Supporting Information 1

- Trophic transfer of anticoagulant rodenticides while managing rodent pests: the fine line between predator-prey regulation and pesticide-pest regulation
- Virgile Baudrot^{1,2} Javier Fernandez-de-Simon¹, 3 Michael Coeurdassier¹,
 Geoffroy Couval¹, 4, Patrick Giraudoux¹, Xavier Lambin⁵, 6
- $_{8}$ 1 Université Bourgogne Franche-Comté UMR CNRS 6249 Laboratoire Chrono-environnement,
- 9 25030 Besançon, France

1

- 2 BioSP, INRA, 84000 Avignon, France
- ³ IREC. Instituto de Investigación en Recursos Cinegéticos
- ⁴ FREDON Franche-Comté, Espace Valentin Est, 12, Rue de Franche-Comté Bât E, 25480
- 13 Ecole-Valentin, France

16

- ¹⁴ School of Biological Sciences, University of Aberdeen, Zoology Building, Tillydrone Avenue,
- 15 Aberdeen AB24 2TZ, Scotland, UK

* Corresponding authors: virgile.baudrot@posteo.net or x.lambin@abdn.ac.uk

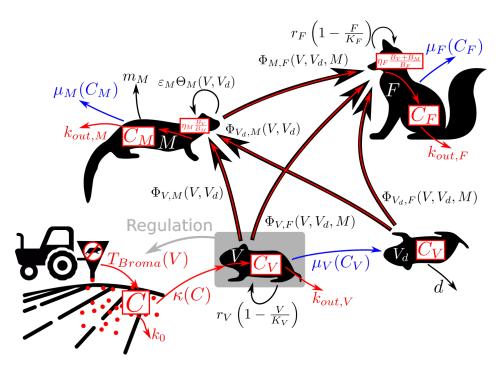


Figure S1 – Dynamics of mustelids (bold grey line) and red foxes (bold black line) after 50 years of simulation (10 years of burn-in period, shaded area, and 40 years considered here for the compute of results, estimation of cost functions, etc.). Letters at the top-right part of each sub-graph corresponds to the results obtained from each farmer functional response. FR (dashed line) and MR (dotted line) indicate the farmer-regulated and mustelid-regulated periods of vole population dynamics.

18 1 Model derivation

19 1.1 Dynamic of populations for the tri-trophic dynamic

20 The population dynamics of a tri-trophic system is commonly described as follow:

$$\begin{cases}
\frac{dV}{dt} = r_V V \left(1 - \frac{V}{K_V} \right) - \Phi_{V,M}(V) M - \Phi_{V,F}(V,M) F \\
\frac{dM}{dt} = \varepsilon_M \Phi_{V,M}(V) M - m_M M - \Phi_{M,F}(V,M) F \\
\frac{dF}{dt} = F r_F \left(1 - \frac{F}{K_F} \right)
\end{cases}$$
(S1)

As described in the manuscript, the vole population, V followed a logistic growth rate, with r_V the maximal reproduction rate, fixed at $r_V = \ln(2 \times 600)/365$ in [day⁻¹], since montane water vole populations can increase from 0 to 600 individuals ha⁻¹ or more in a year [Giraudoux et al., 1997] resulting in the equilibrium density being fixed at $K_V = 600$ individuals.

1.1.1 About the logistic model

33

34

36

37

42

43

45

49

51

52

55

The idea behind this parametrization is the use of a classical Malthusian equation with r_V in $[day^{-1}]$ and V in $[ind\ ha^{-1}]$:

$$\frac{dV}{dt} = r_V V \quad \Rightarrow \quad V(t) = V(0) \exp(r_V t) \quad \Rightarrow \quad r_V = \ln\left(\frac{V(t)}{V(0)}\right) \times \frac{1}{t}$$
 (S2)

So to have the maximal growth rate r_V , we need the greatest V(t) (measured at 600 ind ha⁻¹ but supposed to be more in theoretical optimal condition) obtained in the minimum of time (less than a year). Then, to add the carrying capacity K_V in [ind⁻¹], we use the logistic equation and assumed the carrying capacity to be the one measured in the environment:

$$\frac{dV}{dt} = r_V V \left(1 - \frac{V}{K_V} \right) \tag{S3}$$

The estimation of r using the Malthusian function and then applying the carrying capacity K can be seen as non consistent because of the dependency of r_V to K_V . For isntance, in Ginzburg [1992], the author says that going from $dV/dt = r_V V$ to $dV/dt = r_V V(1 - V/K_V)$ would be correct if r was independent of resource limitation K_V :

