

1 Spring haul-out behavior of seals in the Bering 2 and Chukchi seas

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14 ABSTRACT

15 Ice-associated seals rely on sea ice for a variety of activities, including breeding, molting, pupping, and resting. In
16 the Arctic, many of these activities occur in spring (April – June) as sea ice begins to melt and retreat northward.
17 Rapid acceleration of climate change in Arctic ecosystems is therefore of concern as the quantity and quality
18 of suitable habitat is forecast to decrease. Improved estimates of seal population abundance are needed to
19 properly monitor the impacts of these changes over time. In this paper, we use hourly percent-dry data from
20 satellite-linked bio-loggers deployed between 2005 and 2020 to quantify the proportion of seals hauled out
21 on ice, which is needed to accurately estimate abundance from aerial survey counts of ice-associated seals
22 (i.e., to correct for the proportion of animals that are in the water while surveys are conducted). In addition to
23 providing essential data for survey ‘availability’ calculations, our analysis also provides insights into the seasonal
24 timing and environmental factors affecting haul-out behavior by ice-associated seals. We specifically focused
25 on bearded (*Erignathus barbatus*), ribbon (*Histriophoca fasciata*), and spotted seals (*Phoca largha*) in the Bering
26 and Chukchi seas. Because ringed seals (*Phoca hispida*) can be out of the water but hidden from view in snow
27 lairs, they were not included in this analysis. Using generalized linear mixed pseudo-models to properly account
28 for temporal autocorrelation, we fit models with covariates of interest (e.g., day-of-year, solar hour, age-sex
29 class, wind speed, barometric pressure, temperature, precipitation) to examine their ability to explain variation
30 in haul-out probability. We found evidence for strong diurnal and within-season patterns in haul-out behavior,
31 as well as strong weather effects (particularly wind and temperature). In general, seals were more likely to haul
32 out on ice in the middle of the day and when wind speed was low and temperatures were higher. Haul-out
33 probability increased through March and April, peaking in May and early June before declining again. The timing
34 and frequency of haul-out events also varied based on species and age-sex class. For ribbon and spotted seals,
35 models with year effects were highly supported, indicating that the timing and magnitude of haul-out behavior
36 varied among years. However, we did not find evidence that haul-out timing was linked to annual sea ice extent.
37 Our analysis emphasizes the importance of accounting for seasonal and temporal variation in haul-out behavior,
38 as well as associated environmental covariates, when interpreting the number of seals counted in aerial surveys.

INTRODUCTION

The global climate disruption is causing considerable reduction in Arctic sea ice extent, volume, and seasonal presence (Meier et al., 2014; Wang et al., 2017; Kwok, 2018; Overland, 2021). These changes have tangible effects on Arctic organisms, ecosystems, and the human communities who live in the region (Huntington et al., 2020). Such disruptions are a particular cause of concern for the ice-associated seals that depend on spring and early summer sea ice (March-June) in the Bering and Chukchi seas as a platform for important life history functions, such as pupping, nursing, breeding behavior, and molting (Boveng et al., 2009, 2013; Cameron et al., 2010; Kelly et al., 2010). Limited data and large knowledge gaps complicate predictions about the ultimate effects of changes in sea ice on the behavior, health, abundance, and distribution of these seals. To date, indices of seal health sampled during periods of declining sea ice differ regionally (Crawford, Quakenbush & Citta, 2015; Harwood et al., 2020). Knowledge about evolutionary constraints on the timing of reproductive and molting behavior is generally lacking, so it is difficult to predict how or if ice-associated seal species adapt to future changes (e.g., by adjusting pupping or molting schedules to earlier dates or different locales). This is further complicated by the spatio-temporal variation in the phenology of these life history events within regions and throughout their full ranges. Additionally, trends in abundance of these species are unknown, so it is difficult to assess the effect, if any, declines in sea ice habitat have had, or will have, on seal demography.

Statutory requirements (e.g., United States Marine Mammal Protection Act (MMPA), United States Endangered Species Act (ESA)) for timely estimates of population abundance and trend of these species mean improved aerial survey effort is needed – and, must be paired with improved knowledge of haul-out behavior to ensure appropriate survey design, robust methods, and accurate estimates. Several studies have contributed estimates of the distribution and abundance of ice-associated seal species in the Arctic using aerial surveys (e.g., (Bengtson et al., 2005), (Conn et al., 2014), and (Ver Hoef et al., 2014)) and more recent efforts have significantly expanded on previous survey effort. Such abundance studies are conducted over very large areas and estimation of absolute abundance requires making inference about numerous issues affecting the observation of seals on ice. These include availability (only seals on ice are available to be counted), detection probability (observers or automated detection systems may miss some seals on ice), species misclassification, and possible disturbance of seals by aircraft (Ver Hoef et al., 2014; Conn et al., 2014). Refining these inferences will improve the accuracy of abundance estimates and, hopefully, allow credible predictions about the effects of climate disruptions on the abundance and distribution of Arctic seal populations.

How ice-associated seals use sea ice as a haul-out platform varies between species. Ribbon seals (*Histiophoca fasciata*) haul out of the water almost exclusively on sea ice and are mostly pelagic outside the spring pupping, breeding, and molting season (Boveng & Lowry, 2018). While spotted (*Phoca largha*) and bearded (*Erignathus barbatus*) seals rest on coastal features, they strongly prefer sea ice as a haul-out platform during the spring and early summer (Frost & Burns, 2018). Ringed seals (*Phoca hispida*) — not included in this study — haul out on sea-ice but also within snow lairs during winter and spring.

The remoteness of the Bering and Chukchi seas means direct scientific observation of seal behavior is impractical. Thus, bio-logging devices are especially useful tools for collecting key information on movement and haul-out behavior for these species. Bio-logging records of time spent out of the water provide valuable data for identifying covariates that explain variation in haul-out behavior. For instance, Von Duyke et al. (2020) used satellite-linked bio-loggers to corroborate seasonal changes

83 between diurnal and nocturnal haul-out behavior of ringed seals previously described by Kelly and
84 Quakenbush (1990) using VHF radio tags and direct observation. Bengtson et al. (2005) documented
85 a higher propensity for ringed seal basking near solar noon, as did Ver Hoef et al. (2014) in an analysis
86 of bearded, ribbon, and spotted seals using much larger sample sizes. Olnes et al. (2020) showed
87 that the proportion of time bearded seals spent hauled out progressively increased through spring
88 and summer, and Ver Hoef et al. (2014) found haul-out probabilities increased gradually starting in
89 March and peaked in May and June for bearded, ribbon, and spotted seals. Such analyses have not
90 been limited to the Arctic. In the Antarctic, Bengtson and Cameron (2004) relied on bio-logging data
91 to demonstrate greater haul-out propensity in juvenile crabeater seals (*Lobodon carcinophaga*) than
92 adults, with highest probabilities in February and at times close to solar noon.

93 Knowledge of haul-out patterns is not only important for understanding natural history and
94 ecology, but also for developing “availability” correction factors for aerial surveys. Specifically,
95 researchers need to know the fraction of seals hauled out (versus in the water) when aerial surveys are
96 conducted. Studies estimating availability correction factors for seals typically use logistic regression-
97 style analyses to estimate the time-specific probability of being hauled out based on ‘wet/dry’ data
98 relayed by bio-loggers. In these models, haul-out probabilities were expressed as a function of
99 predictive covariates, such as time-of-day, day-of-year, sex, age class, and environmental conditions
100 (e.g., (Reder et al., 2003), (Bengtson & Cameron, 2004), (Bengtson et al., 2005), (Udevitz et al.,
101 2009), (Ver Hoef et al., 2014), and (Southwell et al., 2008)). However, sample sizes have often been
102 insufficient to permit strong inference about demographic and/or seasonal variation in haul-out
103 probabilities. For instance, Bengtson and Cameron’s (2004) study included 5 adult and 2 juvenile
104 crabeater seals, while Bengtson et al.’s (2005) study was based on 6 ringed seals in the Chukchi and
105 Beaufort seas. These studies were often further limited by logistical constraints on fieldwork and
106 the attachment duration or operational life of bio-loggers. In this study, we addressed some of these
107 limitations by deploying small bio-loggers designed for longer-term attachment on rear flippers of a
108 subset of the study individuals. These devices are designed to collect data through the molt period
109 (when those adhered to the hair would fall off) and, in some situations, provide multiple years of
110 data.

111 In this study, we used 16 years of bio-logging data to investigate the haul-out behavior of bearded,
112 ribbon, and spotted seals in the Bering and Chukchi seas. Our goals were threefold. First, we wished
113 to establish baseline estimates for the chronology of haul-out behavior in the critical spring season
114 for each species across different age and sex classes. Second, we sought to refine estimates of haul-out
115 availability corrections for aerial surveys in order to improve estimates of seal abundance. Previously
116 estimated availability correction factors (e.g., (Bengtson et al., 2005), (Conn et al., 2014), and (Ver Hoef
117 et al., 2014)) accounted for variables such as the time-of-day and day-of-year, but did not investigate
118 meteorological variables that have been shown to influence haul-out behavior of walrus (Udevitz
119 et al., 2009). Third, we aimed to assess the annual variability in haul-out timing and possible linkage
120 to changes in the extent of seasonal sea ice between 2005 and 2020. Our work extends the scope of
121 previous haul-out analyses, includes the influence of meteorological variability, and investigates the
122 potential impact of changing sea-ice extent on the behavior of these species.

Areas of initial deployment location

Bio-loggers were initially deployed on **bearded**, **ribbon**, and **spotted** seals across multiple regions within the Bering and Chukchi seas.

