

REVIEW PAPER

Improving the use of modelling for projections of climate change impacts on crops and pastures

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Received 21 December 2009; Revised 22 March 2010; Accepted 25 March 2010

Abstract

Projections of climate change impacts on global food supply are largely based on crop and pasture modelling. The consistency of these models with experimental data and their ability to simulate the effects of elevated CO₂ and of increased climate variability has been debated. The effects of high temperatures, of increased climate variability and of several limiting factors which interact with elevated CO₂ such as soil nutrients, pests and weeds are neither fully understood nor well implemented in leading models. Targeted model developments will be required based on experimental data concerning: (i) the role of extreme climatic events, (ii) the interactions between abiotic factors and elevated CO₂, (iii) the genetic variability in plant CO₂ and temperature responses, (iv) the interactions with biotic factors, and (v) the effects on harvest quality. To help make better use of the available knowledge, it is envisioned that future crop and pasture modelling studies will need to use a risk assessment approach by combining an ensemble of greenhouse gas emission (or stabilization) scenarios, of regional climate models and of crop and pasture models, as well as an ensemble of adaptation options concerning both management practices and species/varieties.

Key words: Agriculture, ecophysiology, FACE, global change, livestock.

Introduction

For many key parameters, the climate system is already moving beyond the patterns of natural variability within which our society and economy have developed and thrived. There is a significant risk that many of the trends will accelerate, leading to extreme climatic events and to an increasing risk of abrupt or irreversible climatic shifts (IPCC, 2007a; University of Copenhagen, 2009).

Global climate change can be expected to threaten food supply, for example, through changing patterns of rainfall, increasing incidence of extreme weather, and the changing distribution of diseases and their vectors (Tubiello *et al.*, 2007). Increased incidences of extreme climate events could lead to greater variability of production, increased price volatility and changes in trade flows (Lobell *et al.*, 2008). A consensus has emerged that developing countries are more vulnerable to climate change than developed countries because of the predominance of agriculture in their

economies, the scarcity of capital for adaptation measures, their warmer baseline climates, and their heightened exposure to extreme events (Parry et al., 2001). Thus, climate change may have particularly serious consequences in the developing world (Nelson et al., 2009) where about one billion people are currently undernourished (FAO, 2009).

Even in temperate regions, there are some early warning signs of climate change impacts on the yields of some major crops like wheat. A slower increase in grain yields compared to past decades has been reported in a range of countries, including Europe and India. For example, since 1990, winter wheat yields have been increasing at a significantly slower rate in France than over previous decades (Gate, 2009) and this change has foremost been attributed to an increased variability in climate (Gate 2007, 2009). By contrast, circumstantial evidence for climate-induced increasing grain yield of winter wheat (*Triticum aestivum* L.)

from 1981 to 2005 at two locations in China has been challenged by a recent re-analysis (White, 2009).

The Intergovernmental Panel on Climate Change report (IPCC, 2007b) has reviewed model projections of climate change impacts at the plot level. Moderate warming (i.e. in the first half of this century) may benefit crop and pasture yields in temperate regions, while it would decrease yields in semiarid and tropical regions. Modelling studies indicate small beneficial effects on crop yields in temperate regions corresponding to local mean temperature increases of 1–3 °C and associated CO₂ increases and rainfall changes. By contrast, in tropical regions, models indicate negative yield impacts for the major cereals even with moderate temperature increases (1–2 °C). Further warming has negative impacts in all regions.

These conclusions are based on modelling studies which most often do not account adequately for the impacts of an increased climatic variability (Tubiello *et al.*, 2007; Orlandini *et al.*, 2008). The projected rise in climatic variability will tend to be associated with more extreme weather patterns (IPCC, 2007a), leading to a potential for negative surprises that has not been fully explored, thus reducing the level of confidence in regional and global projections. Moreover, several key interactions are currently poorly described by crop and pasture models including: (i) non-linearity and threshold effects in response to extreme weather events; (ii) modification of weed pest and disease incidence; (iii) field response of crops to elevated CO₂ concentration; and (iv) interactions of climate and management variables with elevated CO₂ (Tubiello *et al.*, 2007).