"Two populations of a species live in environments identical in all respects, that is, they have the same resource availability and any other imaginable characteristics. One of the two is subject to higher mortality. There is no doubt that the population growth curves in these two cases will be different. Are the final equilibrium abundances different? Most ecologists will answer that the equilibrium values should be the same and that the higher rate of reproduction just means that the population will 'get there faster', but reach the same level nevertheless".

Then, Ginzburg [1992] states that their is two ways to reflect the extra mortality: either assuming (a) $dV/dt = r_V V(1 - V/K_V) - mV$ or (b) $dV/dt = (r_V - m_V)V(1 - V/K_V)$. In equation (a), the equilibrium depend on r: $V^* = (r_V - m_V)K_V/r_V$, so is wrong with the un-changed equilibrium assumption. And re-writing (b) gives: $dV/dt = rV(1 - V/K_V) - m_V V(1 - V/K_V)$. Since $-m_V V(1 - V/K_V)$ reflects an environmental pressure, it should not be positive when $V > K_V$. Meaning, when the population is over it's carrying capacity, we are going to growth even faster and to infinity. In the present manuscript, we made assumption (a), and we assume that r is not independent of resource limitation K_V . It is for instance well known that vole reproduction rate is linked to the state of resource at the time of reproduction [?].

Many authors respond to Ginzburg [Olson, 1992, Watkinson, 1992, Mackenzie, 1992, Berryman, 1992] and later [Gabriel et al., 2005].

In his response, Olson [1992] says: "While the logistic model is often defined as non-mechanistic, it has the attractive feature of describing population growth with only two parameters, K and r. K is universally defined as the population-carrying capacity. However, there is much disagreement about the definition of r and this disagreement has affected conclusions

derived from the use of the logistic equation." Olson [1992] suggests that the problem is the intuition assigned to the r-independence to equilibirum density K. Again in our line, Watkinson [1992] argues that "Most ecologists, he asserts, would believe that the equilibrium values of these two populations should be the same. In that case I must be one of the minority who does not believe this to be so."

Another answer by Berryman [1992] proposed to rewrite the model in order to see the link between r and K. The approach by Berryman [1992] is interesting as it introduces a relevant effect of considering the exploitation rate (link to the growth rate) within the carrying capacity, meaning a carrying capacity with its on dynamic (e.g., with its resilience depending on how exploitation happen), and non an extra-system static one.

69 1.1.2 Predation of voles, and predator dynamics

The vole population was preyed upon by mustelid and fox populations, denoted M and F respectively in (S1). The vole consumption rate at different vole densities was described by functional responses ($\Phi_{V,M}$ for mustelids, $\Phi_{V,F}$ for foxes).

$$\Phi_{V,M}(V)M = \frac{a_M V}{1 + h_M a_M V}
\Phi_{V,F} = \frac{a_{VF} V}{a_{VF} V + a_{MF} F} \times \frac{(a_{VF} V + a_{MF} M)^2}{1 + h_F (a_{VF} V + a_{MF} M)^2}
\Phi_{M,F}(V,M)F = \frac{a_{MF} V}{a_{VF} V + a_{MF} F} \times \frac{(a_{VF} V + a_{MF} M)^2}{1 + h_F (a_{VF} V + a_{MF} M)^2}$$
(S4)

We treated small mustelids as vole specialist predators [King and Powell, 2006], assuming a Holling Type 2 functional response with attack rate a_M in [day $^-$ 1] and handling time h_M in [day] (equation (S4)). We then represented foxes feeding on voles and mustelids by a multi-species functional response derived from Holling Type 3, referring to generalist feeding behaviour [Baudrot et al., 2016]. For that, we denoted a_{VF} and a_{MF} (both in [day $^-$ 1]) the fox attack rate on voles and mustelids respectively. The parameter h_F was the handling time for foxes in [day].

