



Quantifying the impacts of bias in landcover data on global change analyses

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Blah blah.

landcover | bias | remote sensing | agriculture | crop yield | harvested area carbon agent-based model | landscape

Abbreviations: GTI, GeoTerralmage: SSA, sub-Saharan Africa

he nature and distribution of landcover is a fundamental determinant of many environmental and social processes 3 that drive or are affected by global change (1), such as agri-4 cultural production and food security (2-4), carbon cycling 5 (5, 6), biodiversity loss (7, 8), or demographic changes (9). 6 Landcover maps are therefore critical for understanding the 7 nature and impact of such changes (10), and they need to be 8 accurate at the finest scales at which the underlying processes 9 operate. For example, agricultural productivity and nutri-12 farming still dominates (11, 12). To understand agriculturally 13 driven processes, it is thus necessary to accurately delineate 68 land surface hydrology and climatology, and human geogra-14 fields at their smallest grain size, and to do so at regional to $_{15} \; {\rm global} \; {\rm scales} \; {\rm to} \; {\rm have} \; {\rm a} \; {\rm consistent} \; {\rm set} \; {\rm of} \; {\rm maps}.$

Landcover data can only be developed with satellite imag-17 ing, but often the average size class of the cover type of interest 70 **Study area and landcover data** 18 is smaller than the sensor resolution, or spectrally indistinct 71 Our study focused on South Africa, which comprises nearly 23 ing the most rapid land use changes, where the aforementioned 76 types that are found throughout the rest of SSA. 24 sources of bias tend to be most pronounced (15–17).

36 have a better understanding of the biases between landcover 89 ing cropped from non-cropped areas. 37 datasets or in relation to country-level statistics (16, 17, 21) $_{38}$ than we do of how error changes over spatial gradients or as a 39 function of aggregation scale.

Being unable to fully quantify the errors in landcover maps 41 of course makes it difficult, if not impossible, to quantify their 42 impact on downstream analyses. There has been some work 43 examining how such error influences climate simulations (22), 44 agricultural land use patterns (23), and carbon flux (24) and 45 human population estimates (9), but these either use simu-46 lated landcover errors (22) or compare relevant differences in 47 estimates between different satellite-derived landcover maps 48 (9, 24). The exception is (23), who use a high quality, ground-

49 collected reference map detailing farm land use parcels in cen-50 tral Belgium, but the number of sites and region were both 51 fairly restricted, and the parcels were not spatially contigu-

There is thus an urgent need to more precisely quantify land-54 cover map errors and how these vary over large regions, partic-55 ularly for the regions where landcover is changing most rapidly 56 yet is most poorly known. We address this need in this study, 57 using a unique, high accuracy agricultural landcover map for 58 South Africa to quantify the errors in several latest generation 59 landcover maps that are broadly used in global change studies. 60 We use this information to examine how i) landcover proper-61 ties and related classification schemes influence error, ii) how 62 these errors change with aggregation scale, with the specific 63 goal of determining "safe" scales for making area-based cal-64 culations, and 3) how these errors propagate through several 10 ent loadings can vary greatly between neighboring fields, and
65 different forms of downstream analyses that broadly represent 11 field sizes are often <2 hectares in regions where smallholder 66 the global change research focus areas, including biogeochem-67 ical and land use change studies, food security assessments,

19 from other neighboring covers, which propagates classification 72 6% of sub-Saharan Africa's (SSA) landmass, and has a large, 20 error (10, 13, 14). The result is that landcover datasets are 73 diverse agricultural sector, ranging from large commercial op-21 generally inaccurate at finer scales and greatly differ between 74 erations to smallholder farms (25, 26). This diversity suggests 22 one another, particularly in those parts of the world undergo- 75 that the country's agricultural landcover spans the range of

The South African government commissioned a whole coun-These errors are well-known (16, 17, 10, 18, 19), and there 75 try cropland boundary map in order to stratifying the annual 26 are a variety of efforts underway to improve landcover maps, 79 aerial crop type census used to calculate harvested area es-27 particularly for agriculture (20, 14). What is less known is the ∞ timates (27). The map was made by trained workers who 28 degree to which these errors bias measurements built upon 81 visually interpreted high resolution satellite imagery and man-29 the distributional and areal information in landcover. An im- 82 ually digitized field boundaries following a standardized map-30 pediment to this understanding is that the errors are hard 83 ping protocol. The resulting vectorized field maps, which were 31 to quantify because spatially extensive reference data are not 24 made in 2007 and updated in 2011, provide a unique, high ac-32 available for most regions of the world-particularly over Africa 85 curacy reference dataset of both crop field distribution and 33 and other developing regions. Errors assessment therefore typ- 86 size classes. We converted the vector data into a rasterized es-34 ically rely on a small number of ground truth points or survey 87 timated of cropland percentage at 1 km resolution (henceforth 25 data aggregated to political boundaries. For this reason, we so the "reference map"), which was 97% accurate in distinguish-