It is thus imperative to continue to advance the fundamental knowledge of crop and pasture species responses to climate change, reduce uncertainties in impact projections, and assess future risks (IPCC, 2007a). Stakeholders have increasing demands concerning adaptation to climate change of agricultural management (Howden et al., 2007). Adaptation requires detailed understanding of the likely impacts of climate change in small agricultural regions, which further increases the need for modelling.

Some of the key sources of uncertainties for future crop and pasture productivity under climate change are reviewed here; how they could be better reflected in model results, how they could be reduced in the future, and the extent to which conclusions from these models can already be used to provide outlook and guidance for adapting agriculture to climate change are discussed.

An uncertainty cascade in future climate change drivers

Socioeconomic scenarios (SRES) projecting greenhouse gas emissions in CO₂ equivalents are the backbones of impact studies, providing the basis for assessing the impact of climate change on human activities, including agriculture (Parry et al., 2005). Scenarios are neither predictions nor forecasts in a traditional sense; rather they are images of the future, or alternative futures that are meant to assist in

climate change analysis (<u>Nakicenovic</u>, 2000). As an alternative, mitigation scenarios can be used with atmospheric CO₂ concentrations stabilized by 2100 (<u>Arnell et al.</u>, 2002; Tubiello and Fischer, 2007).

Global and regional climate models

The best available tool for predicting climatic change from greenhouse gas (GHG) emission scenarios is the atmosphere-ocean general circulation model (AOGCM) (Giorgi and Mearns, 2002). The atmospheric component of an AOGCM (the atmospheric GCM) is coupled to a land-surface scheme and used to predict changes in landbased state variables (e.g. soil moisture). GCMs compute future climates under anthropogenic forcing (i.e. present and projected future emissions of GHGs (IPCC, 2001a). Their use in impact assessment studies of climate change is widespread (IPCC, 2001b; Reilly et al., 2001). These models provide internally coherent climate dynamics by solving globally all climate-relevant physical equations. Yet it is recognized that GCM projections present significant uncertainties, in part through issues of scale resolution, and in part through imperfect understanding of key climate dynamics (IPCC, 2001a). Even among GCMs with similar temperature-change simulations, predictions of regional precipitation responses may vary significantly, in part because of the intrinsic chaotic nature of climate and in part because of differences in model approach to resolving local to regional atmospheric dynamics.

For example, GCM predictions of the magnitude and location of soil moisture changes often disagree and are often difficult to compare with one another, since the land-surface schemes employed by the various AOGCMs differ considerably, and their accuracy is questionable (Cornwell and Harvey, 2007). This presents a problem when assessing the impacts of climatic change on crops. For example, many studies project impacts using a single GCM (IPCC, 2007a), thus neglecting a major source of uncertainty. Moreover, the spatial resolution of most GCMs is low, which much reduces the realism of local projections of climate change, especially in areas with complex terrain.

Among the methods used to provide insight on climate description at regional scales, regional climate models forced by GCMs have become a widely used tool for regional climate studies (Giorgi and Mearns, 1999; IPCC, 2007a). Since new information is added to the regional climate models by modelling GCM-subgrid processes, the RCMs' fields may have different statistical properties that those of the original GCM (Mearns et al., 1999). This is especially apparent with highly anisotropic fields such as precipitation (Machenhauer et al., 1998).

Downscaling methods

While working with scenarios, difficulties have often been encountered in relation to the disparity of scales between global and regional climate models (RCM) on the one hand and the local application of models on the other hand. For this reason, GCM or RCM results are usually downscaled

with the help of high resolution atmospheric models (dynamical downscaling) and/or statistical models and stochastic weather generators (statistical downscaling). Dynamical downscaling is most demanding in terms of computational resources. Statistical downscaling has the advantage of being computationally fast, but tends to introduce additional uncertainties into the scenarios (Calanca et al., 2008).