Consequently, the functional responses $\Phi_{X,Y}$ are all in [day $^{-1}$]. And the functions: $\frac{dV}{dt} = r_V V \left(1 - \frac{V}{K_V}\right) - \Phi_{V,M}(V) M - \Phi_{V,F}(V,M) F \text{ and } \frac{dF}{dt} = F r_F \left(1 - \frac{F}{K_F}\right) \text{ are also consistent, all part being in [ind ha}^{-1} \text{ day}^{-1}].$

3 1.1.3 Conversion efficiency and mortality rate in mustelid population

The parameterization of the conversion efficiency of ingested food into new born, ε_M in [n.d.] (non dimensional), and the mortality rate, m_M in [day⁻¹], in equation (S1) is not straightforward since we have to take into account the mortality rate due to starvation. In the situation of no starvation, ageing, diseases and predation of mustelids are the main reason

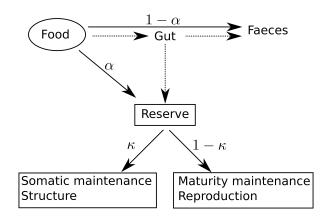


Figure S2 – Representation of the standard dynamic energy budget (DEB model [Sousa et al., 2010]

of death. In the situation of no prey availability at initial condition, then the mortality rate of mustelids should be equal to the mortality rate without food. Using the simplest scheme of Dynamic Energy Budget (DEB) theory (seeFigure S2), the ingested food is assimilated for a part, we denote α and non-assimilated for another part, $(1 - \alpha)$. Then, the assimilated part goes to a reserve (or directly use) in either the somatic maintenance or in the maturity maintenance (to produce reproductive tools). In DEB, a proportion κ from the reserve is the part allocated to somatic maintenance [n.d.] and a proportion $1 - \kappa$ is addressed to the maturity maintenance.

Figure S2 represents the standard dynamic energy budget (DEB) model [Sousa et al., 2010]. Then, assuming the generic functional response $\Phi_{V,M}(V)$, the reproduction rate of the predator is like:

$$\Phi_{V,M}(V)\alpha(1-\kappa)\varepsilon'$$
 (S5)

where ε' in [n.d.] is the conversion efficiency of daily assimilated food for one individual, dedicated for reproduction into newborn mustelids.

Then, $\alpha\kappa\Phi_{V,M}$ is the part allocated to somatic maintenance structure. Parameters α and κ are non-dimensional [n.d.].

We can then take into account the starvation process in the mortality rate, with a foodindependent mortality rate $m_{M,-\Phi}$ in [day⁻¹] and a part depending on food availability $m_{M,+\Phi} = \sigma \alpha \kappa \Phi_{V,M}$ in [day⁻¹], with σ [n.d.] a conversion parameter.

Then, the dynamic of Mustelid given in the manuscript equation (S1) could be rewritten as:

$$\frac{dM}{dt} = \Phi_{V,M}(V)\alpha(1-\kappa)\varepsilon'M - (m_{M,-\Phi} + \sigma\alpha\kappa\Phi_{V,M})M - \Phi_{M,F}(V,M)F$$
 (S6)

Which is equal to:

108

$$\frac{dM}{dt} = \left(\alpha(1 - \kappa)\varepsilon' - \sigma\alpha\kappa\right)\Phi_{V,M}(V)M - m_{M,-\Phi}M - \Phi_{M,F}(V,M)F \tag{S7}$$

Therefore, if we set the food-independent mortality rate $m_M=m_{M,-\Phi}$ and the conversion efficiency of ingested food $\varepsilon_M=\alpha((1-\kappa)\varepsilon'-\sigma\kappa)$, we get back to the equation provided in model (S1) but with a clearer definition of ε_M .

This is why we parameterized m_M in [day⁻¹] as the inverse of life expectancy in [day⁻¹], and ε_M in [n.d.] as an unclear parameter derived from the null model (i.e., without AR) to reflect observed population dynamics.

In any case, the dimension of the equation $\frac{dM}{dt} = \varepsilon_M \Phi_{V,M}(V) M - m_M M - \Phi_{M,F}(V,M) F$ is well defined, since $\varepsilon_M \Phi_{V,M}(V)$ is in [day⁻¹].