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We compared our reference percent cropland estimates to 155 Bias 91 those created from four satellite-derived landcover datasets. 156 Landcover bias. We created the 1 km reference and landcover 92 We obtained South Africa's 30 m resolution National Land- 157 maps, removing all croplands marked as communal or small- 93 cover map (SA-LC) for 2009 (28), the 500 m resolution MODIS 158 holder farmland in the reference vector maps (individual fields ⁹⁴ Landcover for 2011 (29, 30), the 300 m resolution GlobCover ¹⁵⁹ were not mapped with the same precision), as horticulture, or 95 2009 (31), and the new 1 km Geo-wiki hybrid-fusion crop- 160 plantation forestry (SI), and calculated the total cropland ex-96 land map for Africa (18). We chose these particular datasets 161 tent in the remaining area (1,081,000 km², or 90% of South 97 because they are nearly contemporaneous with our reference 162 Africa). The 2011 reference map showed a cropland area of 98 data, and represent the major types of landcover products used 163 104,310 km², which the SA-LC and GeoWiki maps overes-99 by researchers: SA-LC typifies the higher resolution, Landsat- 164 timated by 26 and 5.8%, respectively, and GlobCover and $_{100}$ derived maps that are developed individually for many coun- $_{165}$ MODIS underestimated by 21 and 26.1%. 101 tries (32), MODIS and GlobCover are widely used global-scale 166 We then aggregated each map to 5, 10, 25, 50, and 100 km 102 products (33, 34), while Geo-Wiki incorporates the first three 167 resolutions, and subtracted the four landcover maps from the 103 datasets and is the current state of the art for agricultural 168 reference map at each scale of aggregation to asses the spatial 104 landcover maps. We extracted the cropland classes from the 169 patterns of bias (Fig. 1). 105 first three datasets and converted these to 1 km resolution 106 percent cropland estimates (hereafter simply the "landcover 107 maps"), resulting in 4 maps to compare to our reference crop-108 land map (the "reference map").

109 Quantifying Error

110 We used these maps to first quantify error in cropland area 111 estimates. We calculated error as the difference between the 112 reference and landcover maps at different scales of aggregation 113 (1 to 100 km), in order to estimate bias and how it varies with 114 scale. Next, we assessed how bias correlates with the amount 115 of cropland cover in agricultural landscapes, to gain insight 116 into how landscape patterns may affect error.

We undertook five further analyses to investigate how map 118 error can impact assessments founded on landcover maps. 119 These include first-order analyses, in which values for a vari-120 able of interest are mapped to particular cover type(s), and 121 second-order analyses, in which a process model draws on the 122 cover types' values to calculate an output value. We created 123 four datasets to represent first order analyses. The first was 124 a series of maps of vegetated carbon stocks created following 125 the methodology of Ruesch and Gibbs' (35). The second was 126 constrained cropland percentage maps, which, following Ra-127 mankutty et al (36) were adjusted so that their total cropland 128 areas matched provincial-level reported cropland totals. Using 129 these adjusted cropland percentage maps, we disaggregated 130 district-reported maize harvested area and yields (following 131 37). We then compared differences between total carbon stock 132 estimates calculated from the reference map with those from 133 the four cropland maps, and again examined how these dif-134 ferences changed as a function of aggregation scale. We made 135 the same comparisons for total maize harvested area, average 136 yield, and total production.

137 For the second-order analyses, we examined how cropland 138 cover errors influence 25 km resolution monthly evapotran-139 spiration estimates produced using the Variable Infiltration 141 ample, we used the cropland maps to adjust the seasonally The eastern to northern margins (red areas in Fig. 1). varying, landcover-specific leaf area index (LAI) values that 175 116 average shades 121 142 varying, landcover-specific leaf area index (LAI) values that 176 34% at 1 km, meaning that each map underestimates cropland 142 varying, landcover-specific leaf area linex (LEG) values that I km, meaning that each map directly larger than 143 VIC uses to partition water vapor fluxes into their evaporate tive and transpirative components. In the second example, we sexamined how these errors can impact the parameterization of an agent-based food security model (39). Spatially-explicit, larger transported in land change in land chan The SA-LC map uniformly overestimates cropiand through147 agent-based models are frequently employed in land change
148 science, and require an initialization step to assign landscape
149 resources to model agents (e.g. 40–42). In this case, we used
150 the cropland maps to allocate farmland to agents represent151 ing individual households in political districts, with the model
152 mindividual households its initial cropland holdings using a
153 mindividual household its initial cropland holdings using a
154 The SA-LC map uniformly overestimates cropiand through155 much lower at -8% at 1 km to -6% at 100 km (Fig. S1). The
156 GeoWiki map has a very heterogeneous pattern of bias, hav156 km, which changes to a smaller overestimation bias (-2 %) at
157 mindividual household its initial cropland holdings using a 153 function that considers total district cropland and how much 154 cropland is near the agent's location.

