J. Ocean Univ. China (Oceanic and Coastal Sea Research)

https://doi.org/10.1007/s11802-019-3942-2 ISSN 1672-5182, 2019 18 (4): 913-918 http://www.ouc.edu.cn/xbywb/ E-mail:xbywb@ouc.edu.cn

Effects of Seawater Acidification on Early Development of Clam *Cyclina sinensis*

SUI Yanming*, ZHOU Kai, LAI Qifang*, YAO Zongli, and GAO Pengchen

Engineering Research Center of Saline-alkaline Water Fisheries, East China Sea Fisheries Research Institute, Chinese Academy of Fishery Sciences, Shanghai 200090, China

(Received June 25, 2018; revised August 21, 2018; accepted March 19, 2019) © Ocean University of China, Science Press and Springer-Verlag GmbH Germany 2019

Abstract Anthropogenic emission of atmospheric carbon dioxide (CO₂) has led to a rapid increase in atmospheric CO₂ concentration. Increasing atmospheric CO₂ can reduce seawater pH and carbonate ions, which may adversely affect the survival of the larvae of calcareous animals. *Cyclina sinensis* is a commercially and ecologically important species in several Asian countries. Living in coast shallow waters, this species has experienced the coastal environmental changes frequently throughout its life cycle. In this study, we simulated possible future seawater pH values including 8.2, 7.8 and 7.4 and examined the effects of ocean acidification on the early development of *C. sinensis*. Clam embryos were incubated for 48 h (2 d) in control and high-CO₂ seawater to compare embryogenesis, larval growth and swimming behavior. Fertilization rate was quite sensitive to pH, and moderate acidification could induce a significant decrease in fertilization rate. However, only extreme acidification could bring significant negative effect to hatching rate, body size, and average path velocity of trochophora. Moreover, with seawater acidification, *C. sinensis* needs much more time to reach the same developmental stage, which increases the risk of larva survival. Together with recent studies demonstrating negative impacts of high CO₂ on fertilization and larva swimming behavior, the results imply a future decrease of *C. sinensis* populations in oceans if its acclimation to the predicted environmental alteration does not occur.

Key words seawater acidification; Cyclina sinensis; fertilization; hatching rate; development; average path velocity

1 Introduction

Because of human activities, the extensive use of fossil fuels has released a large amount of CO₂ into atmosphere. It is estimated that about one third of the carbon dioxide released by humans into the atmosphere has dissolved into oceans, rivers and lakes (Millero, 1995; Feely *et al.*, 2004). The pH of the ocean might drop by 0.15–0.4 units at the end of the 21st century. This process is named as ocean acidification (OA) and is expected to affect various marine ecosystems (Hofmann *et al.*, 2010). In particular, organisms that use calcium carbonate to build their shells or skeletons are estimated to be much more negatively impacted (Dupont and Pörtner, 2013; Kroeker *et al.*, 2013).

Compared to the adult stage, calcifying organisms at their early developmental stages are much more susceptible to elevated pCO_2 because larvae cannot keep the acid-base balance well (Dupont *et al.*, 2008; Kurihara, 2008; Byrne, 2012; Andersen *et al.*, 2017). Additionally, the shell in larvae contains a thin layer of amorphous CaCO₃ which is relatively easy to dissolve (Gazeau *et al.*, 2013). Recently, numerous researches have focused on

the early development of marine calcifying organisms under OA (Gazeau et al., 2010; Kroeker et al., 2013; Szalaj et al., 2017). Basically, fertilization, development, metamorphosis, shell growth and survival rate are all significantly reduced at low pH for most species whereas marine calcifying organisms also perform different degrees of susceptibility to OA during their early developmental stages (Kroeker et al., 2013). For instance, in Saccostrea glomerata and Crassostrea gigas, both D-veliger percentage and D-veliger normality decline with pCO₂ increment and significant effects appeared under a moderate acidic condition; however, C. gigas exhibits more tolerance to OA than S. glomerata (Parker et al., 2010). Guo et al. (2015) compared the effects of OA on the early development of 3 species including *Haliotis diversicolor*, *H*. discus hannai and C. angulata, and found that OA negatively affects the 3 species, and the resistance of C. angulata to OA is stronger than the other two. These studies clearly showed that the response to OA in calcifying organisms is species specific.