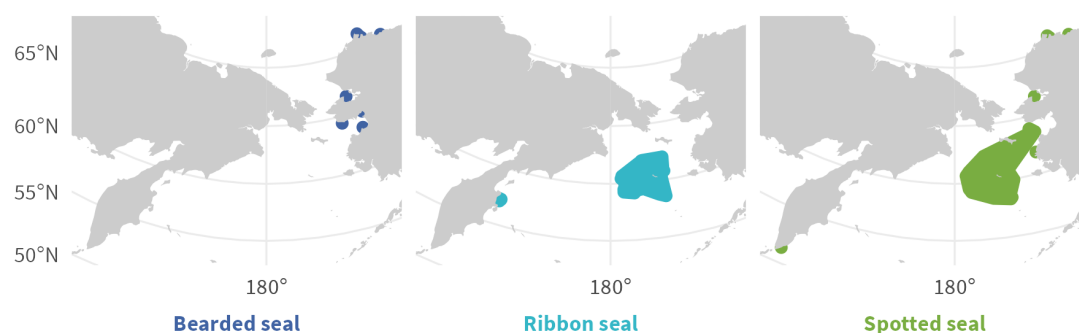


Figure 1. Initial deployment areas for each seal species (bearded, ribbon, and spotted seals) in the study between 2005 and 2020. Colored regions shown are minimum concave polygons buffered by 60 km for enhanced visibility. See Supplemental Material S1 for specific details. Deployments were initiated across a range of months but only data from 1 March to 15 July were included in the analysis. See Figure 3 for the spatial distribution of observed data used in the study.

METHODS

Data collection

For this study we used haul-out behavior data and location estimates from bio-loggers deployed on bearded, ribbon, and spotted seals in the Bering, Chukchi, and western Beaufort seas by multiple organizations as part of collaborative investigations from 2005 through 2020. Seals were captured using nets and bio-loggers were attached during studies based in coastal communities or on research ships. Ship-based capture events occurred during spring near the southern ice edge in the Bering Sea between 2005 and 2018. Land-based capture events occurred between 2005 and 2020 from May to October, generally between the coastal communities of Scammon Bay, Alaska in the Bering Sea, Utqiagvik, Alaska in the Chukchi Sea, and Nuiqsut in the Beaufort Sea. Data from additional deployments along the Kamchatka peninsula in the western Bering Sea are also included. We refer readers to Figure 1 and the primary literature for detailed capture and bio-logger attachment methods (see Supplemental Material, S1).

We subset haul-out behavior data from 249 bio-loggers deployed on 35 bearded, 110 ribbon, and 104 spotted seals to include only records from 1 March to 15 July between 2005 and 2020. Bio-loggers were of the ‘SPLASH’ or ‘SPOT’ family of tags developed by Wildlife Computers (Redmond, Washington, USA) and we either adhered them to the hair on the seal or attached them through the rear flipper inter-digital webbing. The use of bio-loggers adhered to the back or head provides some benefits over flipper mounted devices (e.g. increased satellite transmittal rates, locations at sea) but these are lost during the following annual molt, which, depending on deployment date, limits the duration of haul-out data they provide. Additionally, bio-loggers attached to the head or dorsal region are often dry while the seal is floating at the surface, inducing a slight positive bias in the hourly percent-dry values reported by the bio-logger. For this study, in cases where both bio-logger types were deployed, we preferred hourly percent-dry observations from the flipper tag.

Sex as well as age class (non-dependent *young-of-the-year*, sexually immature *subadults*, and mature *adults*) were estimated at the time of deployment by various combinations of length, claw growth ridges (McLaren, 1958), and pelage characteristics for some species. Seals determined to be less than one year were classified as young-of-the-year. For those bio-loggers deployed on young-of-the-year and transmitting into the next year, the age class was advanced to subadult on 1 March of the following year. Subadults are those seals likely greater than one year of age but less than four years. Adults are individuals that are likely older than four years. Table 1 provides a summary of these deployments and data received from them.

Table 1. Summary of bio-logger data across seal species and age classification from 1 March to 15 July 2005-2020. Total seal hours represents the sum of hourly observations across all seals used in the analysis. Because young-of-the-year are advanced to subadult on 1 March of the following year, some individual seals are represented in both columns in this table

Species	Sex	Age Class		
		Adult	Subadult	Young-of-the-Year
Bearded seal	F	1 (1,776 seal hours)	16 (21,648 seal hours)	
Bearded seal	M	2 (1,676 seal hours)	16 (17,232 seal hours)	
Ribbon seal	F	33 (35,128 seal hours)	18 (15,984 seal hours)	13 (3,734 seal hours)
Ribbon seal	M	24 (27,465 seal hours)	19 (13,046 seal hours)	9 (4,275 seal hours)
Spotted seal	F	23 (20,654 seal hours)	20 (17,542 seal hours)	11 (13,417 seal hours)
Spotted seal	M	20 (28,360 seal hours)	21 (17,210 seal hours)	12 (11,285 seal hours)

Tags that fall off due to molt, attachment failure, or seal mortality and remain on ice or land may continue to send data to satellites; i.e., a detached bio-logger that is dry (either on ice or land) will record and transmit data suggesting the seal is hauled out. Therefore, end times of each deployment were identified by examining bio-logger locations, percent-dry records, and dive behavior (if available) to determine when bio-loggers ceased providing data consistent with seal behavior. For example, a data record that ends with several consecutive days (~10+ days) of 100% dry observations and with locations indicating the tag was on land would be truncated to not include the final stretch of 100% dry observations. The vast majority of deployments end with the device detaching in the water and the deployment end date is obvious. There is no perfect algorithm for identifying deployment end dates and each deployment in question must be considered separately. While not perfect, we are confident our reliance on expert opinion and examination of multiple data streams provides the best option. Data outside of the deployment start and end times were discarded prior to analysis.

Haul-out behavior data were recorded and transmitted via the Argos satellite network as hourly percent-dry timelines. For each hour of a day, the wet/dry sensor was polled by the tag firmware every few seconds and the percent of the hour in the dry state was calculated (Figure 2). On board the bio-logger, hourly percent-dry calculations were rounded to the nearest 10% inclusive of 0% and 100% along with additional values at 3% and 98%. This compression resulted in additional data

Observation records are incomplete because not all data are transmitted

percent-dry actogram for a single ribbon seal across two years

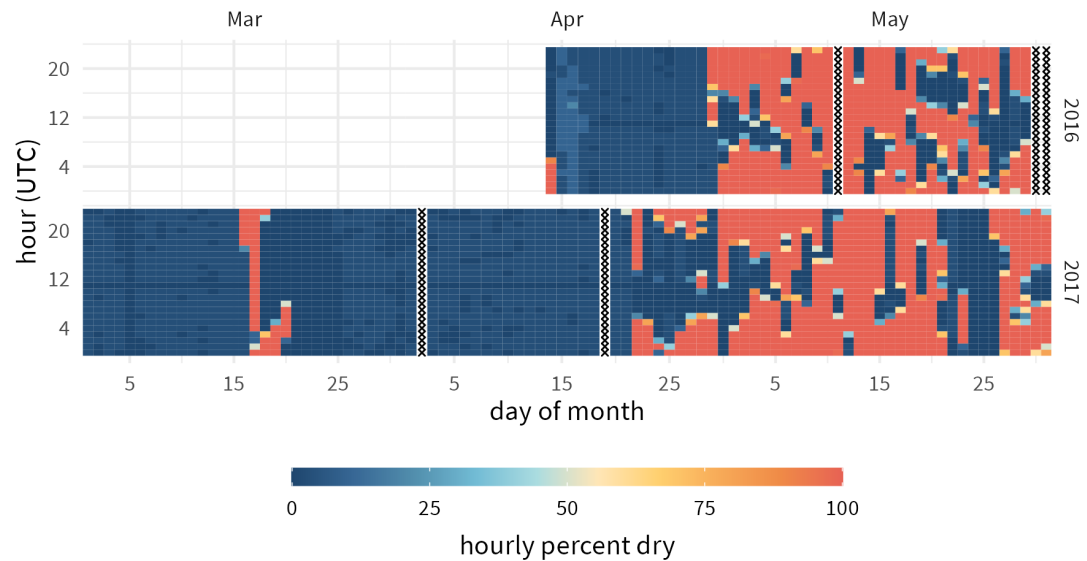


Figure 2. Haul-out behavior observations recorded by a bio-logger deployed on a ribbon seal over two years during the months of March, April, and May. Areas of the plot indicated with an 'X' represent missing data that were not successfully recieved from the tag.

transmission as each message consisted of two complete 24-hour records. Memory capacity allowed caching of percent-dry records for several weeks or months and each message was transmitted several times to ensure reception at the satellite. Bio-loggers were, typically, deployed and programmed in a manner to maximize data transmission during the spring pupping and molting period, though hourly percent-dry data were not always successfully transmitted. This is due to a variety of factors including satellite coverage, tag availability (e.g., tags mounted to the rear flipper often do not transmit while at sea), tag performance, duty cycling, and atmospheric interference. Fortunately, missing records do not substantially bias inference about haul-out probabilities (Conn et al., 2012).

Of key interest in this study was the relationship between haul-out behavior and weather covariates that vary with time and seal location. We explored the use of a continuous-time correlated random walk (Johnson et al., 2008) movement model to predict locations at specific times. However, the sparse nature of data from some bio-loggers, especially those mounted to the rear flipper, resulted in poor modeling performance or convergence issues. For this study, we calculated a weighted average daily location where the inverse of the estimated Argos or FastLoc GPS location error was used for the weight. Each Argos location estimate was assigned an error radius based on either the categorical location quality (3 = 250 m, 2 = 500 m, 1 = 1500 m, 0 = 2500 m (Lopez et al., 2013); we chose 2500 m for location classes A and B) or, when available, the estimated error radius from the Argos Kalman filter algorithm. Location estimates from FastLoc GPS were all assigned an error radius of 50 m. On days when haul-out observations were present but location data were missing we used the seal's last calculated weighted average daily location; days when the location intersected with land were removed from the seal's record. We recognize that bearded and spotted seals haul out on land. However, assessing the relationship between haul-out behavior and weather covariates and

194 seals' availability for aerial surveys on land was outside the scope of this study. Additionally, any
195 daily locations on land were likely more reflective of coordinate averaging and measurement error,
196 rather than actual use of coastal features.