117 1.2 Dynamic of populations with AR

After spreading by farmer, in the model AR is degraded or transferred to voles, which are exposed to the environmental contaminant mainly through ingestion of contaminated food-stuffs (i.e., trophic transfer), and AR accumulates in their tissues and is release with rate $k_{out,V}$. Contaminants in an individual, the body burden, are commonly measured in concentration per mass of individual, [μ g g⁻¹ ind⁻¹] (equivalent to [mg kg⁻¹ ind⁻¹] and denoted [ppm ind⁻¹]). Predators are only exposed to contaminant through the consumption of voles, a trophic transfer, so as a function of functional response.

To consider both toxicological and population dynamics within the same equation, we convert population dynamics into dynamics of their biomasses as in [Huang et al., 2015,

128 Baudrot et al., 2018].

112

113

114

125

We denote V, V_d M, F the densities of voles, dead voles, mustelids and foxes respectively ; B_V , B_{V_d} , B_M and B_F the mean biomass of an individual of those species in [g ind⁻¹], and $B_{T,V}$, B_{T,V_d} , $B_{T,M}$ and $B_{T,F}$ the total biomass of the population in the space unit we used, so in [g ha⁻¹]. In other words, $B_{T,X} = B_X \times X$. The notation C_V , C_{V_d} C_M and C_F holds for the mean concentration of the contaminant in one individual, commonly called the body burden of the prey in ppm [mg kg⁻¹]. The growth function of the prey population is a function $g_X(X,C_X)$ in [day⁻¹] depending on the population density X and the concentration of the contaminant C_X . The dose-response curve $\mu_X(C_X)$ is defined by a log-normal cumulative distribution function as in Loos et al. [2010].

Variables for the population dynamics To convert individual dynamics into biomass dynamics, we just multiplied the variable density X in [ind ha⁻¹] with a constant B_X in [g ind⁻¹], so that the derivative as just to be multiplied by this constant:

$$\frac{dB_{T,X}}{dt} = \frac{dB_XX}{dt} = B_X \frac{dX}{dt} \quad \text{and so we have} \quad \frac{dX}{dt} = \frac{1}{B_X} \times \frac{dB_{T,X}}{dt}$$
 (S8)

Variable for the toxicant dynamics For the dynamic of toxicants, we consider an homoge-141 neous population. 142

We set a new variable $w_X = C_X B_X X$ which is in $[\mu g g^{-1}] \times [g \text{ ind}^{-1}] \times [\text{ind ha}^{-1}]$, so 143 that: w_X is in [mg ha⁻¹] or [mg population⁻¹], which is the quantity of contaminant within 144 the population. 145

The dynamic of the new variable is defined with:

146

147

154

155

156

157

159

166

$$w_X = C_X B_X X = C_X B_{T,X} \quad \Rightarrow \quad \frac{dw_X}{dt} = B_X \frac{dX C_X}{dt} = B_X \left(X \frac{dC_X}{dt} + C_X \frac{dX}{dt} \right)$$
 (S9)

As a consequence, $\frac{dw_X}{dt}$ is in [mg ha⁻¹ day⁻¹].

1.2.1 Dynamic of the contaminant in the environment

In the agro-ecotoxico-logical system, we consider an environmental concentration of contaminant denoted by w_{ext} in [mg ha⁻¹] in order to reflect the spatially explicit spreading of AR following determined in baits per area. From the Figure S1, we can directly provide the 151 dynamic of this concentration: 152

$$\frac{dw_{ext}}{dt} = T_{Broma}(V) - k_0 w_{ext} - \kappa(w_{ext}) V B_V \tag{S10}$$

with $T_{Broma}(V)$ the farmer input of AR in [mg ha⁻¹ day⁻¹] depending on vole density 153 denoted V in [ind ha⁻¹]. AR concentration in baits is 50 μ g g⁻¹ (or mg kg⁻¹). So when farmer spread in grasslands at quantity 7.5 to 20 kg ha⁻¹ day⁻¹, its an amount of 375 to 1000 mg ha⁻¹ of AR spread (i.e., mg kg⁻¹ \times kg ha⁻¹ day⁻¹).

The disappearance of AR in the field, denoted k_0 , in [day⁻¹], so that $k_0 w_{ext}$ is in [mg ha⁻¹ day $^{-1}$]. Such quantity, w_{ext} , was available in the literature, and a proportion disappeared in the environment at rate $k_0=0.0815~{\rm day^{-1}}$ [Sage et al., 2008].