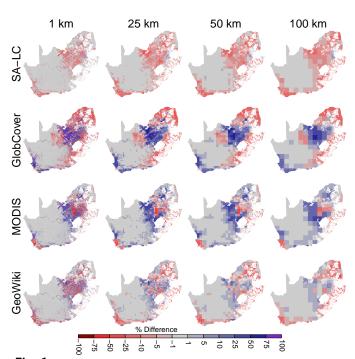


Fig. 1. Differences in percent cropland estimates between the reference map and each of the four landcover maps. Rows indicate the landcover map being assessed (by subtraction from the reference map), while columns refer to resolution of aggregation. White indicates areas with no data where communal farmlands or plantation forests were removed

The most pronounced extreme biases are in the MODIS and 171 GlobCover maps, with both underestimating cropland extent 172 by 10-75% in the center of the country (blue areas in Fig. 173 1, and the dominant production region), and overestimating $_{175}\,\mathrm{The}$ average biases for GlobCover and MODIS are 21% and

186 100 km (Fig. S1).

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188 tigate the degree to which landscape features influence land- 226 type was adjacent to cropland (Table S1), because differences 189 cover map bias, we extract all pixels in agricultural areas 227 between total stock estimates are diluted by the large area $_{190}$ (>0.5% cropland) of the 1 km reference map using the the $_{228}$ (~ 50 -70%) of the country having no cropland. Comparing the 191 boundaries of 354 magisterial districts (South Africa's finest 229 carbon stocks from each map's agricultural area reveals much 192 administrative unit, which average 3,445 km² in size; SI Ap- 230 greater differences (Table S1). SA-LC overestimates carbon 193 pendix, Fig. S3), and averaged the values of these pixels 231 stocks by just 2% when the adjacent cover type is forest, and 194 within each district. These provided an estimate of the inten-232 up to 15% when it is spare cover. MODIS ranges from negli-195 sity of cropping within the farmed parts of the landscape (with 233 gible overestimates in denser carbon classes (forest, secondary 196 districts defining the landscape scale), as well as the degree of 234 forest, and shrublands) to 8-13% underestimates for the grass-197 mixing between cropland and other land covers. We then ex- 235 land and sparse classes. GeoWiki underestimates for all types, $_{198}$ tracted the cropland map biases for the same locations, and $_{236}$ from <1% for sparse cover to 8% for forest. GlobCover shows 199 calculated the mean absolute bias within each district. Abso-237 the largest total carbon stock bias, with over-estimates rang-200 lute bias describes the magnitude of bias but not its direction, 238 ing from for sparse lands to 162% for forest. The magnitude 201 and is more informative about the likelihood of the map being 239 of this bias is due to false positives-GlobCover identified crop-202 biased at any given point in the landscape than actual bias, as 240 land in nearly 50% of pixels, compared to 30% for the other 203 positive and negative biases can cancel each other out within 241 three cropland maps. 204 small areas.

206 bias (log-transformed) shows that absolute bias peaks at 50% 244 timated and the surrounding cover type is of higher carbon 207 cropland cover for all but the GlobCover map (which contin- 245 density than cropland, carbon density is overestimated. For 208 ued to increase with cropland cover), and is lowest when the 246 lower density (grassland and sparse vegetation), there may be 2009 landscape is dominated either by cropland or other cover types 247 slight underestimates. These tendencies are reflected in the 210 (Fig. 2). In other words, bias is highest when cropland cover 248 average of each map's spatial biases over agricultural areas 211 is mixed evenly with other cover types.

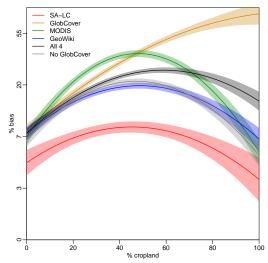


Fig. 2. The relationship between the mean absolute bias in cropland maps and cropland cover within agricultural areas (reference map pixels having >0.5% cropland), averaged within the boundaries of magisterial districts (n = 345), as fit with a generalized additive model. Prediction curves are color-coded to the different cropland maps, with the solid line indicating predicted absolute bias, and the lighter shading the standard error of the coefficients. Models were also fit to the mean absolute biases across all four maps (black curve), and from all maps exlcuding GlobCover (grey

212 The impact of bias on calculating carbon stocks. We used the 213 data of (35) to calculate carbon densities for African forests, 214 shrublands, croplands, grasslands, and sparse habitats. We as-215 signed the cropland carbon values to map cells in proportion 216 to their cropland cover. For the non-cropland proportions, we 217 assigned the carbon value from each of the other types, such 218 that we created four different carbon maps for each landcover 219 map at each aggregation scale (Fig. S4), which allowed us to 220 test how carbon estimates vary as a function of i) bias in the 221 cropland estimates and ii) the characteristics of adjacent cover

At the 1 km scale, the differences between country-wide to-224 tal carbon stocks made using any of the cropland maps and

187 The role of landscape characteristics in shaping bias. To inves- 225 the reference were within $\pm 7.3\%$, regardless of which cover