197 Explanatory variables

198 In addition to sex and age class, we analyzed variables that might help explain variation in haul-out
199 probabilities. These included day-of-year (for seasonal effects) and local solar hour (for diurnal
200 effects). We calculated local solar hour using the {solaR} package (Perpiñán, 2012) within the R
201 statistical environment (R Core Team, 2021) based on the weighted daily average locations. We also
202 linked the weighted average daily locations to weather values from the North American Regional
203 Reanalysis (NARR) model produced by the National Centers for Environmental Prediction (Mesinger
204 et al., 2006). The NARR model assimilates observational data to produce a long-term picture of
205 weather over North America and portions of the surrounding seas. Weather variables are made
206 available across the region 8 times daily. For this study, NARR weather values were subset to the
207 extent of our study area over the Bering and Chukchi seas at 3-hr intervals based on the native grid
208 resolution of 32 km (1024 km²). The following meteorological variables are known to affect haul-out
209 behavior in other Arctic pinnipeds (Reder et al., 2003; Udevitz et al., 2009; Perry, Stenson & Buren,
210 2017) and were interpolated and assigned to daily seal locations using a bilinear method: 1) air
211 temperature at 2 m above the Earth's surface, 2) wind consisting of northerly and easterly vector
212 components converted to wind speed using the Euclidean norm, 3) barometric pressure at sea level,
213 and 4) precipitation (Table 2).

214 For all seal species, we considered the following variables when modeling the hourly haul-out
215 behavior: day-of-year, solar hour, temperature, wind speed, barometric pressure, precipitation, wind
216 chill (represented by a *wind:temperature* interaction (Udevitz et al., 2009)), and day-of-year and
217 time-of-day (solar hour) interactions to permit diurnal patterns to change from March to July. Ribbon
218 and spotted seal models included age-sex class and interactions between day-of-year and age-sex class,
219 but we omitted these from bearded seal models due to poor representation of age-sex classes (Table 1).
220 Bearded seal models included a latitudinal effect (and an interaction with day-of-year), since bearded
221 seals occupy a substantial range during the spring and we were interested in possible differences in
222 the timing of haul-out behavior along a latitudinal gradient. We omitted the latitudinal effect from
223 ribbon and spotted seal models because, during the spring, these species are most prevalent near the
224 southern ice edge in the Bering Sea (Conn et al., 2014).

225 We did assess whether the annual variation in maximum spring sea ice extent in the Bering Sea
226 influenced the seasonal peak of seal haul-out probability. In particular, we used sea ice concentration
227 data from the Nimbus-7 SMMR and DMSP SSM/I-SSMIS Passive Microwave Dataset, Version 1
228 (Cavalieri et al., 1996) to calculate maximum sea ice extent. All sea ice concentration grid cells (25
229 km²) in the study area with greater than 15% concentration were counted daily to get the total sea ice
230 extent for each day between 15 February and 15 July across all years. Maximum spring sea ice extent
231 was simply the largest daily count of grid cells with greater than 15% concentration for each year.

Table 2. Explanatory covariates used in analyses of binary haul-out records for bearded, spotted, and ribbon seals. Note that we also considered select interactions (see article text) between these primary covariates. For instance, wind chill was represented by the interaction temperature:wind.

Covariate	Type	Description
Age-sex class	Categorical	young-of-the-year, subadult, adult male and adult female
Hour	Continuous; Fourier basis	local solar hour using 6 variables of a Fourier-series basis
Day	Continuous	linear, quadratic, and cubic effects of day-of-year
Precip	Continuous	convective precipitation kg/m ²
Pressure	Continuous	atmospheric pressure at sea level (kPa)
Temp	Continuous	air temperature (C) at 2m above the earth's surface
Wind	Continuous	northerly and easterly vector components for wind converted into a single wind speed via the Euclidean norm
Northing	Continuous	latitude divided by the mean latitude across all locations (for bearded seals only)
Year	Continuous	For the set of models examining inter-annual variation in sea ice use, we fitted models with the addition of year by day-of-year interactions.

Haul-out modeling

Haul-out records for seals are often characterized by sequential hours spent basking on ice alternating with long periods in the water (Figure 2). Commonly used statistical models for binary data (e.g. logistic regression) assume independence among responses, an assumption that is clearly violated if hourly responses are modeled. Any analysis that ignores temporal autocorrelation in responses will thus have overstated precision (Betts et al., 2006).

To properly account for temporal dependence within a computationally tractable framework, we used generalized linear mixed pseudo-models (GLMPMs; (Ver Hoef, London & Boveng, 2010)) to model variation in haul-out behavior as a function of (1) covariate predictors, (2) temporally autocorrelated random effects, and (3) individual random effects representing heterogeneity in individual behavior. We used the glmmLDS package (Ver Hoef, London & Boveng, 2010) to implement GLMPMs. We explored two different model formulations for our data, and owing to the large number of records, we fit separate models to bearded, ribbon, and spotted seal data sets. In our first model formulation and for each species, we fitted a year-independent model that predicted average haul-out behavior as a function of demographic, environmental, seasonal, and diurnal effects.

Second, for ribbon and spotted seals (which had considerably more data than bearded seals), we fitted models that included all the effects from the first model, but also permitted annual variation in haul-out timing. This second class of models was used to examine whether haul-out patterns varied by year and to determine the annual timing of apparent peaks in haul-out behavior. For both models, we assumed an hourly Bernoulli response (i.e., whether tags were mostly dry or mostly wet) where the linear predictor was modeled on the logit scale. This is consistent with previous approaches London et al. (2012) and only -99% of our observations fell between 10% and 90% hourly percent-dry.

We followed Ver Hoef et al. (2014) in using linear, quadratic, and cubic effects of day-of-year to represent temporal changes in behavior. However, unlike previous models for harbor seals (London et al., 2012) and ice-associated seals (Ver Hoef et al., 2014), which treated hour-of-day as a 24-level categorical variable to capture diurnal cycles, we adopted a continuous formulation based on Fourier series that provides a flexible model while preserving the inherent circularity needed for time-of-day effects (i.e., hour 0 should be equal to hour 24). It also represents hour-of-day with 6 parameters, which is a considerable reduction when compared to a 24-parameter variable, especially when fitting models including interactions between hour-of-day and other variables (e.g., age-sex class, day-of-year). According to this approach, we used the following specification for hour-of-day effects:

$$H_t = \alpha_1 \cos\left(\frac{\pi t}{4}\right) + \alpha_2 \sin\left(\frac{\pi t}{4}\right) + \alpha_3 \cos\left(\frac{\pi t}{6}\right) + \alpha_4 \sin\left(\frac{\pi t}{6}\right) + \alpha_5 \cos\left(\frac{\pi t}{12}\right) + \alpha_6 \sin\left(\frac{\pi t}{12}\right)$$

where H_t gives the effect for solar hour t and α_i are estimated parameters (regression coefficients).

For the second set of models examining inter-annual variation in sea ice use, we fitted models with year by day-of-year interactions. However, in this case we only included *year:day* and *year:day*², omitting the main effects of year as well as *year:day*³ interactions because models with the latter effects were numerically unstable. However, the modeled interactions were sufficient to allow shifts in haul-out distribution, as one can show mathematically that a simple horizontal shift in timing of haul-out distributions does not affect the main effects or cubic terms in a polynomial regression model. Bearded seals were not included in this examination of inter-annual variation because of limited data across many years in the study.

A typical model fitting exercise would also include a model selection process. However, AIC (and similar criteria) is not suitable when using pseudo-likelihoods, because pseudo-data generated in the model fitting process (Ver Hoef, London & Boveng, 2010) differ between models (Ten Eyck & Cavanaugh, 2018). After fitting GLMPM models, we instead used “type III” *F*-tests to calculate *p*-values (Ver Hoef, London & Boveng, 2010) to evaluate model performance and important terms. We also produced predictions of haul-out behavior as a function of three influential predictors (e.g. solar hour, day-of-year, age-sex). Weather covariates for these predictions were based on daily or hourly smoothed weather covariate values across the study region. Such predictions were then used to develop haul-out probability surfaces, explore conditional effects of weather covariates, and determine annual peaks in haul-out activity. The timing of peak haul-out behavior was further used to regress against the annual maximum sea ice extent in the study region.

Visualizing the marginal or conditional effect of an individual weather covariate on haul-out probability was difficult in this analysis because of the collinearity between covariates as well as the spatial and temporal variation across such a large region. The relationship of each weather covariate with haul-out probability, averaged over the other weather conditions, was more variable than model coefficients would imply. That said, important insights can be gained from plots of marginal effects.

Spatial distribution of haul-out behavior data

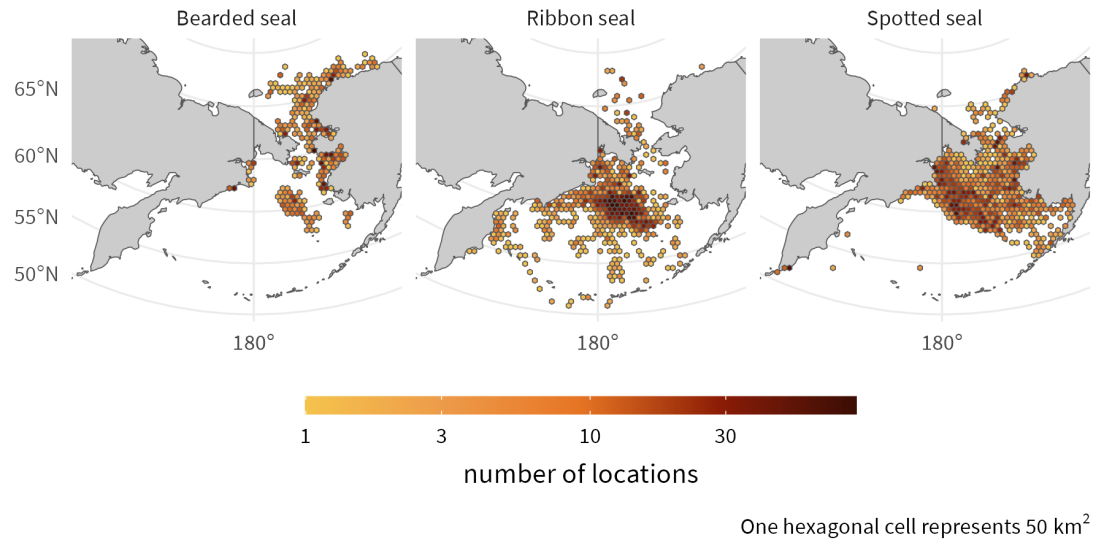


Figure 3. Spatial distribution of haul-out data during the months of March through July 15 for each of the three species. Data were collated across all years between 2005 and 2020. Each hexagonal cell represents an area of 50 km²

To create these plots, we predicted haul-out probability across the full range of weather covariate values while fixing hour of the day at local solar noon and day-of-year at the average day of peak haul-out for each species. The visualizations also include vertical lines representing 95% confidence intervals around the predicted haul-out probability to better communicate the variation in model uncertainty.