Cyclina sinensis is a commercially and ecologically important species in several Asian countries, especially in the coast shallow waters of these countries (Ren et al., 2016). Thus, this species experiences the coastal environmental change frequently throughout its life cycle. So far, little has been known about the effect of OA on C.

^{*} Corresponding authors. E-mail: yanmingsui@foxmail.com E-mail: qifanglai@163.com

sinensis, especially on its early development. In this research, we measured the fertilization, hatching rate, and proportion of larvae at various developmental stages, as well as the larva size and average path velocity to assess the effect of possible OA on the early life stages of *C. sinensis* in the future.

2 Materials and Methods

2.1 Experimental Animals

Mature *C. sinensis* were obtained from Haining City, Zhejiang Province, China. About 1 kg of *C. sinensis* (in average, $31.5 \,\mathrm{mm} \pm 6.0 \,\mathrm{mm}$ in shell length and $13.0 \,\mathrm{g} \pm 2.5 \,\mathrm{g}$ in wet weight) were selected and transported to an aquaculture farm where the clams were acclimated for 72 h at seawater temperature $25\,^{\circ}\mathrm{C}$, salinity 18, dissolved oxygen (DO) concentration >7.0 $\,\mathrm{mg}\,\mathrm{L}^{-1}$ and pH 8.1. The clams were fed with *Isochrysis* spp. twice a day (<10⁵ $\,\mathrm{cells}\,\mathrm{mL}^{-1}$).

2.2 Seawater for Culture and Experiments

The experiment was done in nine 50 L tanks (3 replicate each pCO_2). Three pH levels were designed as pH 8.2

representing the current seawater hydrion concentration; pH 7.8 mimicking the fluctuation state in the sampling waters (8.2-7.7, Li et al., 2014) and other areas (Caldeira and Wickett, 2005; Melzner et al., 2013) and pH 7.4, an extreme pH especially in hypoxic zones (Cai et al., 2011; Chou et al., 2013). All the C. sinensis were transferred to the tanks, respectively, with the same handling condition. The pH was controlled by bubbling pure CO₂ using a pCO₂/pH feedback STAT system (DAQ-M) associated with WTW pH 3310 meters and SenTix 41 pH electrodes (Loligo Systems Inc.), and was manipulated by CapCTRL software (Loligo Systems Inc.). The salinity was guarded throughout the experiment using a multi-parameter meter (model 5200 A, YSI, USA). Total alkalinity (A_T) was measured by titration. Other parameters of the seawater carbonate chemistry, including pCO₂, calcite saturation state (Ω ca), and aragonite saturation state (Ω ar) were calculated from A_T and pH_{NBS} using CO₂sys (Lewis et al., 1998). Calculations depended upon a series of constants, K1 and K2, which were brought from Millero (2010). Seawater was sampled twice a day during the experiment, and the pH, total alkalinity, temperature, salinity and other calculated data are listed in Table 1.

Table 1 Seawater carbonate chemistry variables (mean \pm SD, n=5) over the experimental period

Treatments	Temperature ($^{\circ}$ C)	Salinity	pH_{NBS}	$A_T (\mu mol kg^{-l})$	pCO ₂ (μatm)	Ωca	Ωar
pH 8.2	25.10 ± 0.07	18.12 ± 0.08	8.16 ± 0.04	2263 ± 26	368 ± 45	4.91 ± 0.52	2.95 ± 0.31
7.8	25.08 ± 0.04	18.12 ± 0.17	7.83 ± 0.03	2257 ± 30	872 ± 79	2.49 ± 0.21	1.50 ± 0.12
7.4	25.06 ± 0.05	18.16 ± 0.08	7.37 ± 0.04	2232 ± 17	2631 ± 311	0.91 ± 0.09	0.55 ± 0.05

Notes: Seawater pH on the NBS scale (pH_{NBS}), temperature ($^{\circ}$ C), salinity, and total alkalinity (A_T; μ mol kg⁻¹) were used to calculate dissolved inorganic carbon (DIC), CO₂ partial pressure (pCO₂; μ atm), as well as aragonite (Ω ar) and calcite (Ω ca) saturation states.