The spatial patterns of carbon bias (Fig. S4) are a reflection 242 A generalized additive model fit to district-level absolute 243 of the cropland biases (Fig. 1). Where cropland is underes-²⁴⁹ (Fig. 3). For example, MODIS and GlobCover overestimate 250 carbon by $\sim 50\%$ on average at 1 km resolution when forest is 251 the cropland-adjacent cover (stars in semi-transparent green 252 and gold bars, Fig. 3; Table S2). When it is sparse vegetation 253 (open circles in Fig. 3), MODIS underestimates by nearly 3% 254 at all scales, whereas maps that overestimate cropland (e.g. 255 SA-LC, semi-transparent red) overestimate carbon density for 256 this cover. Overall, GeoWiki has the lowest average spatial 257 biases, for all cover types and all resolutions. It's maximum 258 bias is a 12% overestimate at 1 km for forest cover, thereafter 259 all biases are less than 10%, and converge to \pm 2% above 1 260 km aggregation (Fig. 3, Table S2). All maps' biases are within $_{261}$ +/-10% bias after aggregation to 25 km.

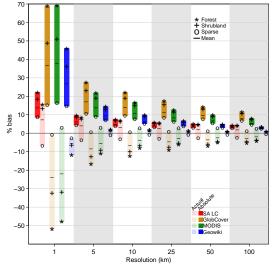


Fig. 3. The average spatial biases in carbon density estimates based on cropland maps, expressed as a percent difference relative to the reference map. The values of actual bias (incorporating bias direction and magnitude) and absolute bias (magnitude only) are presented in semi-transparent and solid colors, respectively, with colors denoting the specific cropland map and symbols indicating which cover type was used to calculate cropland adjacent carbon density. The bar represents the mean biases calculated across each of the 5 cover types. Shrubland and grassland bias values are near zero and not shown for display clarity.

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263 bars in Fig. 3) generally follow that of actual bias, but with 300 GlobCover's mean yield bias declined linearly with aggrega-264 somewhat higher magnitudes and a few important differences. 301 tion (Fig. 5). 265 The most notable is that Geo-Wiki has fairly large absolute 302 In contrast to yields, production errors were unbiased in 266 bias at 1 km, averaging 27% across across cover types (line in 303 aggregate (Fig. 5). The statistical constraints on harvested 267 solid blue bar, Fig. 3, Table S2), which is close to the 36-37% 304 and cropland areas led to the canceling out of spatial errors in 268 for GlobCover and MODIS. SA-LC has the lowest absolute 305 production estimates, which is evident in the checkerboard-like 269 bias across scales, averaging (across cover types) from 14% at 306 pattern in maps of production biases (Figure S7). However, 270 1 km to 3% at 100 km. (Fig. 3, Table S2). The increase in 307 this pattern of bias means that the average absolute biases 271 Geo-Wiki's absolute spatial bias relative to SA-LC's can be at- 308 in production estimates are large, between 40 to 55% for Ge-272 tributed to the spatially heterogeneous nature of its cropland 300 oWiki, MODIS, and GlobCover at 1 km, and remaining gen-273 biases (Fig. 1) compared to the other three cropland maps. 310 erally high (<10%) for these three datasets even up to 50-100

275 used the reference dataset to calculate total cropland area 313 Kill).

275 used the reference dataset to calculate total cropland area 314 Absolute mean biases in yield were also substantial, and within South Africa's nine provinces, and then adjusted the the four cropland maps so that they each summed to the same within each province (per 36). Despite this statistical to be satisfied as the statistical to be satisfied to be satis 279 constraint, the updated cropland maps still had substantial $_{280}\,\mathrm{spatial}$ biases similar in pattern (Fig. S5) to those in the un-281 adjusted maps (Fig. 1), and we evaluated how these biases 318 Bias in evapotranspiration estimates. Compared to carbon $_{284}$ (n = 354, mean area = 3,445 km²) agricultural census data $_{321}$ averaged to less than +/-1%. There are several hotspots of 287 aggregation scale.

 $_{289}$ markedly different to those on the reference map, particularly $_{326}$ ern coasts, particularly the latter where the ET overestimates 290 in the lower density cropland areas in the center of the coun- 327 reach 25%. These locations correspond primarily to margins try, where GlobCover overestimated yields and MODIS and 328 of major crop production regions—in the center is the western-²⁹² GeoWiki (to a lesser extent) underestimated them at 1 km ³²⁹ most boundary of the summer rainfall growing region, where 293 resolution (Fig. S6). However, the mean spatial bias in yields 330 maize is the primary crop, while the bias hotspot up the west 294 at this resolution was pronounced only for GlobCover, equiv- 331 coast marks one of the two boundaries of the winter rain-²⁹⁵ alent to nearly 60% of the mean reference yield of 3.4 tons ³³² fall region where wheat farming dominates(25). These two ²⁹⁶ ha⁻¹, compared to +/-5% for the other maps (Fig. 5). Inter- ³³³ boundaries correspond respectively to approximately 400 and $_{297}$ estingly, GeoWiki and MODIS biases increased with aggrega- $_{334}$ 200 mm of growing season rainfall. 298 tion, peaking at 10 km where both datasets underestimated 335 SA-LC and GeoWiki also produced biased ET estimates