RESULTS

Figure 3 shows the spatial distribution of weighted locations with available haul-out behavior data used for analysis of each species across the study area.

Figure 4 shows the temporal distribution of all haul-out data across the study season for each species. Observations for ribbon and spotted seals were concentrated in the months of May and June due to the timing of deployment (April and May) and the timing of molt (May and June). During molt, seals (and their attached bio-loggers) spend more time out of the water and more data are transmitted. Molt timing also impacts when many deployments end as any bio-loggers adhered to the hair will fall off. Relative to the other species in the study, there were fewer deployments of bio-loggers on bearded seals. This resulted in fewer data observations overall and noticeably lower in numbers May and June. The majority of bearded seal deployments started later in the summer and, by May, bio-loggers had either fallen off or their batteries were depleted.

Models omitting year effects suggested that day-of-year, solar hour, age-sex class, temperature, and wind substantially influenced haul-out behavior of all three species, with *F* tests producing *p*-values less than 0.05 for variables embodying these effects and/or their interactions. Haul-out probabilities typically increased throughout March and April, reaching a peak in May and early June

Density of haul-out behavior observations varies from March through July
data are grouped by day-of-year and presented in seal-hours collated across years (2005-2020)

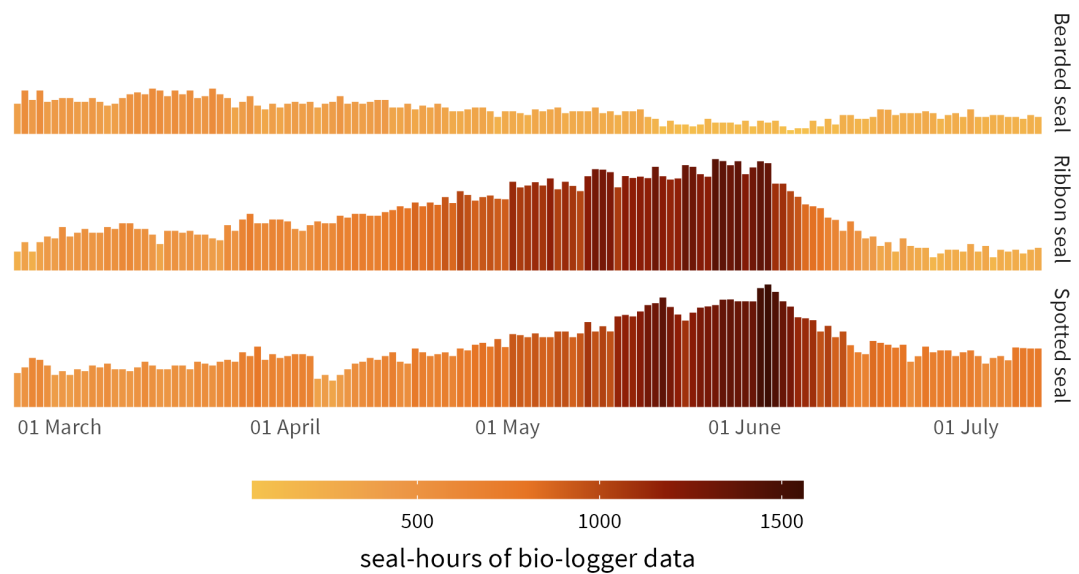


Figure 4. Distribution of hourly percent-dry bio-logger data from 1 March to 15 July for each species. Data are grouped by day-of-year and presented in seal-hours collated across all years between 2005 and 2020. The higher density of data from May and June in ribbon and spotted seals coincides with peak molting when seals (and their attached bio-loggers) are more likely hauled out. Additionally, many bio-logger deployments started in April and May. The overall reduced quantity of observations from bearded seals is reflective of the lower number of bio-logger deployments in the study.

Bearded seals rest on sea ice around solar noon and solar midnight

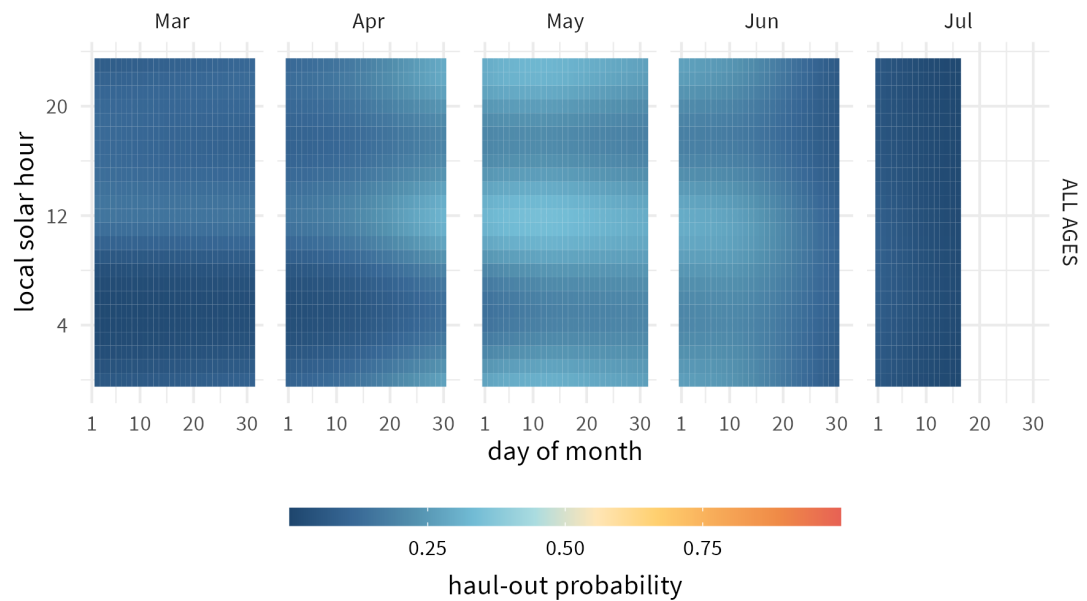


Figure 5. Predicted hourly haul-out probability of bearded seals (all ages and sex classes) from 1 March to 15 July for all and sex classes. Hour of the day was fixed at local solar noon.

before declining again. Diurnal patterns were present, with maximum haul-out behavior centered around local solar noon.

Bearded Seals

Age and sex class were not included in the model for bearded seals due to our lower sample size for adult and young-of-year age classes. As such, results are shown for all ages (Figure 5). Additionally, after approximately 9 months, 7 devices deployed on the rear flipper of bearded seals reported implausible hourly percent-dry data (100% dry for several weeks but indicative of movement and increasing transmission rates (see Boveng & Cameron (2013))). All data after the first instance of unrealistic values were censored from this analysis. Overall, bearded seals were less likely to haul out and had a bi-modal distribution of haul-out probability across the day. In addition to a peak around local solar noon, the bearded seal model predicted additional haul-out activity around local midnight. In concert with the lower magnitude of haul-out probability, bearded seal haul-out behavior was also more protracted throughout the spring season compared to ribbon and spotted seals.

When exploring the influence of weather, bearded seal haul-out probability appeared most influenced by wind ($F_{1,42302} = 185.426$; $p = <0.001$) and temperature ($F_{1,42302} = 14.915$; $p = <0.001$) (Figure 6). Additionally, wind chill was also a significant influence ($F_{1,42302} = 13.58$; $p = <0.001$). Any influence of barometric pressure or precipitation was less apparent.

Ribbon Seals

Ribbon seals exhibited a pattern of gradually increasing haul-out probability in April that peaks in late May for subadults and in early June for adults (Figure 7). The behavior was clearly centered around local solar noon and expanded to other hours later in the spring as seals entered their molting period. Subadults showed an earlier start and more intense haul-out activity in April and May. The

Bearded seal haul-out probability is influenced by temperature, wind, and pressure. Marginal effects are calculated for local solar noon on the day of peak haul out

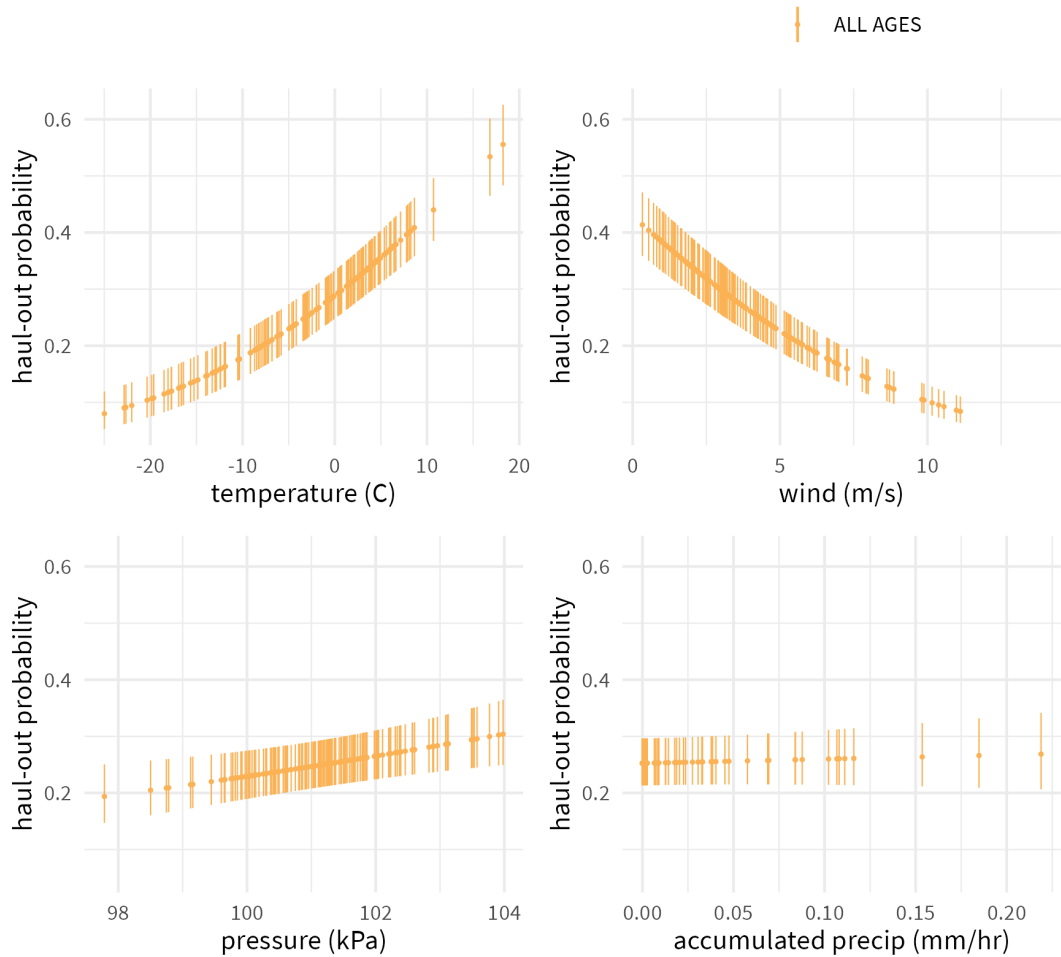


Figure 6. Marginal effects of temperature, wind, pressure, and precipitation on the predicted haul-out probability of bearded seals combined across all age and sex classifications. Hour of the day was fixed at local solar noon and day-of-year held at the day of peak haul-out for bearded seals. Vertical lines represent the 95% confidence interval around the predicted haul-out probability.