2.3 Ocean Acidification Effect on the Embryonic Development of *C. sinensis*

The air-exposure and flow stimulation methods were adopted for spawning (Guo *et al.*, 2016). Briefly, clams which experienced a 4h air-exposure were transported to corresponding seawater with a flow generated by a submersible pump. Then the clams produced sperms and eggs. Based on the measured gamete concentrations, 10:1 was adopted as experimental sperm-egg ratio to avoid possible error caused by unreasonably high sperm-egg ratio (Shi *et al.*, 2017a). After 10 minutes, large tissue debris was removed from tank using a 200-mesh sieve and the eggs were collected with a 400-mesh sieve. The eggs were rinsed with corresponding seawater and then transferred to corresponding tanks (about 15 eggs mL⁻¹). Larvae were fed

with a mix of diet of *Chaetoceros* sp. and *Isochrysis galbana* from 8 hours post fertilization (hpf). At least 50 eggs or larvae were selected each tank and fixed with 4% paraformaldehyde at 2, 8, 16, 24, 48 hpf for further observation.

Samples for measuring fertilization were taken at 2 hpf and the fertilization rate (%) was calculated by checking the cleaved embryos among eggs:

Fertilization rate =
$$\frac{\text{Number of cleaved embryos}}{\text{Number of total eggs}} \times 100$$
.

Samples for hatching rate (%) were taken at 16 hpf. Hatching rate was quantified by observing the trochophore larvae and the D-shaped larvae that swam out of the fertilization eggs:

Hatching rate =
$$\frac{\text{Number of Trochophores} + \text{Number of D-shaped larvae}}{\text{Total number of eggs} \times \text{Fertilization rate}} \times 100 \text{ .}$$

For developmental measurement of *C. sinensis*, the proportions of each developmental stage were calculated at 24, 48 hpf. Classification of deformed and normal larvae depended on morphological criteria set by Guo *et al.*. (2016) for clam.

Larval size was measured at 16, 24, 48 hpf using Image-Pro Plus software. A total of 50 larvae were measured

each replicate each pCO_2 .

2.4 Ocean Acidification Effects on Average Path Velocity of Trochophore in *C. sinensis*

Trochophore velocity was measured at 8 hpf using a dissecting microscope (Leica MZ 125, 4 objective) equipped with a camera Sony Exwave HAD (Suquet *et al.*,



2012). Average path velocity (VAP) was measured using a CASA plug for the Image J software. Calibration parameters adopted in this experiment were as following: frame rate, 25 frame s⁻¹; larval size range, 1 to 1000 pixels; minimum VAP for motile larvae, 30 μm s⁻¹; minimum number of larvae observed, 30; minimum track length, 25 frames.

2.5 Statistical Analysis

Percentage data were square-root and arcsine transformed before using the Shapiro-Wilks normality test and Levine's test for variance homogeneity. Larva size and average path velocity data were not transformed. Oneway ANOVA and Duncan's multiple comparisons were used to compare differences among experimental groups when equal variances were assumed. The results are expressed as the means±standard deviation (SD).

3 Results

3.1 Seawater Chemistry

Three pH were maintained stable throughout the experiment (Fig.1). Salinity was controlled at 15 ± 0.4 , and temperature was kept at $25\,^{\circ}\mathrm{C}\pm0.3\,^{\circ}\mathrm{C}$. Seawater carbonate chemistry parameters measured and calculated for all treatments were reported in Table 1.

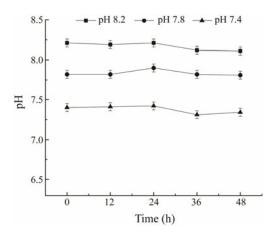


Fig.1 Daily pH (mean±SD) of seawater during a period of 48 h of different pH groups.

3.2 Ocean Acidification on Fertilization and Hatching Rate of *C. sinensis*

Both fertilization and hatching rate of *C. sinensis* were significantly affected by OA. They declined with seawater pH. Averagely, $56.5\% \pm 5.3\%$ eggs were successfully fertilized at pH 8.2, but the percentage was $53.6\% \pm 2.4\%$ at pH 7.8 and $53.2\% \pm 2.8\%$ at pH 7.4, respectively. Similarly, the hatching rate at pH 8.2 was $67.0\% \pm 1.0\%$, the percentage at low pH was $65.7\% \pm 1.2\%$ and $53.3\% \pm 2.5\%$, respectively (Fig.2).