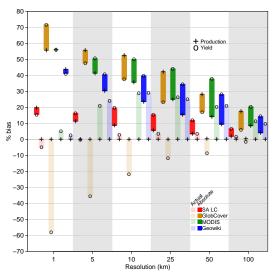


Fig. 4. The mean spatial biases in disaggregated maize yield estimates, and the bias in total crop production estimates calculated from these yields and constrained crop area estimates. The values of actual bias (incorporating bias direction and magnitude) and absolute bias (magnitude only) are presented in semi-transparent and solid colors, respectively, with colors denoting the specific cropland map and symbols indicating whether the bias is related to one of two yield aggregation methods and associated crop production estimates.

The patterns of mean absolute spatial bias (solid colored 299 yields by 30%, thereafter declining to 10% at 100 km, whereas

311 km of aggregation (Fig. 5). SA-LC production biases were Bias in harvested areas, yield, and production estimates. We 312 lowest across all spatial scales (20% at 1 km to 2% in 100

282 affected crop production estimates. We derived maize yield 319 and crop related examples, actual and absolute biases in evap-283 and harvested area estimates by disaggregating district-level 320 otranspiration calculated using the VIC model were small and 285 (43) following Monfreda et al. (37), and used that to calcu- 322 discrepancy, however, which are evident in maps of spatial bi-286 late spatial biases in yield and production estimates at each 323 ases (Fig. 6). The most pronounced of these are the 5-15% 324 overestimates by VIC initialized with MODIS and GlobCover The yields disaggregated onto the cropland maps were 325 in the center of the country, and along the southern and west-

> 336 along the southern and western coasts, but here the tendency 337 was to underestimate ET, while biases in the center of the

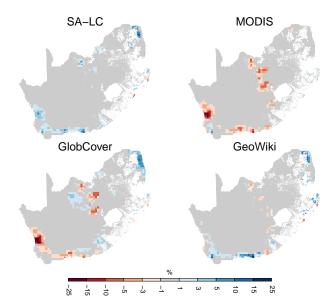


Fig. 5. Differences in annual mean evapotranspiration estimates from 29-year runs of the VIC land surface hydrology model when initialized with LAI response curves derived from the reference map, versus those from the four cropland maps.

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338 country were either negligible to absent. All but MODIS un- 393 and the proximity of other agents) within which cropland is $_{339}$ derestimated ET by 5-15% in the northern tip of the country. $_{394}$ sought when agents are seeded onto the landscape.

340 Initialization biases in agent-based models. We used an agent-341 based model of food security that represents the interactions 342 between hundreds of individual farming households over multi-343 ple seasons (39). The model is thus computationally intensive, 344 and, like many spatial ABMs, run over smaller geographic 345 domains (e.g. districts, rather than an entire country) and 346 at higher spatial resolutions (10s to 100s of meters) that are 347 needed to represent the different land units of single farm-348 ers. To match these computational characteristics, we selected 349 four contiguous magisterial districts (ranging from 1,040-1,343 350 km², Fig. S8) in the eastern part of the country, having be-351 tween 28-45% of their areas devoted to cropland, according to 395 **Discussion** 352 the reference map.

binary cropland/non-cropland cover types with 100 m resolu398 tions/considerations you see from the results. $_{355}$ tion, which matches the typical field size (1 ha) for smallholder $_{399}$ 356 farmers in household survey data (collected in Zambia) used in 357 developing the agent-based model (39). These surveys found 400 • $_{358}$ mean household crop field area to be 2 ha, which we divided $_{401}$ $_{359}$ into reference cropland areas to estimate the total households $_{402}$ 360 within each district. We then initialized the model by as- $_{361}\,\mathrm{signing}$ each household agent two cropland pixels. In order 403 $_{362}\,\mathrm{to}$ emulate the natural groupings of communities, the model 404 $_{363}$ only assigns a household fields that are within 1.5 km of other 405 $_{364}$ agents' fields, and if those pixels have not already been al- 406 $_{365}$ located to another agent. The model thus iteratively grows $_{407}$ $_{366}$ "communities" until all households are assigned cropland, or $_{408}$ 367 all available cropland is allocated.