The uncertainty originating from the choice of global and regional model formulation in climate-change downscaling experiments has been analysed (Beniston et al., 2007), by conducting a co-ordinated set of climate modelling experiments. While there are a number of contrasted approaches for statistical downscaling (e.g. anomalies, variable corrections, Déqué et al., 2005; statistical disaggregation, Boé et al., 2006), they all require accurate climatic data and long time series and are therefore most applicable in regions with a high density of weather stations. Moreover, it should be noted that the validation of a statistical downscaling algorithm for present day climate conditions does not necessarily imply the validity of its climate change projections, the plausibility of which needs to be compared with direct GCM model outputs (Boé et al., 2006). The task of deriving better estimates of changing variability and extreme events is thus still intrinsically difficult (IPCC, 2007a) and this limits the confidence in climate model projections.

Finally, it should be noted that a number of past impact studies with a focus on agriculture have taken arbitrary ad hoc incremental annual, seasonal or monthly anomalies for the long-term mean temperature and precipitation. Therefore, part of the literature does not reflect the likely range of climate change impacts but only the effects of a certain degree of temperature and/or precipitation change.

In order to assess uncertainties arising from future climatic and atmospheric drivers on crop and pasture productivity better, an ensemble of regional climate models carefully downscaled to reflect better the local distribution of key climatic variables like rainfall is needed (Challinor et al., 2007, 2009). This approach could be generalized by comparing several crop and pasture models for a range of greenhouse gas emission scenarios and/or of mitigation scenarios in order to quantify the probability of avoided impacts brought by greenhouse gas mitigation.

Statistical and mechanistic models of crop and pasture impacts

Assessment studies of the impacts of climate change on agriculture at farm to regional levels need to analyse complex interactions of climate, agro-ecosystem function, and human management. To this end, researchers typically link climate predictions to crop models and land management decision tools (Tubiello et al., 2002). For instance, Decision Support System for Agrotechnology Transfer (DSSAT) (Rosenzweig et al., 1995, Tsuji et al., 1995), Erosion Product Impact Calculator (EPIC) (Williams, 1995), Terrestrial Ecosystem Model (TEM) (Felzer et al.,

2004), and the AEZ model (Tubiello and Fischer, 2007) have been used at the global scale. Simulation models often provide the only available tool to investigate complex interactions and feedbacks of many variables.

Process-based models

Most studies have made use of process-based crop and pasture models to evaluate the impacts of climate change scenarios either at a local or at the global scale (Rosenzweig and Parry, 1994; IFPRI, 2009) for crop and pasture species. However, substantial time, data, and expertise are needed to calibrate these models for particular locations (Wallach et al., 2006) and the lack of process-based models for many minor crops, restricts the scope of this approach. Nevertheless, mechanistic models have recently been expanded in the context of climate change studies to cover an extended range of crop species, including potato (Wolf and Van Oijen, 2003), sugar beet (Launay et al., 2009), cassava, sorghum, millet (Liu et al., 2008), and quinoa (Lebonvallet et al., 2009).

Generic models

An intermediate approach uses detailed agronomic-based knowledge to simulate availability and use of land resources, farm-level management options, and crop production potentials as a function of climate (AEZ, Fischer et al., 2002, 2005). It computes potentially attainable, rather than actual, crop yields. Thus, wherever the gap between actual and potential yields is large, such as in many developing countries, there is uncertainty in translating the calculated impacts of climate change into changes in actual crop productivity. In those regions with large yield gaps, this approach may predict larger positive impacts of climate change on crop production than is actually possible in real fields (Tubiello and Fischer, 2007). It should be noted that results from such approaches, as well as of some of the generic crop modelling (e.g. CROPSYST; Stöckle et al., 2001), can be used to compare the future suitability of land for a range of crops, which allows mapping of the likely future climatic envelopes of the main cropping systems.

Statistical models

Statistical models of crop responses to climate change, based on historical datasets of crop and climate variables have recently been used (Lobell et al., 2008; Paeth et al., 2008; Schlenker and Roberts, 2009) to address climatic change impacts on food security in developing countries. This approach can be used for all crops and all regions, but has several deficiencies (as acknowledged by the authors): (i) errors in statistical data of production and in gridded climate data, and (ii) in some instances, low weight of climatic factors for the between year variation in crop yields. Moreover, statistical models cannot be extrapolated without further assumptions to predict production impacts for future conditions (e.g. higher temperature than any historical year, elevated CO₂ concentration).