3.3 Ocean Acidification on the Larval Development of *C. sinensis*

At 24h after fertilization, 80.67% of C. sinensis larvae

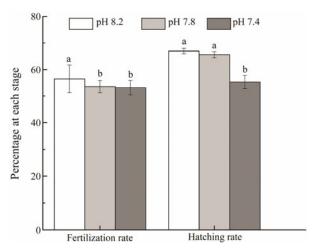


Fig.2 Fertilization and hatching rates of C. sinensis when it experiences OA. The error bars represent the standard deviation. Data are the averages (n=3). Different lower-case means significant differences between fertilization rate or hatching rate at different pH (P < 0.05).

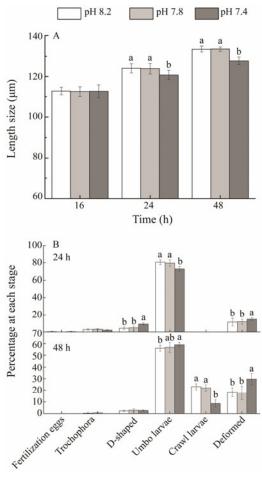


Fig.3 Percentage distribution at each developmental stage of *C. sinensis* embryos/larvae after 24, and 48 h when it experiences OA. The error bars represent the standard deviation. Data are averages (n = 3). Different lowercase means the significant difference between percentages at different pH but the same developmental stage (P < 0.05).

at pH 8.2 and 79.67% at pH 7.8 developed into umbo larvae. However, only 73% completed this process at pH 7.4, which was significantly lower than that of other two



pH. Conversely, deformation percentage at pH 7.4 was significantly higher than those at other two pH. At 48 h post fertilization, 23% larvae at pH 7.4 and 22% at pH 7.8 had entered the crawl larva stage, only 8.67% at pH 7.8 did so. Elevated deformation was found at all pH, and the number of deformations at pH 7.4 was still larger than those at the other two pH (Fig.3).

3.4 Effect of Ocean Acidification on Larval Size of C. sinensis

The size of *C. sinensis* larvae was measured at 16 hpf, 24 hpf and 48 hpf, respectively. There was no significant difference in larval length and height at 16 hpf. The larval length and height at pH 7.4 were significantly smaller than those at pH 7.8 and pH 8.2 since 24 hpf (Fig.4).

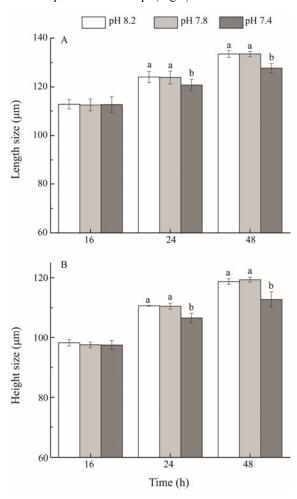


Fig.4 Larvae size of C. sinensis when it experiences OA. The error bars represent the standard deviation. Data are averages (n = 3). Different lowercase means significant difference between larval size at different pH but the same sampling time (P < 0.05).

3.5 Ocean Acidification Effect on Trochophores Average Path Velocity

Average path velocity (VAP) of trochophores at pH 7.4 was significantly depressed but not at pH 7.8 and 8.2. VAP at pH 8.2 reached up to $103\,\mu\text{m}\,\text{s}^{-1}$. At pH 7.8, this was $99\,\mu\text{m}\,\text{s}^{-1}$, and at pH 7.8, VAP was only $62\,\mu\text{m}\,\text{s}^{-1}$ (Fig.5).

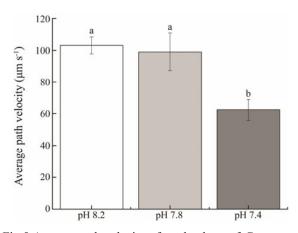


Fig.5 Average path velocity of trochophore of *C. sinensis* when it experiences OA. The error bars represent the standard deviation. Data are averages (n = 3). Different lowercases mean a significant difference between average path velocity of trochophore at different pH (P < 0.05).

4 Discussion

This study provided a scientific evidence for that OA brings a negative effect to calcifying organisms by affecting their development parameters at their early stages, such as fertilization, hatching success, size and average path velocity of larva.