Ribbon seals rest on sea ice in late spring and around solar noon.

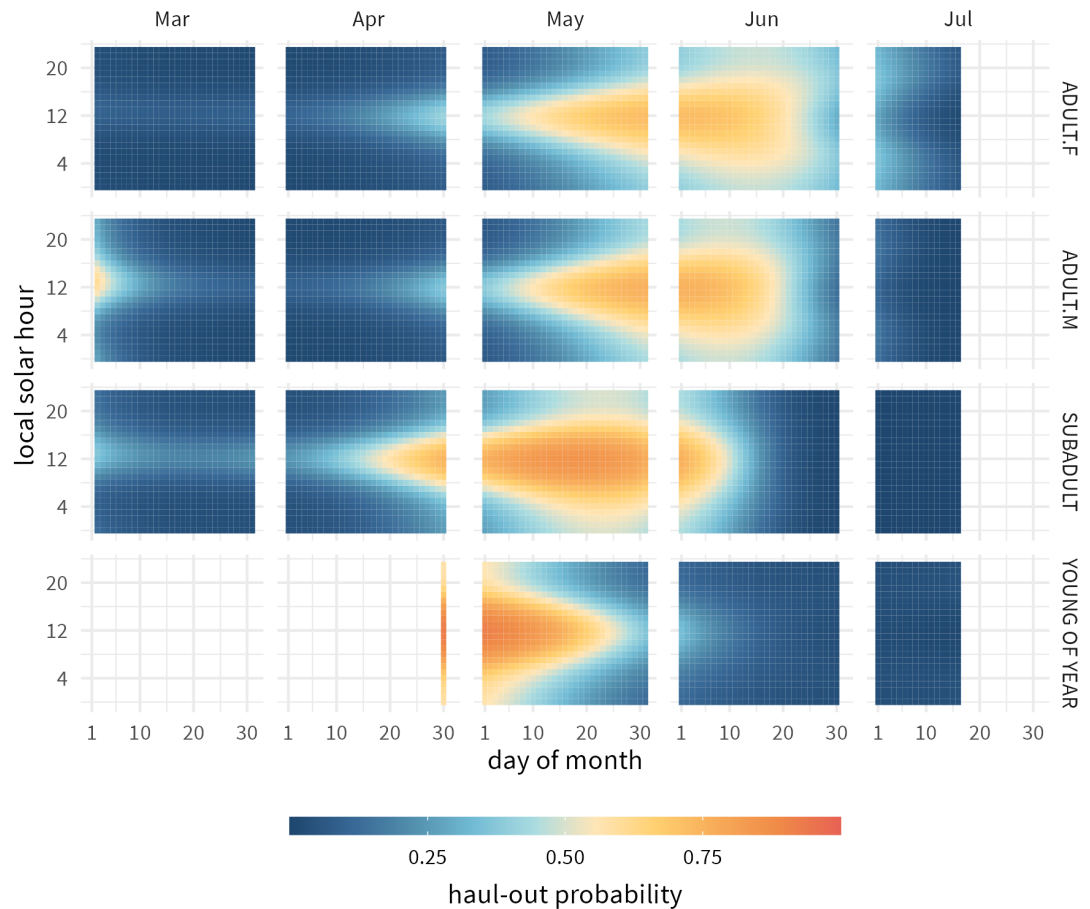


Figure 7. Predicted hourly haul-out probability of ribbon seals from 1 March to 15 July for each age and sex class used in the model. Adult females are indicated by 'ADULT.F' and 'ADULT.M' indicates adult males. 'SUBADULT' and 'YOUNG OF YEAR' age classes include both sexes. Hour of the day was fixed at local solar noon.

young-of-the-year records begin after weaning and the model predictions demonstrated the ontogeny of in-water activities (e.g. diving, foraging) in May. Adult females had a more protracted haul-out season compared to males, and more time was spent resting in June and July. The model suggested adult male ribbon seals completed their molt by the end of June. Lastly, for adults of both sexes there was some indication for a shift to a crepuscular haul-out pattern in late June and July.

The haul-out probability for ribbon seals was strongly influenced by each of the weather covariates included within the model except precipitation. The impacts of temperature ($F_{1,99593} = 15.688$; $p < 0.001$), wind ($F_{1,99593} = 72.392$; $p < 0.001$), and barometric pressure ($F_{1,99593} = 9.521$; $p = 0.002$) were especially noticeable as ribbon seals were less likely to haul out at higher winds and lower pressure values and more likely to haul out when temperatures were relatively warm. Wind chill ($F_{1,99593} = 2.794$; $p = 0.095$) showed a moderately negative influence on haul-out probability. As with bearded seals, Figure 8 presents the predicted haul-out probability of ribbon seals across the range of weather conditions encountered in the observed data. Because our ribbon seal model included age

Ribbon seal haul-out probability is influenced by temperature, wind, and pressure; marginal effects are calculated for local solar noon on the day of peak haul out

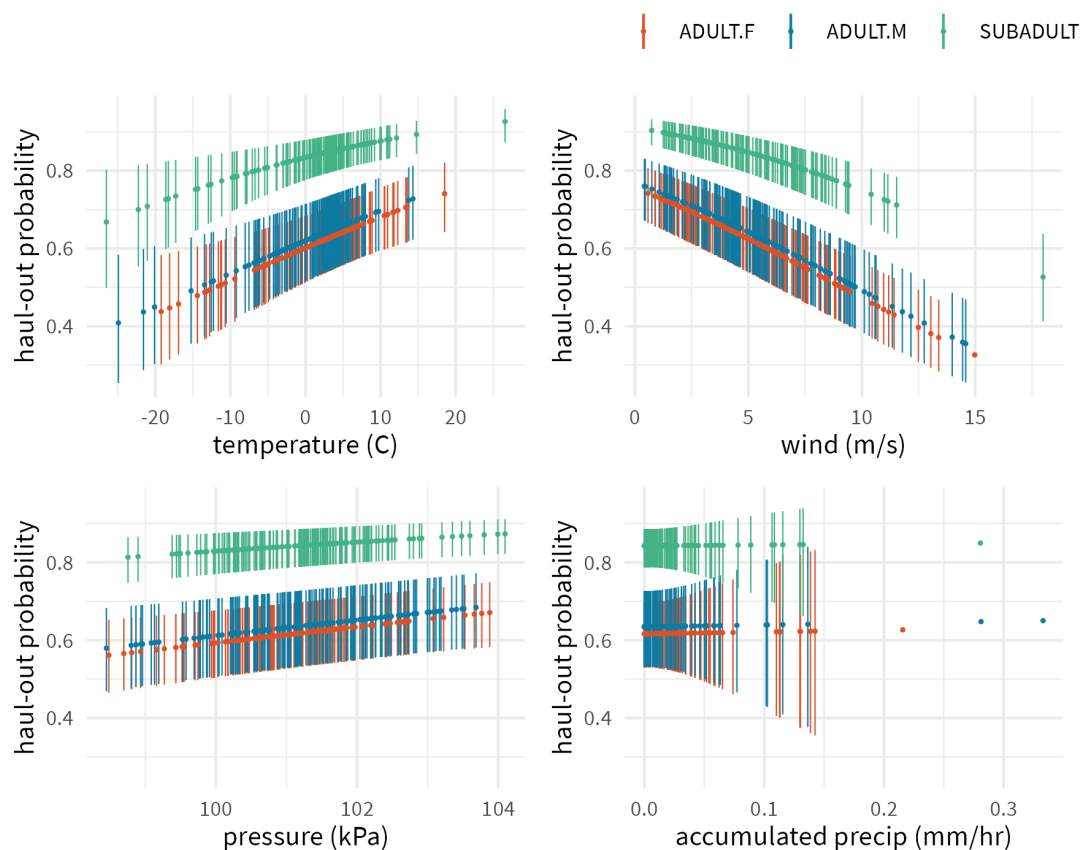


Figure 8. Marginal effects of temperature, wind, pressure, and precipitation on the predicted haul-out probability of ribbon seals within each age and sex classification. Hour of the day was fixed at local solar noon and day-of-year held at the average day of peak haul-out for ribbon seals. Vertical lines represent the 95% confidence interval around the predicted haul-out probability.

and sex class, we can visualize the different influences of weather covariates on those classes.

Spotted Seals

Compared to ribbon seals, spotted seals showed a longer spring haul-out season that was less intensely centered on solar noon (Figure 9). Adults of both sexes spent considerable time in April, May and June hauled out. In contrast to ribbon seals, adult spotted seal males had a more protracted haul-out season compared to females, and more time out of the water in June and July (Figures 7 and 9). As with ribbon seals, the young-of-the-year records began after weaning and the model predictions demonstrated the ontogeny of in-water activities (e.g. diving, foraging) in May.

Spotted seal haul-out behavior was most strongly influenced by temperature ($F_{1,108429} = 3.432$; $p = 0.064$) and wind ($F_{1,108429} = 25.821$; $p < 0.001$) with barometric pressure having a moderate influence ($F_{1,108429} = 5.05$; $p = 0.025$). Spotted seals were less likely to haul out at higher winds and more likely to be on the ice when temperatures were relatively warm. Wind chill (*temperature:wind*) and precipitation were not as influential as the other covariates. Differences in the magnitude

Spotted seals rest on sea ice in late spring and around solar noon.