We used each of the five maps to initialize the model, and 410 $_{369}$ assessed how bias in cropland cover estimates impacted the $_{411}$ $_{370}$ initialization process, in terms of how many agents were not $_{412}$ $_{371}$ assigned fields, and how many fields were left unallocated. For $_{413}$ $_{372}\,\mathrm{the}$ first metric, there was a one-to-one relationship between $_{414}$ $_{373}\,\mathrm{the}$ percent by which cropland area underestimated and the $_{415}$ $_{\rm 374}\,{\rm percent}$ of households left without and farmland (Fig. 6, left $_{\rm 416}$ $_{\rm 375}$ panel), the most extreme examples occurring when MODIS $_{\rm 417}$ 376 cropland was used to initialize the model in districts 1 and 2, 377 where ~85% of agents could not be allocated farmland because 418 • Bias as a function of cropland cover $_{378}$ MODIS underestimated cropland by that amount $\sim 85\%$. No $_{419}$ $_{379}$ agents were left without fields where maps overestimated crop- $_{420}$ 380 land extent (GeoWiki, SA-LC), but in these cases the percent 381 of cropland left unallocated matched the size of the overesti- $_{382}$ mate (e.g. $\sim\!\!20\%$ for SA-LC, Fig. 6 right panel). Interestingly, $_{_{423}}$ 383 the overall relationship between land allocation and cropland 424 $_{\rm 384}$ bias was U-shaped, as the model also failed to give land to $_{\rm 425}$ $_{385}$ agents when cropland was underestimated by 50% or more $_{386}$ (Fig. 6, right panel), as exemplified by the MODIS result for 426 $_{387}$ districts 1 and 2, where 7-12% of cropland was left unallocated 427 $_{388}$ despite the fact that only 85% of agents had no land. This cu- $_{428}$ 389 rious relationship occurred because cropland tends to cluster, 429 390 and when it is underestimated, the size of these clusters is 430 391 small, resulting in islands of cropland that fall outside of the 431 392 search radius (which is constrained by an absolute distance 432

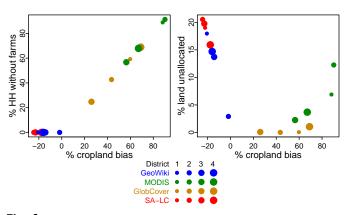


Fig. 6. Biases in agent-based model initialization relative to cropland map biases, measured in terms of the percent of households having no cropland allocated (left) and the percent of cropland left unallocated (right). Dot sizes correspond to district numbers, colors represent the landcover map.

396 Some points to make thrown out now because out of time. We disaggregated the cropland percentages in all maps to 397 Others will be added. Please add any important implica-

Some points from my notes last year:

- Bias as a function of scale
 - 1. At 1 km resolution all landcover products are still fairly biased.
 - 2. Bias drops to acceptable levels quickly for geowiki-at 5X5 km, mean bias is just 1% (overestimated). The absolute bias for this dataset is 10% or lower from 10X10 km resolution and coarser.
 - The SA dataset's bias is fairly consistent but low across all levels of aggregation, amounting to no more than an 8% overestimate of cropland with absolute bias of similar magnitude.
 - 4. MODIS and GlobCover biases (mostly of underestimation) do not dissipate until the higher levels of aggregation. MODIS's actual bias (under-estimation) falls below 10% at 20 km resolution, but the absolute bias remains above 10% until more than a 100-fold aggregation is done (>100 km resolution). For GlobCover, it is still too high.
- - 1. Classification algorithms are thus more error-prone where landcover is mixed/heterogenous.
 - 2. The exception to this lies in the GlobCover dataset, where bias primarily increases as a function of cropland cover. The reason for this is that GlobCover's cropland classes do not provide for 100% cropland cover, so aggregation tends to exacerbate underestimates.
 - 3. Thus caution is needed when aggregating a mixed pixel
 - An example illustrats this: take 4 1 km pixels, 2 of which are 100% cropland, 2 of which are other cover types. Imagine a landcover product classifies 3 of these as cropland (2 correct, 1 an error of commission), using a cropland class that is defined as 50%



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cropland. Aggregating the actual fraction by a fac- 491 More blather. Materials and Methods tor of 4 will result in a new 4 km pixel having 50% 492 Methods. Perhaps it is right SI Materials and Methods. cropland, whereas aggregating the landcover product's pixels will give just 38% cropland, even when 493 factoring in the incorrect classification.)

- Bias as a function of method
 - 1. Higher resolution and ensemble-based approaches have less bias
 - 2. geowiki represents a fusion of multiple coarse resolution data sources that has undergone extensive validation using a crowdsourcing approach
 - 3. the SALC dataset is based on 30 m landsat data, but incorporates a range of ancillary data and expert judge-
 - 4. MODIS and GlobCover data are effectively single $_{500}$ 1. Lambin EF (1997) Modelling and monitoring land-cover source/single algorithm.
 - 5. Newer points begin here
 - 6. Statistically constrained constrained landcover estimation approaches provide accurate area-based inferences 505 when aggregated. But spatial errors are still high, as 506 seen with GeoWiki and production/yield estimates. Us- 507 ing these to identify yield gaps at specific map locations 508 is inappropriate, or even for a larger location if it does 509 not coincide with the geographic boundaries of the sta- 510 tistical unit.
 - 7. Constrained estimates are also dependent on the accuracy of the statistics.
- Fix above to have section on bias for global change studies 515 460
 - 1. Scales at which it is safe to estimate values of say carbon $_{\scriptscriptstyle{517}}$ stocks.
 - 2. Above point about bias in disaggregated yield estimates 519 - no point mapping these out. A new approach might be 520 to take these statistically reported yields and then com- 521 bine them with satellite data to estimate yield variability $\,^{522}$ within the district. That way would have meaningful 523 reason for disaggregating yields, and would be pegged 524 to real yield values, which would help minimize errors $^{525}\,$ in remote sensing of yields.
 - 3. Something on ET doesn't seem to matter much, but $_{528}$ land-atmosphere interactions can make these discrepan- $_{529}$ cies meaningful, particularly since biases occur in arid 530 areas where a lot of irrigation happens-can cause sig-531 10. See L et al. (2015) Improved global cropland data as an nificant impacts on regional climate. etc. etc. Also we $_{532}\,$ didn't change out land cover types, and the vegetation $_{533}$ in SA around the cropland will have reasonably simi- 534 11. lar LAI and ET responses (I think), thus impact more $_{535}$ muted than it might be elsewhere (e.g. in forested land- 536
 - 4. Agent-based models. Tom, Peng, something of significance/implications of this, please
- Will need a section on way forward for data, etc. Key role 541 483 484 ods, vectorized field boundaries seem to be highly valuable, 543485 Mapping Africa, Stephanie's paper, Geo-Wiki, etc are the 544 14. Estes L et al. (2015) DIYlandcover: Crowdsourcing the 486 way ahead. 487