The results of this study on fertilization revealed that decreased seawater pH only significantly inhibited the fertilization of C. sinensis when pH value dropped to 7.4, and moderate acidification brought a slight negative effect to this process. Similarly, Van Colen et al., (2012) found that moderate seawater acidification brought a slight adverse effect on the fertilization of Macoma balthica, but under extreme acidic condition, fertilization was significantly reduced. Shi et al. (2017b) demonstrated that the fertilization increased with decreased seawater pH and attributed this phenomenon to the reductions in the opportunity of sperm-egg collisions by decreasing sperm velocity and the chance of gamete fusion for every gamete collision activity and the intracellular Ca²⁺ oscillation disturbance when the organisms experience OA. However, conclusions about the effect of OA on calcifying organism fertilization were different. Parker et al., (2010) pointed out that OA brought a significant effect to the oyster from Port Stevens (Australia) whereas Kurihara et al. (2007) and Maggs and Samuela (2009) showed that there was no sign of decreased fertilization under OA in C. gigas obtained from the East China Sea and off Western Sweden. This implies that the effect of OA on organism fertilization may be species-specific.

The swimming velocity has been proposed as a larva quality indicator (Myrina et al., 2015). In this experiment, the VAP of trochophora at 8 pfh was depressed at pH 7.4. Similar results were obtained in Ahiura filiformis. Chan et al. (2016) pointed out that the reduced pH brought significant negative effects on swimming velocity of larval brittle stars. However, researches in Dendraster excentricus demonstrated that seawater acidification had no significant effect on larva swimming (Chan et al., 2011).



One possible explanation is that the three species showed various physiological responses to OA, i.e., lowered swimming velocity for C. sinensis and A. filiformis, but stable swimming velocity for D. excentricus. Subsequently, larva body size was measured. At the initial moment, there was no significant difference among different groups. With the development going on, the body size of larvae at pH 7.3 was significantly depressed. The result was consistent with the findings documented previously (van Colen et al., 2012). This may be attributed to energy allocations, e.g., under OA, calcifying organisms need much more energy to maintain acid-base balance (O'Donnell et al., 2010; Stumpp et al., 2012; Lewis et al., 2016). Alternatively, lowered swimming velocity may reduce the chance of larva initiation of feeding (Qiu et al., 2015). As documented in other calcifying organisms, OA delayed C. sinensis larva development and increased the deformed individuals (Wang et al., 2016). The delay in development implied that larvae should take more time to complete the same developmental stage, which might increase the probability of loss by predation in nature (Dupont and Thorndyke, 2009).

The results of our study showed that early development of *C. sinensis* was susceptible to OA. Reduced survival rate and delayed development would likely bring population scale impacts. The predicted OA might reduce the number of *C. sinensis* recruit and lead to a decline in fishery.

Acknowledgements

This work was supported by the Central Public-interest Scientific Institution Basal Research Fund (Nos. 2016HY-ZD0601 and 2018HYXKQ01-10), the Fund of Key Laboratory of Sustainable Development of Marine Fisheries, Ministry of Agriculture, P. R. China, the National Modern Agricultural Industry Technology System Construction Project (No. CARS-49).

References

- Andersen, S., Grefsrud, E. S., and Harboe, T., 2017. Sensitivity towards elevated pCO₂ in great scallop (*Pecten maximus* Lamarck) embryos and fed larvae. *Biogeosciences*, 14: 529-539.
- Byrne, M., 2012. Global change ecotoxicology: Identification of early life history bottlenecks in marine invertebrates, variable species responses and variable experimental approaches. *Marine Environmental Research*, **76** (2): 3-15.
- Cai, W. J., Hu, X., Huang, W. J., Murrell, M. C., Lehrter, J. C., Lohrenz, S. E., Chou, W. C., Zhai, W., Hollibaugh, J. T., Wang, Y., Zhao, P., Guo, X., Gundersen, K., Dai, M., and Gong, G. C., 2011. Acidification of subsurface coastal waters enhanced by eutrophication. *Nature Geoscience*, 4 (11): 766-770.
- Caldeira, K., and Wickett, M. E., 2005. Ocean model predictions of chemistry changes from carbon dioxide emissions to the atmosphere and ocean. *Journal of Geophysical Research: Oceans*, 110: 919-931.
- Chan, K. Y., Grunbaum, D., and O'Donnell, M. J., 2011. Effects of ocean-acidification-induced morphological changes on lar-

