02	0.08	0.781
ID × N		4	3.59	2.87	0.024
	<i>Home</i>	<i>1</i>		<i>3.34</i>	<i>10.69</i>
	<i>ID × N</i>	<i>3</i>		<i>0.25</i>	<i>0.26</i>
ID × S		2	0.63	0.99	0.373
N × S		2	2.37	3.78	0.024
ID × N × S		4	3.20	2.55	0.040
	<i>Home × S</i>	<i>1</i>		<i>2.82</i>	<i>9.00</i>
	<i>ID × N × S</i>	<i>3</i>		<i>0.38</i>	<i>0.40</i>
Residuals		244	76.53		

755 Key to abbreviations: df, degree of freedom; %SS, percentage of the total sum of squares explained by each term; the numbers in bold indicate
756 significant effects; in italics are the decompositions of previous terms (namely ID × N and ID × N × S) and corresponding statistical values.

757

758 **Table 2.** Results of general linear models presenting the effects of block, soil-inoculum treatments (sterilized vs. unsterilized inoculum) and
759 community composition (different combination of species) on NE, CE and SE in soil supplied with a mix of N forms. The interaction term soil-
760 inoculum treatments \times community composition was not significant and therefore omitted from the analysis.

Source of variation	df	NE			CE			SE		
		SS %	<i>F</i>	<i>P</i>	SS %	<i>F</i>	<i>P</i>	SS %	<i>F</i>	<i>P</i>
Intercept	1	16.48	14.98	<0.001	18.47	18.78	<0.001	11.00	10.88	0.002
Block	9	17.72	1.79	0.090	18.26	2.06	0.048	21.42	2.35	0.025
Soil-inoculum treatments	1	1.80	1.64	0.206	4.17	4.24	0.044	4.18	4.13	0.047
Community composition	3	1.29	0.39	0.759	3.03	1.03	0.387	5.73	1.89	0.144
Residuals	57	62.71			56.06			57.67		

761 Key to abbreviations: df, degree of freedom; %SS, percentage of the total sum of squares explained by each variable; the numbers in bold indicate
762 significant effects.

763 **Table 3.** Results of general linear model analyzing the relationships between plant
764 species richness and community productivity (the total of plant above- and
765 belowground biomass) in soils supplied with a mix of N forms but with different soil-
766 inoculum treatments (sterilized vs. unsterilized inoculum). Overall analysis (Model A)
767 followed by separate analyses for soils with sterilized (Model B) and unsterilized
768 inoculum (Model C).

Model A: overall model				
Source of variation	Community productivity			
	df	% SS	<i>F</i>	<i>P</i>
Block	9	7.20	1.82	0.072
Soil-inoculum treatments (S)	1	7.46	16.96	<0.001
Plant species richness (SR)	1	2.51	5.70	0.019
S × SR	1	0.34	0.77	0.381
Community composition	5	30.11	13.68	<0.001
Residuals	119	52.38		

Model B: sterilized inoculum				
Source of variation	Community productivity			
	df	% SS	<i>F</i>	<i>P</i>
Block	9	12.44	1.54	0.159
Plant species richness (SR)	1	1.40	1.55	0.217
Community composition	5	40.40	9.00	<0.001
Residuals	51	45.77		

Model C: unsterilized inoculum				
Source of variation	Community productivity			
	df	% SS	<i>F</i>	<i>P</i>
Block	9	21.76	2.80	0.009
Plant species richness (SR)	1	3.78	4.38	0.041
Community composition	5	27.78	6.43	<0.001
Residuals	54	46.67		

769 Key to abbreviations: df, degree of freedom; %SS, percentage of the total sum of
770 squares explained by each variable; the numbers in bold indicate significant effects.

771

772 **Table 4.** Results of general linear model analysis presenting the effect of community
 773 niche (CN) on community productivity (the total of plant above- and belowground
 774 biomass) in unsterilized soil supplied with a mix of N forms.