Modelling the incidence of extreme climatic events

Increased frequency of heat stress, droughts, and floods (IPCC, 2007a; Beniston et al., 2007) negatively affect crop yields and livestock beyond the impacts of mean climate change, with impacts that are larger, and occurring earlier, than predicted using changes in mean variables alone (IPCC, 2007b).

High temperatures

In Europe, during the summer of 2003, temperatures were up to 6 °C above long-term means, and precipitation deficits up to 300 mm. Crop and pasture yields were reduced by 20-36% in regions affected, leading to uninsured economic losses for the agriculture sector in the European Union which were estimated at 36 billion euros (IPCC, 2007b). It is highly likely (more than 90% chance) that growing season temperatures by the end of the 21st century will exceed even the most extreme seasonal temperatures recorded from 1900 to 2006 for most of the tropics and subtropics (Battisti and Naylor, 2009). High seasonally averaged temperatures will challenge food production in the future, unless major adaptations are made. Indeed, above-optimal temperatures have been reported to induce severe damages to corn and soybean crops in the US, leading to large potential negative impacts in the future (Schlenker and Roberts, 2009).

Yield damaging climate thresholds spanning periods of just a few days for cereals and fruit trees include absolute temperature levels linked to particular developmental stages that condition the formation of reproductive organs, such as seeds and fruits (Wheeler et al., 2000; Wollenberg et al., 2003). Three research needs were identified by Wheeler et al. (2000) in order to provide a framework for predicting the impact of episodes of hot temperatures on the yields of crops: (i) reliable seasonal climate projections, (ii) robust predictions of crop development, and (iii) crop simulation models which are able to quantify the effects of brief episodes of hot temperatures on seed yield. Crop simulation models often do not, at present, capture the effects of brief hot temperature episodes on the number of reproductive structures (Wheeler et al., 2000). Inclusion of detailed phenology and aboveoptimal temperature effects on crops is therefore required to address this problem (Porter and Semenov, 2005).

Variability in temperature can also affect yield quality. Heat damage effects on wheat grains quality caused by episodes of hot temperatures are reasonably well documented (Wheeler et al., 2000), but we are not aware of any crop model that explicitly simulates such effects, despite the increasingly large impacts of these damages on grain price and on bread making. Moreover, information on the effects of variability in temperature on the yield and quality of other annual and perennial crops is sparse, which prevents the development of models.

Drought

During the 20th century a major drought index has increased over a number of regions, including Southern Europe and most of Africa (Bates et al., 2008). Further drying of a large part of the subtropics is likely by the end of this century (IPCC, 2007a) as a result of elevated greenhouse gas concentrations. Amplification of the hydrological cycle as a consequence of global warming is forecast to lead to more extreme intra-annual precipitation regimes characterized by larger rainfall events and longer intervals between events (IPCC, 2007a). More extreme rainfall regimes are expected to increase the duration and severity of soil water stress in mesic ecosystems as intervals between rainfall events increase (Knapp et al., 2008). Drought in particular plays an important role in pasture dynamics and can lead to longterm degradation (Briske et al., 2005).

Plant strategies to cope with drought normally involve a mixture of stress avoidance and tolerance 'strategies' that vary with species and with genotype. For instance, to survive severe droughts, most perennial pasture species exhibit a combination of responses. An increased senescence of mature shoots leads to a reduction in leaf transpiration surface. An increase in rooting depth and water uptake at low soil water potential (Durand et al., 2007) contributes to a delay in dehydration. The accumulation of carbohydrates and dehydrins in basal tissues, including buds and meristems that ensure perenniality, allows tolerating dehydration in the surviving organs (Volaire, 2002). The genotypes surviving best in the most arid conditions exhibit responses associated with summer dormancy (Norton et al., 2008, 2009). The combination of these plant strategies confers very contrasting inter- and intra-specific adaptations to drought. An example of threshold effect is the abrupt cessation of the ascending sap flow, which occurs as a result of xylem cavitation. With herbaceous plants this process has been evidenced in maize, rice, sunflower, and sugar-cane (Tyree et al., 1986; Cochard et al., 2002).