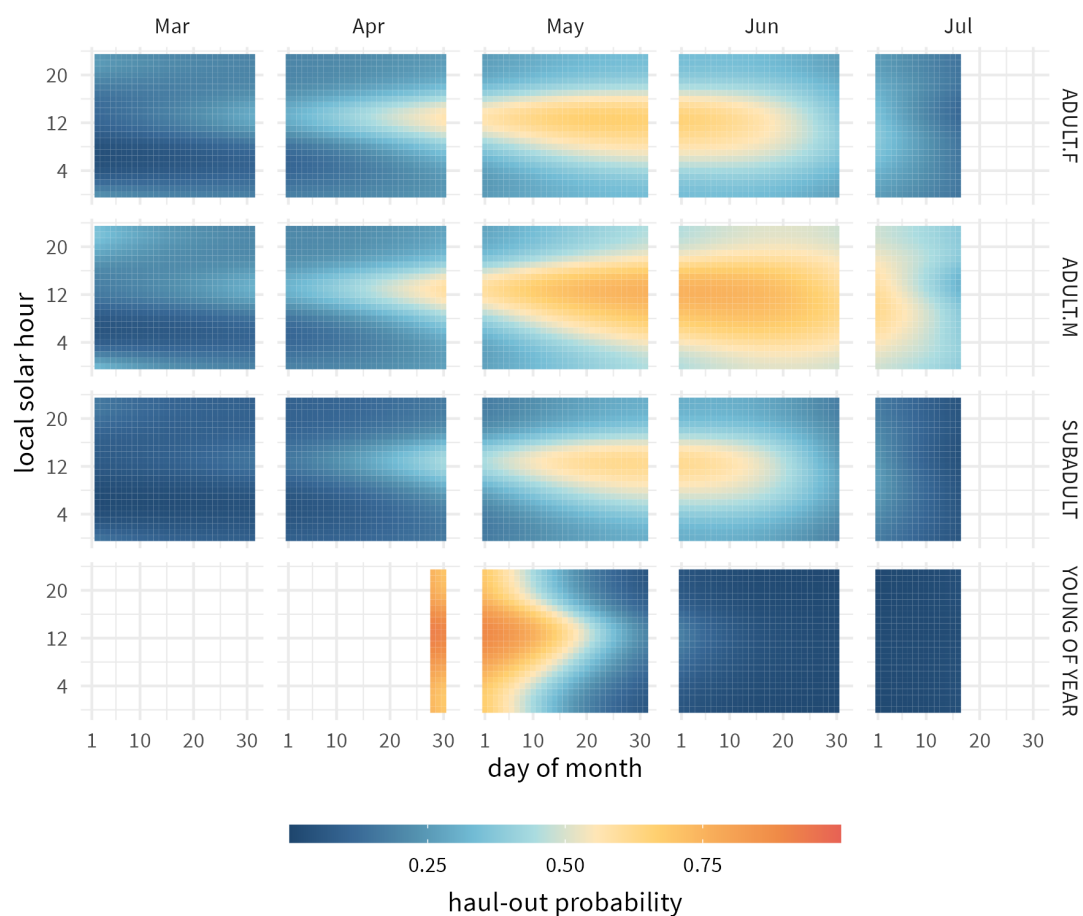


Figure 9. Predicted hourly haul-out probability of spotted seals from 1 March to 15 July for each age and sex class used in the model. Adult females are indicated by 'ADULT.F' and 'ADULT.M' indicates adult males. 'SUBADULT' and 'YOUNG OF YEAR' age classes include both sexes. Hour of the day was fixed at local solar noon.

Spotted seal haul-out probability is influenced by temperature, wind, and pre
marginal effects are calculated for local solar noon on the day of peak haul out

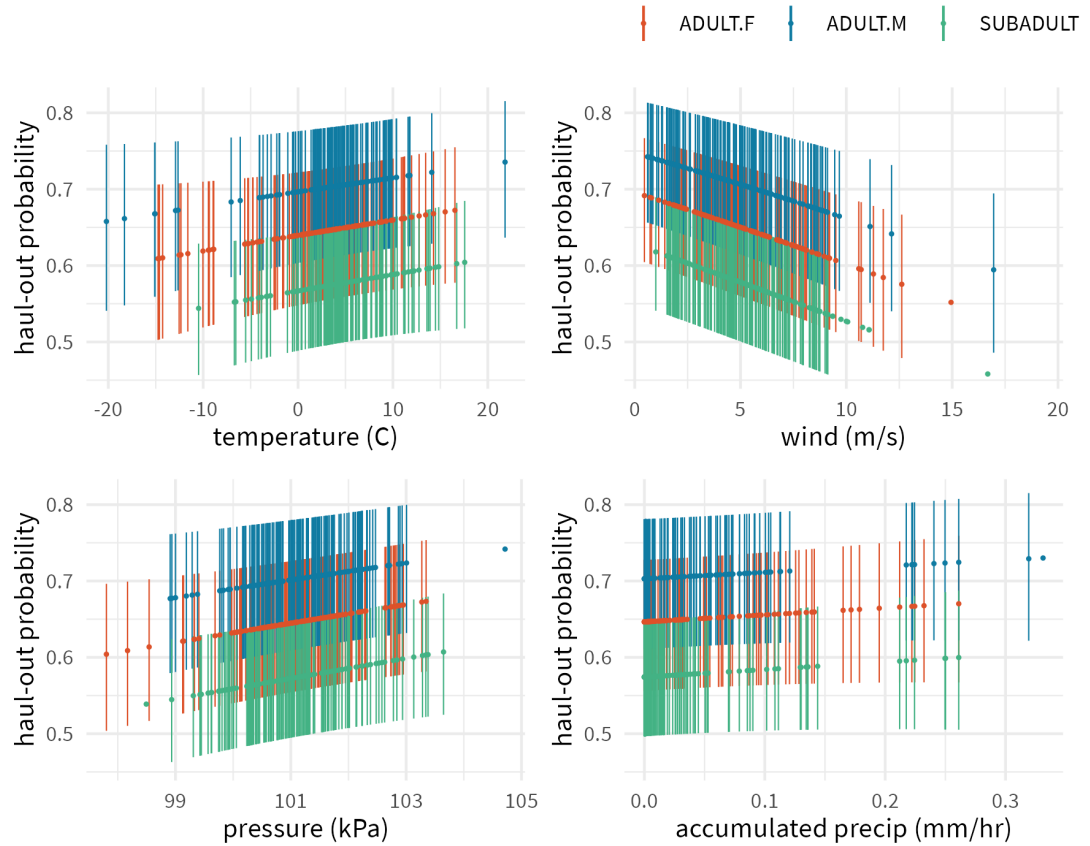


Figure 10. Marginal effects of temperature, wind, pressure, and precipitation on the predicted haul-out probability of spotted seals within each age and sex classification. Hour of the day was fixed at local solar noon and day-of-year held at the average day of peak haul-out for spotted seals. Vertical lines represent the 95% confidence interval around the predicted haul-out probability.

357 of response between the age-sex classes were present and consistent across each of the weather
358 covariates (Figure 10). There was a consistent ranking of adult males being the most likely to haul
359 out, followed by adult females, and, then, subadults.

360 Annual variation in haul-out timing

361 The second set of models, which included annual variation in haul-out patterns, uncovered significant
362 contributions for linear and quadratic interactions between day and year in both ribbon seals (day:year,
363 $F_{10,99563} = 4.051$; $p = <0.001$; day²:year, $F_{10,99563} = 4.582$; $p = <0.001$) and spotted seals (day:year,
364 $F_{14,108387} = 4.732$; $p = <0.001$; day²:year, $F_{14,108387} = 4.848$; $p = <0.001$). Predicted distributions of
365 haul-out activity were largely unimodal, but varied some among and within years with respect to the
366 timing and magnitude of haul-out peaks (Figure 11). It is important to note that predicted variation
367 in annual haul-out patterns likely reflected both process error and sampling variability. While we did
368 remove any years where only one deployment in a species + age:sex group was present, there were
369 still some years where the pattern shown was informed by a small number of individuals that may

Annual variability in the timing of peak haul-out probability

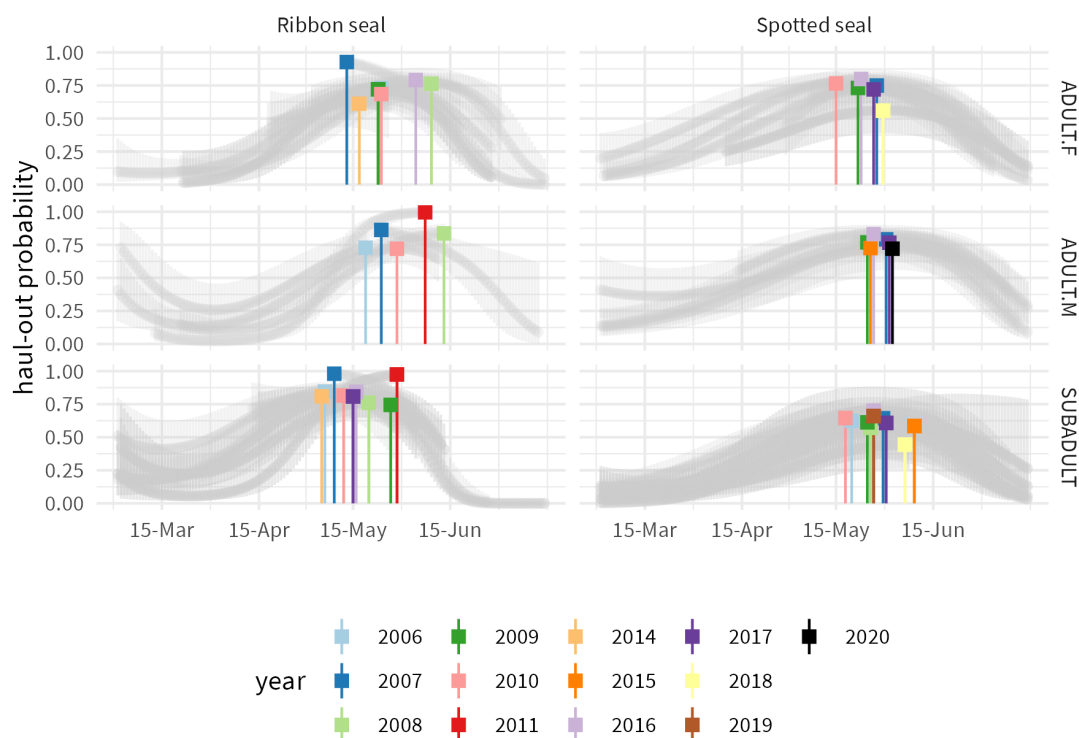


Figure 11. Annual variability in the timing of peak haul-out probability (colored markers) for ribbon and spotted seals across 13 years. Predictions are shown for local solar noon and under smoothed weather conditions. Only those groups (age:sex + year) that included observations from more than one seal are shown. Additionally, any groups where data were only available after 1 June or before 1 May are not included.

not represent population-level patterns.