Source of variation	Community productivity			
	df	% SS	<i>F</i>	<i>P</i>
Block	9	21.76	2.80	0.009
Community niche	1	3.57	4.13	0.047
Community composition	5	28.00	6.48	<0.001
Residuals	54	46.67		

775 Key to abbreviations: df, degree of freedom; %SS, percentage of the total sum of
 776 squares explained by each variable; the numbers in bold indicate significant effects.

777

778

779

780

781

782

783

784

785

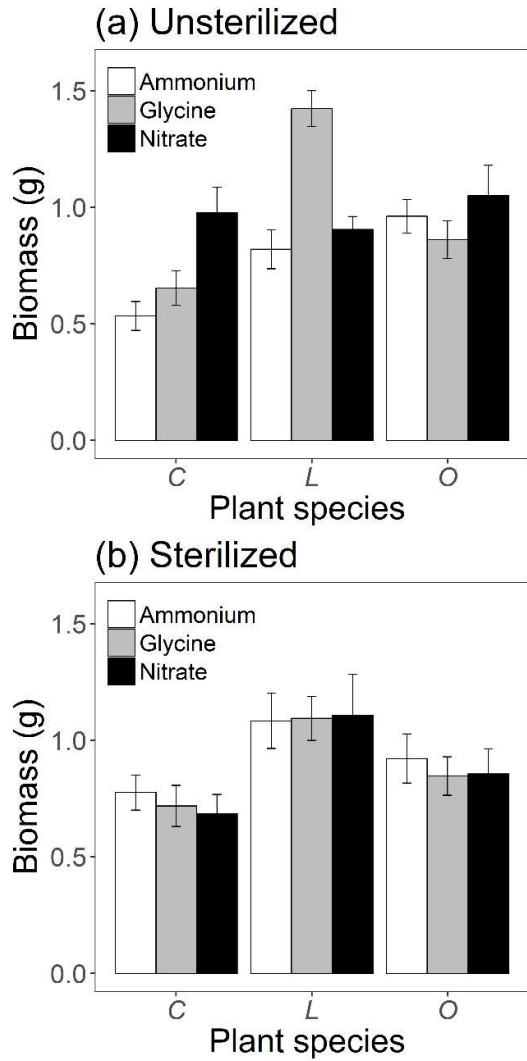
786

787

788

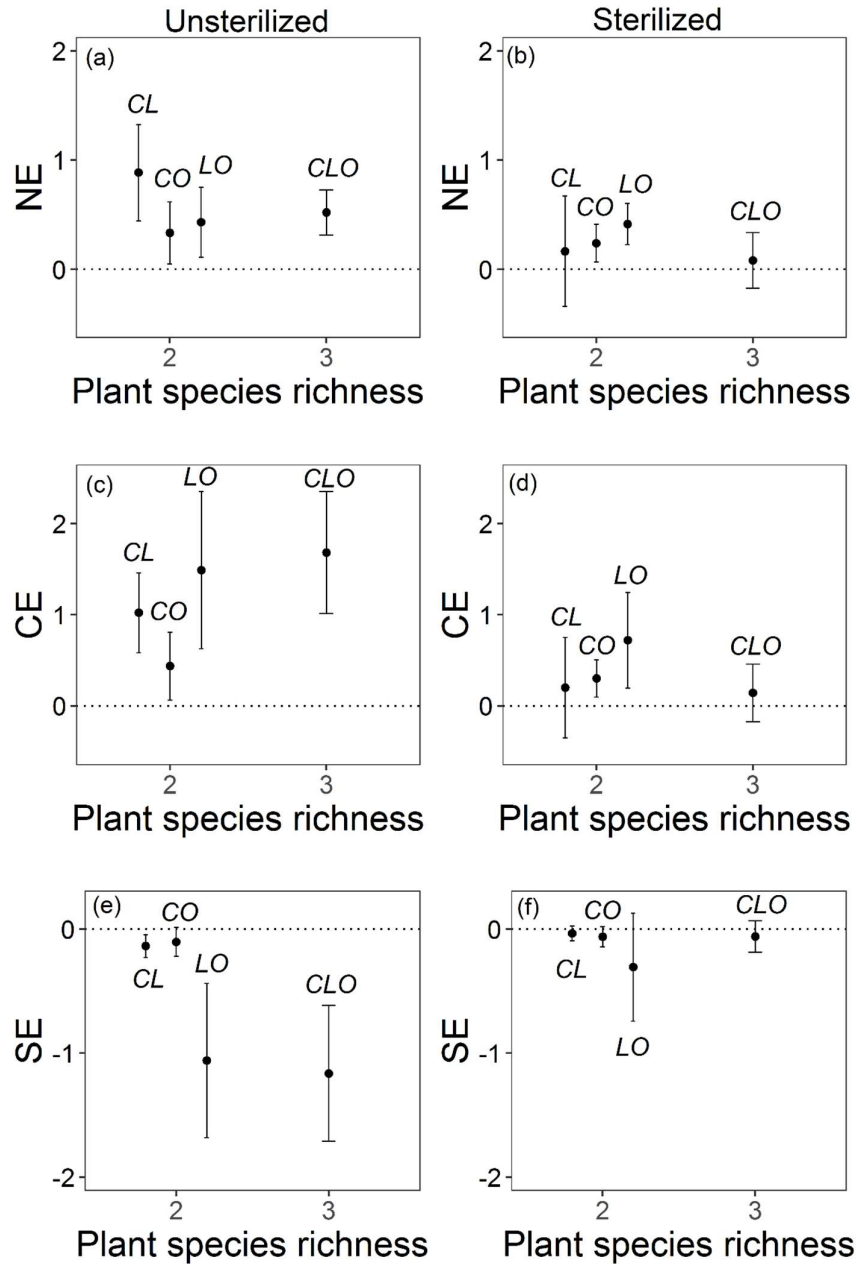
789

790



791

792 **Fig. 1.** Mean biomass (the total of above- and belowground biomass dry weight) of
 793 plant individuals in soils supplied with different N forms (ammonium, glycine or nitrate)
 794 but with either unsterilized (a) or sterilized inoculum (b). Data shown are means \pm SEM
 795 ($N = 15$). Species code: C, *Cryptocarya concinna*; L, *Lithocarpus litseifolius*; O,
 796 *Ormosia glaberrima*.



797
 798 **Fig. 2.** Net effect (NE; a-b), complementarity effect (CE; c-d) and selection effect (CE;
 799 e-f) of different plant species composition (2 or 3 species) in soils supplied with a mix
 800 of N forms but with either unsterilized (left panel) or sterilized inoculum (right panels).
 801 Data shown are means \pm SEM ($N = 10$). Species code: C, *Cryptocarya concinna*; L,
 802 *Lithocarpus litseifolius*; O, *Ormosia glaberrima*.