Moreover, drought has a number of interactive effects on plant-soil processes. For example, it sharply reduces microbial biomass in soils, leading to the transient accumulation of soluble forms of N and to high N uptake when rainfall resumes (Lemaire and Denoix, 1987). Longer-term effects will be affected by the increase in the atmospheric CO₂ concentration (Loiseau and Soussana, 2000; Pinay et al., 2007) and may include shifts in functional microbial types (e.g. nitrifiers, denitrifiers, heterotrophs) (Barnard et al., 2006), increased nutrients losses and, hence, changes in biogeochemical cycles and carbon sequestration.

Furthermore, a number of potential feedbacks that may increase water stress and vulnerability have not yet been simulated at the scale of a river basin. For example, a decline in vegetation cover and soil organic matter as a result of warming and drought would increase run-off and reduce soil water storage, thereby accentuating waterrelated extremes. Contrasted conclusions may be obtained with crop models depending on the assumptions made, or not, on changes in capillary rise and in groundwater level which may provide a substantial part of evapotranspiration, especially during low precipitation periods. Scaling up individual processes such as run-off, infiltration, soil water redistribution, is uneasy since the boundaries of the system are at a higher scale than the field (Schulze, 2000). Process-oriented ecosystem and crop simulation models do not adequately represent the rise in plant mortality that may occur as a result of climatic variables exceeding thresholds and the resulting complex plant-soil interactions described above.

The ability of current crop and pasture models to simulate future summer heat and drought events could be tested by using severe heat and drought pulses like the summer 2003 event in Europe as a benchmark. Further benchmarks could be obtained through the development of crop and ecosystem manipulation experiments imposing extreme climatic events of the same magnitude as that predicted by downscaled climate scenarios.

Modelling interactions between elevated CO₂ and abiotic factors

Crop and pasture models used in impact assessment studies simulate the effects of elevated CO₂ on growth and yield with a variety of methods (reviewed by Tubiello and Ewert, 2002), ranging from simple multipliers of final biomass and yield or variable multipliers of daily growth, based on, for example, water and nutrient stress, phenological stage, leaf area feedback, etc. Only a few models have mechanistic equations of leaf photosynthesis scaled to canopy. For instance, the AEZ-simulated crop response to elevated CO₂ is modelled rather simply, as a multiplier of the harvest yield obtained under current CO2 levels (Tubiello and Fischer, 2007). In the AEZ assessment, these factors are applied to land units without production limitations. Where suitability is limited, the multipliers are reduced proportionally to the magnitude of the limitations. While this type of approach is generic and robust it does not necessarily cover the main interactions between elevated CO2 and environmental drivers which have been demonstrated experimentally.

Modelled versus observed effects of elevated CO₂

Some concerns have also been expressed (Long et al., 2006; Ainsworth et al., 2008b) that most simulation models would overestimate the CO₂ effect compared with Free Air Carbon dioxide Enrichment (FACE) experiments.

Tubiello et al. (2007) reported that some key models used in climate change impact assessments have not been evaluated against FACE data, but where this has been carried out, the models reproduce FACE results well. This statement was based on a previous assessment of five widely used crop models with wheat grain yield data from the Maricopa FACE experiment (Tubiello and Ewert, 2002). However, this conclusion was challenged by Long et al. (2006) and Ainsworth et al. (2008b) who showed experimental mean response ratios of 1.08 and 1.18 under wellwatered and water-stress conditions, respectively, whereas the average model outputs estimated response ratios of 1.18 and 1.28, respectively. Thus, wheat crop models which were compared to the Maricopa FACE data would overestimate the CO₂ fertilization effect both under well-watered conditions and under water-stress conditions (Ainsworth et al., 2008a).