- val swimming and feeding. *Journal of Experimental Biology*, **214** (22): 3857-3867.
- Chan, K. Y. K., Grünbaum, D., Arnberg, M., and Dupont, S., 2016. Impacts of ocean acidification on survival, growth, and swimming behaviours differ between larval urchins and brittlestars. *ICES Journal of Marine Science*, 73 (3): 951-961.
- Chou, W. C., Gong, G. C., Hung, C. C., and Wu, Y. H., 2013. Carbonate mineral saturation states in the East China Sea: Present conditions and future scenario. *Biogeoscience Discussion*, 10: 5555-5590.
- Dupont, S., Havenhand, J., Thorndyke, W., Peck, L. S., and Thorndyke, M., 2008. Near-future level of CO₂-driven ocean acidification radically affects larval survival and development in the brittlestar *Ophiothrix fragilis*. *Marine Ecology Progress*, 373 (1): 285-294.
- Dupont, S., and Pörtner, H., 2013. Marine science: Get ready for ocean acidification. *Nature*, 498 (7455): 429.
- Dupont, S., and Thorndyke, M. C., 2009. Impact of CO₂-driven ocean acidification on invertebrates early life-history—What we know, what we need to know and what we can do. *Iop Conference Series Earth and Environmental Science*, **6**: 3109-3131.
- Feely, R. A., Sabine, C. L., Lee, K., Berelson, W., Kleypas, J., Fabry, V. J., and Millero, F. J., 2004. Impact of anthropogenic CO₂ on the CaCO₃ system in the oceans. *Science*, **305** (5682): 362-366.
- Gazeau, F., Gattuso, J. P., Dawber, C., Pronker, A. E., Peene, F., Peene, J., Heip, C. H. R., and Middelburg, J. J., 2010. Effect of ocean acidification on the early life stages of the blue mussel *Mytilus edulis*. *Biogeosciences*, 7 (2): 2051-2060.
- Gazeau, F., Parker, L. M., Comeau, S., Gattuso, J. P., O'Connor,
 W. A., Martin, S., Pörtner, H. O., and Ross, P. M., 2013.
 Impacts of ocean acidification on marine shelled molluscs.
 Marine Biology, 160 (8): 2207-2245.
- Guo, X., Huang, M., Pu, F., You, W., and Ke, C., 2015. Effects of ocean acidification caused by rising CO₂ on the early development of three mollusks. *Aquatic Biology*, **23** (2): 147-157.
- Guo, X., Xu, X., Zhang, P., Huang, M., Luo, X., You, W., and Ke, C., 2016. Early development of undulated surf clam, *Paphia undulate* under elevated *p*CO₂. *Journal Experimental Marine Biology and Ecology*, **484**: 23-30.
- Hofmann, G. E., Barry, J. P., Edmunds, P. J., Gates, R. D., Hutchins, D. A., Klinger, T., and Sewell, M. A., 2010. The effect of ocean acidification on calcifying organisms in marine ecosystems: An organism-to-ecosystem perspective. *Annual Review Ecology Evolution and Systematics*, 41 (1): 127-147.
- Kroeker, K. J., Kordas, R. L., Ryan, C., Hendriks, I. E., Laura, R., Singh, G. S., Duarte, C. M., and Jean-Pierre, G., 2013.
 Impacts of ocean acidification on marine organisms: Quantifying sensitivities and interaction with warming. *Global Change Biology*, 19 (6): 1884-1896.
- Kurihara, H., 2008. Effects of CO₂-driven ocean acidification on the early developmental stages of invertebrates. *Marine Ecology Progress Series*, **373**: 275-284.
- Kurihara, H., Kato, S., and Ishimatsu, A., 2007. Effects of increased seawater *p*CO₂ on early development of the oyster *Crassostrea gigas*. *Aquatic Biology*, **1** (1): 91-98.
- Lewis, C., Ellis, R. P., Vernon, E., Elliot, K., Newbatt, S., and Wilson, R. W., 2016. Ocean acidification increases copper toxicity differentially in two key marine invertebrates with distinct acid-base responses. *Scientific Reports*, 6 (5): 21554.
- Lewis, E., Wallace, D., and Allison, L. J., 1998. Program developed for CO₂ system calculations. Oak Ridge National Laboratory, Oak Ridge, TN, USA.