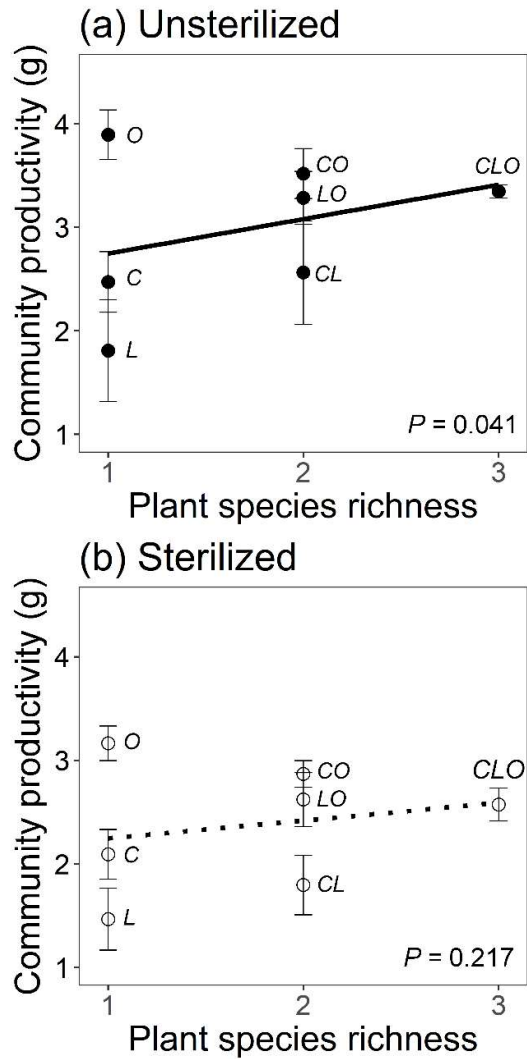


Fig. 3. Relationships between plant species richness and community productivity (the total of plant above- and belowground biomass) in soils supplied with a mix of N forms but with either unsterilized (a) or sterilized inoculum (b). Solid line indicates significant ($P < 0.05$) and dotted line indicates insignificant ($P > 0.2$) relationship between plant species richness and community productivity (see Table 3). The mean (\pm SE) is shown for each community composition under different treatments ($N = 10$). Species code: C, *Cryptocarya concinna*; L, *Lithocarpus litseifolius*; O, *Ormosia glaberrima*.

Functional Ecology

Supporting Information

Appendix S1. Supporting analyses

Soil microbes promote complementarity effects among co-existing trees through soil nitrogen partitioning

Shan Luo¹, Bernhard Schmid², Gerlinde B. De Deyn³, Shixiao Yu^{1*}

¹Department of Ecology, School of Life Sciences/State Key Laboratory of Biocontrol, Sun Yat-sen University, Guangzhou 510275, China

²Department of Evolutionary Biology and Environmental Studies, University of Zürich, Winterthurerstrasse 190, CH-8057 Zürich, Switzerland

³Department of Environmental Sciences, Wageningen University, P.O. Box 47, 6700 AA Wageningen, The Netherlands

* Corresponding author:

Prof. Dr. Shixiao YU

Email: lssysx@mail.sysu.edu.cn

Table S1. Percentage of root mycorrhizal colonization of each tree species under unsterilized and sterilized soils.

Species	Unsterilized	Sterilized
<i>Cryptocarya concinna</i>	73.7 % \pm 4.8%	9.3% \pm 2.3%
<i>Lithocarpus litseifolius</i>	47.0 % \pm 6.0%	5.8% \pm 1.1%
<i>Ormosia glaberrima</i>	77.3 % \pm 6.7%	9.1% \pm 1.1%

Table S2. Result of general linear model analyzing the relationships between plant species richness and community productivity (the total of plant above- and belowground biomass) in unsterilized soil without N addition.

Source of variation	Community productivity			
	df	% SS	<i>F</i>	<i>P</i>
Block	9	29.03	3.40	0.002
Plant species richness (SR)	1	0.02	0.03	0.876
Community composition	5	21.63	4.56	0.002
Residuals	52	49.32		

Key to abbreviations: df, degree of freedom; %SS, percentage of the total sum of squares explained by each variable; the numbers in bold indicate significant effects.

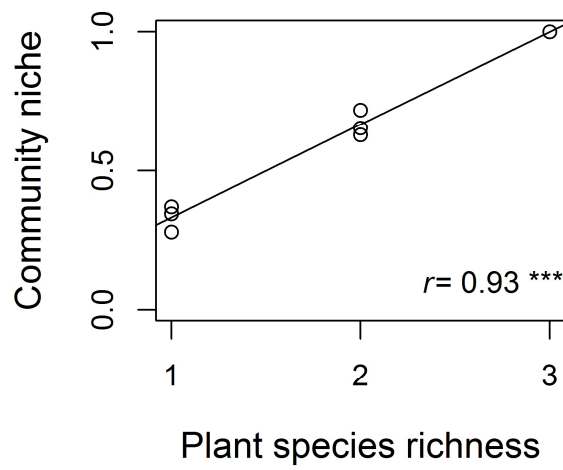


Fig. S1. Relationship between plant species richness and community niche.

Significance level for correlation coefficient $\neq 0$: ***, $P < 0.001$