There is, however, no consensus yet in the literature on the experimental reference to be used for model calibration. By scaling the relative yields to a ratio of 700 to 370 µmol mol⁻¹ CO₂, Ziska and Bunce (2007) have shown that relative yield stimulations in response to future CO₂ concentrations obtained using a number of enclosure methodologies are quantitatively consistent with FACE results for three crops of global importance: rice, soybean, and wheat. By contrast, a direct approach was taken by Ainsworth et al. (2008a) who limited the comparison of FACE experiments and chamber studies to those with similar ambient [CO₂] and similar elevated [CO₂]. It should be stressed, however, that ambient CO₂ concentration itself has changed appreciably during the experimental era (e.g. from 320 in early enhancement studies to a current concentration of 380 µmol mol⁻¹). This could cause a higher response ratio in earlier studies conducted with enclosures than in more recent studies with FACE technology.

In any case, the best test of model parameterization and model design is validation of model output against observed experimental data comparisons. Therefore, a greater number of comparisons with elevated CO₂ and warming experiments would be valuable since the magnitude of the CO₂ effect may vary among sites, according to crop types, as well as to environment and management drivers.

Interactions between elevated CO₂ and temperature

As an increase in temperature enhances photorespiration in C₃-species (Long, 1991), the positive effects of atmospheric CO₂ enrichment on photosynthetic productivity are usually greater when temperature rises close to optimal values. However, the CO₂ effect for seed yield would be lower at above-optimal temperatures (Matsui et al., 1997; Batts et al., 1998; Amthor, 2001). For rice, increasing CO₂ and temperature may, in fact, negate any yield enhancement related to increasing CO₂ (Ziska et al., 1996; Moya et al., 1998) as a result, in part, of reductions in transpirational cooling, higher canopy temperatures and increased pollen sterility (reviewed by Ziska and Bunce, 2007). This decline in the magnitude of the CO₂ effect at high temperatures is not necessarily well captured by current simulation crop models and would require more research, since aboveoptimal temperatures may become prevalent by the end of the century for a number of crops and of regions (Lobell et al., 2008).

Interactions between elevated CO2 and nutrients

Over a number of studies it has been found that plants grown in conditions of high nutrient supply respond more strongly to elevated atmospheric CO₂ concentrations than nutrient-stressed plants (Poorter, 1998; Schneider et al., 2004; Soussana and Lüscher, 2007). FACE experiments confirm that high soil N contents increase the relative

response to elevated atmospheric CO₂ concentrations (Nowak et al., 2004). With a temperate pasture grass, in the Swiss FACE experiment, the response in above-ground dry matter (DM) yield to elevated atmospheric CO₂ concentrations increased from being not significant to a significant positive proportional response of 0.17 when the rate of application of N fertilizer was increased (Schneider et al., 2004). The CO₂-induced N limitation was alleviated in the high N-fertilizer treatment only by supplying a significant external input of N. These results confirm that N is a major limiting factor in the response of grasslands to elevated atmospheric CO₂ concentrations.

As a result of these interactions with soil processes, experiments which impose sudden changes in temperature or atmospheric CO₂ concentrations, and which last only a few years are unlikely to predict the magnitude of the long-term responses in crop productivity, soil nutrients (Thornley and Cannell, 2000) and C sequestration. This may imply that the actual impact of elevated atmospheric CO₂ concentrations on yields in farmers' fields could be less than earlier estimates which did not take into account limitations in availability of nutrients and plant—soil interactions.

When other nutrients are not strongly limiting, a decline in N availability may be prevented by an increase in biological N₂-fixation under elevated atmospheric CO₂ concentrations (Gifford, 1994). Indeed, in fertile grasslands, legumes benefit more from elevated atmospheric CO₂ concentrations than non-fixing species (Hebeisen *et al.*, 1997; Lüscher *et al.*, 1998) resulting in significant increases in symbiotic N₂ fixation (Soussana and Hartwig, 1996; Zanetti *et al.*, 1997). Nevertheless, other nutrients, such as phosphorus, may act as the main limiting factor restricting growth and responses in yield in legumes to atmospheric CO₂ concentrations. This has been demonstrated both in calcareous grasslands (Stöcklin *et al.*, 1998) and under controlled environmental conditions (Almeida *et al.*, 2000).

Such experimental results underline the need to model nutrients cycling under climate change and not just plant growth. This can be best achieved by coupling crop/pasture models with soil models (Thornley and Cannell, 2000). However, simpler approaches have been used in some models like AEZ where multipliers of yield are reduced proportionally to the magnitude of the limitations. This conclusion is extended by the model to interactions between CO₂, nitrogen, and other major nutrients (P, K, and S) which are widespread limiting factors especially in tropical crops and pastures.