The timing of peak haul-out probability for ribbon and spotted seals appeared to have little to no relationship with the amount of yearly maximum sea ice extent. Adult female and subadult spotted seals show a negative relationship, but there was no indication that the observed trend is meaningful (spotted seal adult female: $R^2 = 0.767$, $p = 0.022$; spotted seal subadults: $R^2 = 0.449$, $p = 0.034$). For ribbon seals and adult male spotted seals, p -values were substantially larger than 0.05 (ribbon seal adult females: $R^2 = 0.035$, $p = 0.689$; ribbon seal adult males: $R^2 = 0.116$, $p = 0.575$; ribbon seal subadults: $R^2 = 0.013$, $p = 0.773$; spotted seals adult males: $R^2 = 0.107$, $p = 0.592$).

DISCUSSION

We modeled data from bio-loggers deployed on bearded, ribbon, and spotted seals to examine factors affecting haul-out behavior on sea ice in the Bering and Chukchi seas. Our analysis shows all three species of seal haul out progressively more through the spring and peak near mid-May to early June before declining again. This pattern aligns well with what has been previously documented qualitatively (Boveng et al., 2009; Cameron et al., 2010; Boveng & Cameron, 2013) and confirms our haul-out data are likely quantifying population-level behavioral patterns. Seals preferentially

haul out on ice shortly after solar noon, which allows seals to maximize absorption of solar radiation thought to facilitate the molting process (Feltz & Fay, 1966). Interestingly, bearded seals appear to have two peaks in haul-out activity within a day, one shortly after solar noon, and one centered near solar midnight. This, of course, could be an artifact of our limited sample size for bearded seal deployments across all age classes. However, a similar bi-modal pattern has been seen in ringed seals (Von Duyke et al., 2020) and suggests that bearded and ringed seals may be operating under different constraints than ribbon and spotted seals. Bearded and ringed seals are distributed across higher latitudes and the extended daylight hours may allow more flexibility in optimizing resting periods with foraging. Other factors such as predation by polar bears (which is rare for ribbon and spotted seals in the Bering Sea) may also explain differing haul-out patterns. The change in haul-out behavior during the season was less pronounced in bearded seals compared to ribbon and spotted seals. This aligns with findings from Thometz et al. (Thometz et al., 2021) who observed a mean molting period of 119 ± 2 days and a relatively stable resting metabolic rate during that time. While ribbon seals were not considered in that same study, spotted and ringed seals underwent molt periods of just 33 ± 4 and 28 ± 6 days and had increased resting metabolic rates.

Unlike previous analyses of seal haul-out behavior in spotted and ribbon seals (e.g. (Ver Hoef, London & Boveng, 2010), (Conn et al., 2014)), we also investigated the influence of sex-age class on haul-out probabilities for both species (not for bearded seals because of low sample size). Field identification of age class can be inexact, particularly when differentiating subadults from adults. In the case of ribbon seals, subadults often have less distinct ribbons than adults. Spotted seal pelage cannot be used to reliably discern adults from subadults. Despite these challenges, we feel the age classifications used in this analysis are useful in testing for age-related effects on haul-out behavior.

Although ribbon and spotted seals exhibited a unimodal diel haul-out pattern generally centered around local solar noon, there were key differences across species, age, and sex that match our understanding from natural history descriptions of their ecological behavior. Spotted seals are known to form triads during the breeding season where a female and dependent pup are accompanied on the ice by a suitor male (Frost & Burns, 2018). The male waits for the female to wean the pup and enter estrus, and fends off any other potential suitor males. Triad formation results in both males and females spending a large portion of the day hauled out on ice and a protracted spring haul-out season for both sexes. While females are still nursing the pup and not yet in estrus they may be less inclined to interrupt their haul out and enter the water where the suitor male would attempt mating. We see this reflected in the predicted haul-out patterns, with both males and females exhibiting a broad distribution of time out of the water throughout the solar day and the season. Ribbon seals are not known to form triads and our model predicts a progression of increased haul-out behavior with females starting earlier in the season than males. Notably, female ribbon seals spend a large portion of the day in the water during the pupping period, aligning with the hypothesis that ribbon seal females continue foraging while nursing. Subadult ribbon and spotted seals begin elevated haul-out behavior earlier in the spring and follow a pattern seen in other phocids where yearlings and subadults molt first followed by adult females and males Kirkman et al. (2003).

We also investigated the influence of weather on haul-out probabilities, including wind speed, temperature, barometric pressure, precipitation, and wind chill. These have been investigated for walruses (e.g. Udevitz et al. (2009)) and a few select studies of ice-associated seals Perry, Stenson & Buren (2017). In our study, ribbon seals seemed to be the most influenced by weather, with wind, temperature, and barometric pressure all being important components of the model. Spotted seals

were most affected by wind and barometric pressure. For bearded seals, the model indicated wind and temperature had the greatest impact. In general, and as might be expected, seals were more likely to haul-out when daily temperatures were warmer, winds speeds were lower, barometric pressure was higher, and precipitation was lower. Those weather conditions are general indicators of increased solar radiation and lower convective heat loss, both of which provides energetic benefits. Low winds and precipitation could also enhance predator detection. Our results highlight the importance of incorporating weather covariates when analyzing haul-out behavior and calculating availability corrections for aerial surveys.

Notably missing from our list of explanatory variables is any spatial-temporal representation of sea ice concentration, area, or extent. This may seem counterintuitive when modeling the haul-out behavior of seal species with such a close association to sea ice; seals haul out in the presence of sea ice, and we could assess the local concentration of sea ice during these events (see (Olnes et al., 2020)). This, however, expands the scope of our analysis into the realm of habitat selection and many of our deployments consisted of a single device attached to the rear flipper of the seal which did not provide at-sea locations, limiting our ability to fully evaluate fine-scale habitat preferences related to sea-ice. Insight into how seals use and interact with sea-ice during an extended period when the availability and characteristics of sea-ice is rapidly changing is important but ancillary to the focus of this analysis. The focus of this study was to develop models applicable for aerial survey correction factors and using sea ice as a covariate would almost certainly bias haul-out predictions towards those seals that are on or near ice and therefore more likely to haul out. Since aerial surveys can only detect seals on ice, abundance estimates would be missing a correction for those seals that are away from ice. Consider the case of employing haul-out probabilities as an aerial survey correction in a population inhabiting two areas: an ice-free region and a region with sea ice. Further, assume that half of the population is in each area on average, and that the probability of hauling out is 1.0 in the ice covered region and 0.0 in the region without ice. Denoting the total population as N , let us examine what happens when we use (1) population level availability, and (2) ice-specific availability as a correction factor. We'll further assume, for sake of this example, that detection probability of seals on ice is 1.0, and that the entirety of the ice covered region is censused. In case (1), our aerial survey count is $C = 0.5 \cdot N$, with a population-level availability probability of $\hat{a} = 0.5$. A Horvitz-Thompson-type estimator for abundance is simply $\hat{N} = C \div \hat{a}$, which has expectation $\mathbb{E}(\hat{a}) = \frac{0.5 \cdot N}{0.5} = N$, as desired. Now consider case (2). In this case, the probability of hauling out is 1.0 for seals in the ice-covered area, so we have $\mathbb{E}(\hat{N}) = \frac{0.5 \cdot N}{1.0} = 0.5N$. That is, we bias abundance because we are not accounting for seals that are away from ice, and therefore have a zero chance of hauling out. The same logic holds for any other covariate, such as distance from the ice edge, that has different values in the surveyed and unsurveyed habitat of the seals (assuming distance from ice edge is coded with opposite signs inside and outside the pack ice). Lastly, our study was limited to the spring season when seal haul-out behaviors are strongly influenced by pupping, nursing, breeding behavior, and molt and these drivers are likely more influential than specific sea-ice concentration. Crawford et al. (2019) compared haul-out probability models for ringed seals and found those that only included season (and not sea-ice concentration) were the most parsimonious. For these reasons, we have elected not to use sea ice concentration *as a predictor for haul-out probability* in the present study.

Our models detected small deviations in the timing and magnitude of annual peaks in haul-out behavior for ribbon and spotted seals. The timing of peak haul-out activity appears to fall within a relatively narrow time window of 3-4 weeks in late May and early June. This consistency across

15 years is likely a reflection of the relationship between a critical photoperiod and the timing of important life history stages (Temte, 1994; Bronson, 2009). However, along with a critical photoperiod, ribbon and spotted seals are dependent upon the presence of sea ice for pupping and molt. We did not find any support in our models for a relationship between the timing of peaks in haul-out behavior and the amount of yearly maximum sea ice. This could indicate that, while the extent of spring sea ice in the Bering Sea varied widely during our study period, seals were still able to locate sea ice and haul out. Only a small portion of our data was from 2018 – 2019, years of extreme low spring ice extent in the Bering Sea that appeared to impact body condition of ribbon and spotted seals (Boveng et al., 2020), so we may currently lack sufficient contrast in ice extent to characterize an effect on haul-out probability. We should expect, however, that some minimal threshold in the spatial extent or density of sea ice will have a meaningful impact on the timing of peak haul-out behavior — if there is no sea ice, seals will not haul out or be forced to use terrestrial haul-out sites which were not part of the evolution of their normal behaviors. Additionally, while from an ecological perspective the haul-out behavior appears consistent, the interannual differences in timing and magnitude are large enough to have important ramifications on calculations of abundance and trend. Those ramifications will only be exacerbated if climate variability amplifies interannual differences.