- Li, H. M., Shi, X. Y., Chen, P., and Zhang, C. S., 2014. Effects of pH and DO on the migration and transformation of phosphate in the process of mixing in the Changjiang Estuary. *Marine Environment Science*, **33**: 497-502.
- Maggs, A., and Samuela, P., 2009. Near-future levels of ocean acidification do not affect sperm motility and fertilization kinetics in the oyster *Crassostrea gigas*. *Biogeosciences Discussions*, **6** (2): 3009-3015.
- Melzner, F., Thomsen, J., Koeve, W., Oschlies, A., Gutowska, M. A., Bange, H. W., Hansen, H. P., and Körtzinger, A., 2013. Future ocean acidification will be amplified by hypoxia in coastal habitats. *Marine Biology*, 160: 1875-1888.
- Millero, F. J., 1995. Thermodynamics of the carbon dioxide system in the oceans. *Geochimica et Cosmochimca Acta*, 59 (4): 661-677.
- Millero, F. J., 2010. Carbonate constants for estuarine waters. *Marine and Freshwater Research*, **61** (2): 139-142.
- Myrina, B., Charlotte, C., Arnaud, H., Ismaël, B., Claudie, Q., Virgile, Q., Caroline, F., and Marc, S., 2015. Assessment of oocyte and trochophore quality in pacific oyster, *Crassostrea* gigas. Aquaculture, 437: 201-207.
- O'Donnell, M., Todgham, A. E., Sewell, M. A., Hammond, L. M., Ruggiero, K., Fangue, N. A., Zippay, M. L., and Hofmann, G. E., 2010. Ocean acidification alters skeletogenesis and gene expression in larval sea urchins. *Marine Ecology Progress Series*, 398 (6): 157-171.
- Parker, L. M., Ross, P. M., and O'Connor, W. A., 2010. Comparing the effect of elevated *p*CO₂ and temperature on the fertilization and early development of two species of oysters. *Marine Biology*, **157** (11): 2435-2452.
- Qiu, T., Liu, Y., Zheng, J., Zhang, T., and Qi, J., 2015. A feeding model of oyster larvae (*Crassostrea angulata*). *Physi*ology Behavior, 147: 169-174.
- Ren, Y., Pan, H., Yang, Y., Pan, B., and Bu, W., 2016. Molecular cloning, characterization and functional analysis of a heat shock protein 70 gene in *Cyclina sinensis*. Fish and

- Shellfish Immunology, 58: 663-668.
- Shi, W., Han, Y., Guo, C., Zhao, X., Liu, S., Su, W., Wang, Y., Zha, S., Chai, X., and Liu, G., 2017a. Ocean acidification hampers sperm-egg collisions, gamete fusion, and generation of Ca²⁺ oscillations of a broadcast spawning bivalve, *Tegillarca granosa*. *Marine Environmental Research*, **130**: 106-112.
- Shi, W., Zhao, X., Han, Y., Guo, C., Liu, S., Su, W., Wang, Y., Zha, S., Chai, X., Fu, W., Yang, H., and Liu, G., 2017b. Effects of reduced pH and elevated pCO₂ on sperm motility and fertilisation success in blood clam, *Tegillarca granosa*. *New Zealand Journal of Marine and Freshwater Research*, **151** (4): 543-554.
- Stumpp, M., Hu, M. Y., Melzner, F., Gutowska, M. A., Dorey, N., Himmerkus, N., and Bleich, M., 2012. Acidified seawater impacts sea urchin larvae pH regulatory systems relevant for calcification. *Proceedings of the National Academy Sciences of the United States of America*, 109 (44): 18192-18197.
- Suquet, M., Mercier, A. L., Rimond, F., Mingant, C., Haffray, P., and Labbe, C., 2012. Setting tools for the early assessment of the quality of thawed pacific oyster *Crassostrea gigas* D-larvae. *Theriogenology*, **78** (2): 462-467.
- Szalaj, D., De Orte, M. R., Goulding, T. A., Medeiros, I. D., DelValls, T. A., and Cesar, A., 2017. The effects of ocean acidification and a carbon dioxide capture and storage leak on the early life stages of the marine mussel *Perna perna* (Linneaus, 1758) and metal bioavailability. *Environmental Science and Pollution Research*, 24 (1): 765-781.
- Van Colen, C., Debusschere, E., Braeckman, U., Van Gansbeke, D., and Vinex, M., 2012. The early life history of the clam *Macoma balthica* in a high CO₂ world. *PLoS One*, 7 (9): e44655.
- Wang, W., Liu, G., Zhang, T., Chen, H., Tang, L., and Mao, X., 2016. Effects of elevated seawater pCO₂ on early development of scallop Argopecten irradias (Lamarck, 1819). Journal of Ocean University of China, 15 (6): 1073-1079.

(Edited by Qiu Yantao)