The function $\kappa(w_{ext})$, in [mg kg⁻¹ day⁻¹], reflects the transfer of contaminant from the en-160 vironmental compartment to the vole compartment. As a consequence, the term $\kappa(w_{ext})VB_V$ 161 is in $\lceil mg \ ha^{-1} \ day^{-1} \rceil$ ($\lceil mg \ kg^{-1} \ day^{-1} \rceil \times \lceil kg \ ha^{-1} \rceil$). 162

This rate $\kappa(w_{ext})$ was assumed to be an increasing function characterized by a maximum 163 intake rate M_{in} in [mg kg⁻¹ day⁻¹], and a half-saturation constant for ingestion D_{in} in [mg ha^{-1}] since w_{ext} is in [mg ha^{-1}]:

$$\kappa(w_{ext}) = M_{in} \times \frac{w_{ext}}{D_{in} + w_{ext}}$$
 (S11)

So the dimension of the ingestion rate $\kappa(w_{ext})$ is the same as M_{in} that is [mg kg⁻¹ day⁻¹].

1.2.2 Dynamics of contaminant within populations

- 168 With the previous assumption, the general model is described as follow:
- Population dynamics From the equation (S8), we have: $\frac{dB_{T,X}}{dt} = B_X \frac{dX}{dt}$

Convertion of w_V into C_V

$$\frac{dC_V}{dt} = \frac{dw_V/B_{T,V}}{dt} = \frac{1}{B_{T,V}^2} \left(\frac{dw_V}{dt} B_{T,V} - w_V \frac{dB_{T,V}}{dt} \right)$$
(S12)

$$\begin{cases} \frac{dw_{ext}}{dt} &= T_{Broma}(V) - k_0 w_{ext} - \kappa(w_{ext}) B_{T,V} \\ \frac{dB_{T,V}}{dt} &= B_{T,V} g_V(V, C_V) - \mu_V(C_V) B_{T,V} - B_V \Phi_{V,M}(V) M - B_V \Phi_{V,F}(V, M) F \\ \frac{dw_V}{dt} &= \kappa(w_{ext}) B_{T,V} - k_{out,V} w_V - w_V \mu(C_V) - C_V B_V \Phi_{V,M}(V) M + \gamma_V w_V g_V(V, C_V) \\ \frac{dB_{T,V_d}}{dt} &= \mu_V(C_V) B_{T,V} \\ \frac{dw_V_d}{dt} &= \frac{dw_V}{dt} \\ \frac{dB_{T,M}}{dt} &= B_{T,M}(\varepsilon_M B_V \Phi_{V,M}(V) M - m_M - \mu_M(C_M)) - B_M \Phi_{M,F}(V, M) F \\ \frac{dw_M}{dt} &= -k_{out,M} w_M - w_M(m_M + \mu(C_M)) + \eta_M C_M B_M \Phi_{V,M}(V) M - C_M B_M \Phi_{M,F}(V, M) F + \gamma_M w_M g_M(M, C_V) \\ \frac{dB_{T,F}}{dt} &= B_{T,F} g_F(F, C_F) - \mu_F(C_F) \\ \frac{dw_F}{dt} &= -k_{out,F} w_F - w_F \mu(C_F) + \eta_F C_M B_M \Phi_{M,V} F + \eta_F C_V B_V \Phi_{V,V} F + \gamma_F w_F g_F(F, C_F) \end{cases}$$
(S13)

Thus, it is straightforward to express the dynamics as:

$$\begin{cases} \frac{dV}{dt} &= \frac{growth\ rate}{r_{V}V(1-V/K_{V})} - \frac{poisoning}{\mu_{V}(C_{V})V} - \frac{predation}{\Phi_{V,M}(V,V_{d})M} - \frac{predation}{\Phi_{V,F}(V,V_{d},M)F} \\ \frac{dV_{d}}{dt} &= \frac{new\ dead\ voles}{\mu_{V}(C_{V})V} - \frac{predation}{\Phi_{V_{d},M}(V,V_{d})M} - \frac{predation}{\Phi_{V_{d},F}(V,V_{d},M)F} - \frac{dV_{d}}{dV_{d}} \\ \frac{dM}{dt} &= \frac{growth\ rate}{(\varepsilon_{M}\Theta_{M}(V,V_{d})-m_{M})M} - \frac{poisoning}{\mu_{M}(C_{M})M} - \frac{predation}{\Phi_{M,F}(V,V_{d},M)F} \\ \frac{dF}{dt} &= \frac{growth\ rate}{r_{F}F(1-F/K_{F})} - \frac{p_{F}(C_{F})F}{\mu_{F}(C_{F})F} \\ \frac{dC}{dt} &= T_{Broma}(V) - \frac{k_{0}C}{k_{0}C} - \frac{\kappa(C)}{\kappa(C)} \\ \frac{dC_{V}}{dt} &= \frac{intake}{R} - \frac{excretion}{R} \\ \frac{dC_{W}}{dt} &= \frac{growth\ rate}{R} - \frac{excretion}{R} - \frac{excretion}{R} \\ \frac{dC_{W}}{dt} &= \frac{growth\ rate}{R} - \frac{growth\ rate}{R} - \frac{excretion}{R} - \frac{ex$$

71 References

170

172 Virgile Baudrot, Antoine Perasso, Clémentine Fritsch, Patrick Giraudoux, and Francis Raoul.

The adaptation of generalist predators' diet in a multi-prey context: insights from new

functional responses. *Ecology*, 97(7):1832–1841, 2016.

175 Virgile Baudrot, Clémentine Fritsch, Antoine Perasso, Malay Banerjee, and Francis Raoul.

Effects of contaminants and trophic cascade regulation on food chain stability: Application

to cadmium soil pollution on small mammals-raptor systems. Ecological Modelling, 382:

178 33–42, 2018.

Alan A Berryman. Intuition and the logistic equation. *Trends in ecology & evolution*, 7(9):316,

180 1992.

Jean-Pierre Gabriel, Francis Saucy, and Louis-FÃ(c)lix Bersier. Paradoxes in the logistic

- equation? Ecological Modelling, 185(1):147 151, 2005. ISSN 0304-3800. doi
- http://dx.doi.org/10.1016/j.ecolmodel.2004.10.009. URL http://www.sciencedirect.
- com/science/article/pii/S030438000400554X.
- Lev R Ginzburg. Evolutionary consequences of basic growth equations. *Trends in ecology & evolution*, 7(4):133, 1992.
- P Giraudoux, P Delattre, M Habert, JP Quéré, S Deblay, R Defaut, R Duhamel, MF Moissenet,
- D Salvi, and D Truchetet. Population dynamics of fossorial water vole (arvicola terrestris
- scherman): a land use and landscape perspective. Agriculture, Ecosystems & Environment,
- 190 66(1):47–60, 1997.
- Qihua Huang, Hao Wang, and Mark A Lewis. The impact of environmental toxins on predatorprey dynamics. *Journal of Theoretical Biology*, 378:12–30, 2015.
- ¹⁹³ Carolyn M King and Roger A Powell. *The natural history of weasels and stoats: ecology, behavior,*¹⁹⁴ *and management.* Oxford University Press, 2006.
- 195 Mark Loos, Ad MJ Ragas, Rinus Plasmeijer, Aafke M Schipper, and A Jan Hendriks. Eco-space:
- an object-oriented, spatially explicit model to assess the risk of multiple environmental
- stressors on terrestrial vertebrate populations. Science of the Total Environment, 408(18):
- 198 3908–3917, 2010.
- Aulay Mackenzie. Intuition and the logistic equation. *Trends in ecology & evolution*, 7(9):316, 1992.
- Mark H Olson. Intuition and the logistic equation. *Trends in ecology & evolution*, 7(9):314, 1992.
- 203 Mickaël Sage, Michaël Cœurdassier, Régis Defaut, Frédéric Gimbert, Philippe Berny, and
- 204 Patrick Giraudoux. Kinetics of bromadiolone in rodent populations and implications for
- predators after field control of the water vole, arvicola terrestris. Science of the total envi-
- 206 ronment, 407(1):211-222, 2008.
- ²⁰⁷ Tânia Sousa, Tiago Domingos, J-C Poggiale, and SALM Kooijman. Dynamic energy budget
- theory restores coherence in biology. Philosophical Transactions of the Royal Society B: Bio-
- logical Sciences, 365(1557):3413-3428, 2010.
- Andrew Watkinson. Intuition and the logistic equation. *Trends in ecology & evolution*, 7(9): 316, 1992.