Previous attempts to estimate the abundance of phocid seals from aerial survey data in the Bering and Chukchi seas (e.g. (Bengtson et al., 2005), (Conn et al., 2014), (Ver Hoef et al., 2014)) have used estimated haul-out probabilities to correct for the proportion of animals that are in the water and thus unavailable to be counted. Although several of these studies allowed haul-out probabilities to vary by day-of-year and time-of-day, they have not accounted for variability among years, in weather conditions, and in the age-sex class of the sample. In this paper, we have shown that there can be considerable differences in the number of seals hauled out on ice based on these factors. We recommend that future abundance analyses employ availability models that account for them. For instance, it is relatively straightforward to obtain weather reanalysis products for times and locations that are surveyed and to construct a relevant correction factor based on predictions of GLMPMs. The most challenging element in developing availability correction factors is with annual variability. It can be difficult to get a sufficient sample size to estimate year-specific correction factors, particularly because research teams would likely need to tag seals and conduct aerial surveys concurrently, requiring considerably more personnel and money. One possible suggestion is to estimate a “shift” parameter within models for aerial survey counts that allow the peak of haul-out distributions to be adjusted earlier or later in the year based on the frequency of counts observed over time. Regardless, researchers should anticipate there being some unmodeled heterogeneity in availability when using aerial surveys to estimate Arctic seal abundance. This may require consideration in trend estimation, as one will not know if moderate differences in abundance estimates are attributable to changes in abundance or changes in haul-out behavior.

Predictions of absolute haul-out probability in this paper were somewhat different than those previously reported for these species, especially for bearded seals. For instance, Ver Hoef et al. (2014) and Conn et al. (2014) used haul-out correction factors with maximums of 0.66 for bearded seals, 0.62 for ribbon seals, and 0.54 for spotted seals, where maximums corresponded to times near solar noon in mid-late May. Applying models that ignore age, sex, and year effects, these probabilities were 0.38, 0.72, and 0.60, respectively, under the current analysis framework. Our current estimates of haul-out probability reflect increased sample sizes in terms of number of individuals, but also improvements to the way data were prepared prior to analysis.

517 We focused this paper on haul-out behavior of bearded, ribbon, and spotted seals. Ringed seals
518 are also present in the Bering and Chukchi seas but exhibit a unique complicating factor. Adult
519 ringed seals build subnivean lairs under the snow on top of the sea ice, where they haul out and
520 where females rear pups until conditions are good for basking (Frost et al., 2004). Thus, the wet-dry
521 sensor on a bio-logger could indicate that an animal is hauled out, but if it is within a lair, it is not
522 available to be detected during an aerial survey. We hope to address availability of ringed seals using
523 data from satellite tags, replicate survey tracks, and auxiliary information about snow depth and
524 timing of melt in a future study.

525 AUTHOR CONTRIBUTIONS

- 526 • **Josh M. London:** investigation, conceptualization, methodology, formal analysis, validation,
527 software, writing: original draft, writing: review and editing, visualization, and data curation
- 528 • **Paul B. Conn:** conceptualization, methodology, formal analysis, software, validation, writing:
529 original draft, writing: review and editing
- 530 • **Stacie K. Hardy:** investigation, data curation, methodology, validation, writing: review and
531 editing
- 532 • **Erin L. Richmond:** data curation, investigation, methodology, validation, writing: review and
533 editing
- 534 • **Jay M. Ver Hoef:** conceptualization, methodology, software, writing: review and editing
- 535 • **Michael F. Cameron:** investigation, project administration, writing: review and editing
- 536 • **Justin Crawford:** investigation, methodology, validation, data curation, writing: review and
537 editing
- 538 • **Lori T. Quakenbush:** investigation, methodology, supervision, project administration, writing:
539 review and editing
- 540 • **Andrew L. Von Duyke:** investigation, methodology, validation, data curation, writing: review
541 and editing
- 542 • **Peter L. Boveng:** investigation, conceptualization, supervision, project administration, writing:
543 review and editing

544 DATA AVAILABILITY

545 This manuscript was developed as a reproducible research compendium. All data and code are avail-
546 able on GitHub (<https://github.com/jmlondon/berchukHaulout>) and major versions
547 archived at Zenodo (<https://doi.org/10.5281/zenodo.4638221>). Original data sources
548 for telemetry are archived at the United States Animal Telemetry Network, archived at Movebank, or
549 associated with other published manuscripts (see supplemental material S1).

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722 **SUPPLEMENTAL MATERIAL**

723 **1 S1. SUPPLEMENTAL TABLE OF BIO-LOGGER DEPLOYMENT DETAILS**

724 The following table provides additional details about the timing, location, and agencies responsible for the bio-logger deployments used in this study.

Data Contributions From Collaborators and Additional Deployment Details

	Year Deployed	Location	Publication(s)	Permits	No. Seals	Sex	Age Class
Alaska Dept. Fish & Game							
Bearded seal	2005	Kotzebue Sound	Cameron et al 2018	358-1585	6	M, F	subadult
Bearded seal	2006	Kotzebue Sound	Cameron et al 2018	358-1585	2	F, M	subadult
Bearded seal	2009	Kotzebue Sound	Breed et al 2018	358-1787	4	F, M	subadult
Bearded seal	2014	Norton Sound, Koyuk River	Olnes et al 2020	15324	2	M	subadult
Bearded seal	2014	Norton Sound, Nome	Olnes et al 2020	15324	1	M	subadult
Bearded seal	2015	Norton Sound, St. Michael	Olnes et al 2020	15324	1	M	subadult
Bearded seal	2016	Elson Lagoon, Utqiagvik	Olnes et al 2020	15324	1	F	subadult
Bearded seal	2016	Norton Sound, Koyuk River	Olnes et al 2020	15324	2	F, M	subadult
Bearded seal	2016	Norton Sound, Nome	Olnes et al 2020	15324	1	M	subadult
Bearded seal	2016	Norton Sound, St. Michael	Olnes et al 2020	15324	2	M, F	subadult
Bearded seal	2017	Colville River, Nuiqsut	Olnes et al 2020	15324	1	F	subadult

Bearded seal	2017	Norton Sound, Koyuk River	Olness et al 2020	15324	1	F	subadult
Bearded seal	2017	Norton Sound, Nome	Olness et al 2020	15324	1	F	subadult
Bearded seal	2019	Dease Inlet, Utqiagvik	Olness et al 2021	20466	1	M	adult
Spotted seal	2005	Kotzebue Sound	Von Duyke et al in prep	358-1585	3	F, M	subadult, adult
Spotted seal	2016	Dease Inlet, Utqiagvik	Von Duyke et al in prep	15324	4	M, F	adult
Spotted seal	2017	Colville River, Nuiqsut	Von Duyke et al in prep	15324	1	F	subadult
Spotted seal	2017	Scammon Bay	Von Duyke et al in prep	15324	3	F, M	subadult, adult
Spotted seal	2018	Dease Inlet, Utqiagvik	Von Duyke et al in prep	20466	1	F	subadult
Spotted seal	2018	Scammon Bay	Von Duyke et al in prep	20466	1	M	subadult
Spotted seal	2019	Dease Inlet, Utqiagvik	Von Duyke et al in prep	20466	6	M	adult, subadult

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Bearded seal	2005	Kotzebue Sound	NA	358-1585	1	F	subadult
Bearded seal	2009	Kotzebue Sound	McClintock et al 2017	782-1765	2	M	subadult, adult
Bearded seal	2011	Kotzebue Sound	McClintock et al 2017	15126	3	F, M	subadult
Bearded seal	2012	Kotzebue Sound	McClintock et al 2017	15126	1	F	adult
Ribbon seal	2005	Ozemoy Gulf, Russia	NA	782-1765	9	F, M	young of year, adult, subadult
Ribbon seal	2006	Bering Sea	NA	782-1765	7	M, F	adult, young of year
Ribbon seal	2007	Bering Sea	NA	782-1765	28	M, F	subadult, adult, young of year

Ribbon seal	2008	Bering Sea	NA	782-1765	1	M	subadult
Ribbon seal	2009	Bering Sea	NA	782-1765	28	F, Madult, subadult, young of year	
Ribbon seal	2010	Bering Sea	NA	358-1787, 15126	17	M, Fyoung of year, adult, subadult	
Ribbon seal	2014	Bering Sea	NA	15126	13	M, Fsubadult, adult, young of year	
Ribbon seal	2016	Bering Sea	NA	19309	7	M, F	subadult, adult
Spotted seal	2006	Bering Sea	NA	782-1676	5	M, F	young of year, subadult
Spotted seal	2007	Bering Sea	NA	782-1676	12	F, Madult, young of year, subadult	
Spotted seal	2009	Bering Sea	NA	358-1787	23	F, Madult, subadult, young of year	
Spotted seal	2010	Bering Sea	NA	358-1787, 15126	8	F, Myoung of year, adult, subadult	
Spotted seal	2014	Bering Sea	NA	15126	5	M, F	young of year, adult
Spotted seal	2016	Bering Sea	NA	19309	6	M, F	adult
Spotted seal	2018	Bering Sea	NA	19309	5	F	adult

North Slope Borough

Bearded seal	2012	Elson Lagoon, Utqiagvik	NA	15324	1	M	subadult
Bearded seal	2019	Pittalugruaq Lake	NA	20466	1	F	subadult
Spotted seal	2012	Tiny Island	Von Duyke et al in prep	15324	1	F	adult
Spotted seal	2014	Oarlock Island	Von Duyke et al in prep	15324	6	M, F	subadult, adult
Spotted seal	2014	Seal Island	Von Duyke et al in prep	15324	1	M	subadult

Spotted seal	2015	Oarlock Island	Von Duyke et al in prep	15324	6	M, F	subadult, adult
Spotted seal	2016	Pittalugruaq Lake	Von Duyke et al in prep	15324	3	F	subadult
Spotted seal	2017	Pittalugruaq Lake	Von Duyke et al in prep	15324	1	M	subadult
N. Pacific Wildlife Consortium							
Spotted seal	2009	Kamchatka Peninsula	NA	NA	3	F	adult