Mixed landscapes increase the chances of omission and com- 547 15. 489 mission errors by increasing the number of cover classes, or 548 490 because such landscapes are less spectrally distinct (14)

Describe weighted mean bias reasons

494 Digital RCD Analysis.

495 Appendix: App 1

496 Appendix

497 This is an example of an appendix without a title.

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499 References

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- change processes in tropical regions. Progress in Physical Geography 21(3):375-393.
- 2. Lark TJ, Salmon JM, Gibbs HK (2015) Cropland expansion outpaces agricultural and biofuel policies in the United States. Environmental Research Letters 10(4):044003.
- 3. Wright CK, Wimberly MC (2013) Recent land use change in the Western Corn Belt threatens grasslands and wetlands. Proceedings of the National Academy of Sciences 110(10):4134-4139
- 511 4. Licker R et al. (2010) Mind the gap: how do climate and agricultural management explain the yield gap of croplands around the world? Global Ecology and Biogeography 19(6):769-782.
 - Asner GP et al. (2010) High-resolution forest carbon stocks and emissions in the Amazon. Proceedings of the National Academy of Sciences 107(38):16738–16742.
 - Gaveau DLA et al. (2014) Major atmospheric emissions from peat fires in Southeast Asia during non-drought years: evidence from the 2013 Sumatran fires. Scientific Reports 4.
 - Newbold T et al. (2015) Global effects of land use on local terrestrial biodiversity. Nature 520(7545):45–50.
 - Luoto M, Virkkala R, Heikkinen RK, Rainio K (2004) Predicting bird species richness using remote sensing in boreal agricultural-forest mosaics. Ecological Applications 14(6):1946-1962.
 - 9. Linard C, Gilbert M, Tatem AJ (2010) Assessing the use of global land cover data for guiding large area population distribution modelling. GeoJournal 76(5):525–538.
 - essential ingredient for food security. Global Food Security 4:37-45.
 - Jain M, Mondal P, DeFries RS, Small C, Galford GL (2013) Mapping cropping intensity of smallholder farms: A comparison of methods using multiple sensors. Remote Sensing of Environment 134:210–223.
- Debats S, Luo D, Estes L, Fuchs T, Caylor K (year?) A 538 12. generalized computer vision approach to mapping agri-539 cultural fields in Sub-Saharan Africa. Remote Sensing of Environment.
- of accurate landcover, particularly agricultural. New meth- 542 13. Lobell DB (2013) The use of satellite data for crop yield gap analysis. Field Crops Research 143:56-64.
 - creation of systematic, accurate landcover maps. PeerJ PrePrints 3:e1266.
 - Estes LD et al. (2013) Projected climate impacts to South African maize and wheat production in 2055: a comparison of empirical and mechanistic modeling approaches.





- $Global\ Change\ Biology\ 19(12):3762-3774.$
- 551 16. Fritz S, See L, Rembold F (2010) Comparison of global 600 and regional land cover maps with statistical information $_{\rm 601}$ 552 for the agricultural domain in Africa. International Jour- 602 31. 553 nal of Remote Sensing 31(9):2237-2256. 554
- 555 17. Fritz S et al. (2011) Cropland for sub-Saharan Africa: A 604 synergistic approach using five land cover data sets. Geo- 605 32. Fry J, Coan M, Homer C, Meyer D, Wickham J (2009) 556 physical Research Letters 38:L04404. 557
- 558 18. Fritz S et al. (2015) Mapping global cropland and field 607 559 size. Global Change Biology 21(5):1980–1992.
- 560 19. Verburg PH, Neumann K, Nol L (2011) Challenges in us- 609 33. ing land use and land cover data for global change studies. 610 561 Global Change Biology 17(2):974–989. 562
- 563 20. Fritz S et al. (2012) Geo-Wiki: An online platform for 612 34. improving global land cover. Environmental Modelling & $_{613}$ 564 Software 31:110–123.
- Kaptu Tchuent AT, Roujean JL, De Jong SM (2011) Com- 615 566 21. parison and relative quality assessment of the GLC2000, 616 35. 567 GLOBCOVER, MODIS and ECOCLIMAP land cover 617 568 data sets at the African continental scale. International 618 Journal of Applied Earth Observation and Geoinformation 619 570 13(2):207-219.571
- 572 22. Ge J et al. (2007) Impacts of land use/cover classifica-621 tion accuracy on regional climate simulations. Journal of 622 573 $Geophysical\ Research:\ Atmospheres\ 112 (D5): D05107.$ 574
- Schmit C, Rounsevell MDA, La Jeunesse I (2006) The lim- 624 575 23. itations of spatial land use data in environmental analysis. 625 576 Environmental Science & Policy 9(2):174-188. 577
- 578 24. Quaife T et al. (2008) Impact of land cover uncertainties 627 37. on estimates of biospheric carbon fluxes. Global Biogeo- 628 chemical Cycles 22(4):GB4016. 580
- Hardy M, Dziba L, Kilian W, Tolmay J (2011) Rainfed $_{630}$ 581 25. Farming Systems in South Africa in Rainfed Farming Sys- 631 38. 582 tems, eds. Tow P, Cooper I, Partridge I, Birch C. (Springer 632 Netherlands), pp. 395-432. 584
- 585 26. Estes LD et al. (2014) Using changes in agricultural util- 634 ity to quantify future climate-induced risk to conservation. 635 39. 586 Conservation Biology 28(2):427–437.
- 588 27. Fourie A (2009) Better Crop Estimates in South Africa. 637 40. $ArcUser\ Online\ (1).$
- 590 28. SANBI (2009) National Landcover 2009, (South African 639 National Biodiversity Institute; National Department 640 591 of Environmental Affairs and Tourism, Pretoria, South 641 41. 592 Africa), Technical report.
- 594 29. DAAC) LPDAACL (2011) MODIS MCD12q1 Land Cover 643 (NASA EOSDIS Land Processes DAAC, USGS Earth Re- 645 596 sources Observation and Science (EROS) Center, Sioux 646 Falls, South Dakota), Technical report.

- 599 30. Friedl MA et al. (2010) MODIS Collection 5 global land cover: Algorithm refinements and characterization of new datasets. Remote Sensing of Environment 114(1):168–182.
 - Arino O et al. (2012) Global land cover map for 2009 (GlobCover 2009). (European Space Agency & Universit Catholique de Louvain).
 - Completion of the National Land Cover Database (NLCD) 1992-2001 Land Cover Change Retrofit Product, (U.S. Geological Survey), USGS Numbered Series 2008-1379.
 - Gross D et al. (2013) Monitoring land cover changes in African protected areas in the 21st century. Ecological Informatics 14:31–37.
 - Shackelford GE, Steward PR, German RN, Sait SM, Benton TG (2015) Conservation planning in agricultural landscapes: hotspots of conflict between agriculture and nature. Diversity and Distributions 21(3):357-367.
 - Ruesch A, Gibbs HK (2008) New IPCC Tier-1 global biomass carbon map for the year 2000. CarbonDioxideInformationAnalysisCenter(CDIAC),Oak Ridge National Laboratory, Oak Ridge, Tennessee.Available online at: http://cdiac. $gov/epubs/ndp/global_carbon/carbon_documentation.$ html.
- Ramankutty N, Evan AT, Monfreda C, Foley JA (2008) 623 36. Farming the planet: 1. Geographic distribution of global agricultural lands in the year 2000. Global Biogeochemical Cycles 22:19 PP.
 - Monfreda C, Ramankutty N, Foley JA (2008) Farming the planet: 2. Geographic distribution of crop areas, yields, physiological types, and net primary production in the year 2000. Global Biogeochemical Cycles 22:GB1022
 - Liang X, Lettenmaier DP, Wood EF, Burges SJ (1994) A simple hydrologically based model of land surface water and energy fluxes for general circulation models. Journal of Geophysical Research 99(D7):14415.
 - Chen P, Plale B, Evans T (2013) Dependency Provenance in Agent Based Modeling. pp. 180–187.
 - Manson SM, Evans T (2007) Agent-based modeling of deforestation in southern Yucatn, Mexico, and reforestation in the Midwest United States. Proceedings of the National Academy of Sciences 104(52):20678-20683.
 - Evans TP, Kelley H (2004) Multi-scale analysis of a household level agent-based model of landcover change. Journal of Environmental Management 72(1-2):57-72.
- Type Yearly L3 Global 500 m SIN Grid. Version 5.01, 644 42. Kelley H, Evans T (2011) The relative influences of landowner and landscape heterogeneity in an agent-based model of land-use. Ecological Economics 70(6):1075–1087.
 - 647 43. Africa SS (2007) Commercial Census of Agriculture, South Africa.





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