Interactions between elevated CO2 and water

Most crop and pasture models (Tubiello and Ewert, 2002) include an interaction between elevated atmospheric CO₂ and water availability. An increased productivity from increased water-use efficiency is the major response to elevated atmospheric CO₂ concentrations in C₃- or C₄-crops that are exposed frequently to water stress (Casella and Soussana, 1997; Drake *et al.*, 1997; Aranjuelo *et al.*, 2005). Moreover, elevated atmospheric CO₂ concentrations can

reduce depletion in soil moisture content (Morgan *et al.*, 2004). These results support a general view that elevated atmospheric CO₂ concentration reduces the sensitivity to low precipitation in grassland ecosystems (Volk *et al.*, 2000; Morgan *et al.*, 2004). However, it is still unclear whether, or not, elevated CO₂ can reduce mortality and increase recovery during severe water stress events and further research will be needed in order to model the impacts of such episodes on annual crops and perennial vegetation better.

Interactions between elevated CO₂ and atmospheric O₃ concentration

Another uncertainty is the role of air pollution and especially of increased ozone exposure. IPCC emission scenarios for many cropland regions project elevated mean ozone levels in surface air for 2050 and beyond. At the same time, crop sensitivity may decline in areas where warming is accompanied by drying (e.g. in southern and central Europe). Besides uncertainties in climate projections, parameters in models for ozone risk assessment are also uncertain and model improvements are necessary to define specific targets for crop improvements to identify regions most at risk from ozone in a future climate and to set robust effect-based ozone standards (Fuhrer et al., 2009).

Semi-empirical versus mechanistic models of photosynthesis

The complexity of crop ecophysiology under elevated CO₂, the large range of interactions with abiotic factors, and the possibility for substantial between (Kimball *et al.*, 2002; Nowak *et al.*, 2004) and within (Ziska and George, 2004) species variation in plant responses clearly challenge crop and pasture modellers.

Although, semi-empirical models may seem robust and easily applicable, they tend to oversimplify interactions between elevated CO₂ and abiotic factors which may create concerns for their accuracy. On the other hand, mechanistic models based on physical and biochemical laws are theoretically valid for any combination of soil, climate and crop conditions but are difficult to parameterize. For instance, responses of leaf photosynthesis to variations of light, temperature, and CO2 concentration have been successfully represented by the biochemical model of C₃ photosynthesis by Farquhar et al. (1980), which has pioneered the mechanistic representation of the dark (Wc) and light (Wj) biochemical processes. It assumes that the carboxylation/oxygenation of RuBP by the enzyme Rubisco, and the regeneration of RuBP by the electron transport chain are the main limitations of photosynthesis. Possibly because of the large effort required for parameterizing such detailed photosynthetic models, we are not aware of any crop model being currently used for climate change impact simulation that incorporates the equations for leaf photosynthesis by Farquhar et al. (1980) and upscales these equations from leaf to canopy level. Nevertheless, some

comparisons between simplified photosynthetic equations with Farguhar's equations have sometimes been performed successfully (e.g. for the CROPGRO model; Alagarswamy et al., 2006), but for a limited range of environmental conditions. Incorporating detailed ecophysiological knowledge in widely used crop models would help in reducing the probability for surprises caused by possible failures of semiempirical models of photosynthesis outside the environmental domain for which they were calibrated.

Modelling biotic interactions under elevated CO₂ and climate change

Growth-reducing factors reduce or hamper crop growth and comprise biotic factors such as weeds, pests, and diseases (Van Ittersum et al., 2003). Biotic yield-reducing factors such as insects, pathogens, and weeds have long been ignored in climate change impact simulations (Coakley et al., 1999).

Pests and diseases

CO₂-temperature interactions are recognized as a key factor determining plant damage from pests in future decades; CO₂-precipitation interactions will also be important (Stacey and Fellows, 2002; Zvereva and Kozlov, 2006). Most studies continue to investigate pest damage as a separate function of either CO₂ (Agrell et al., 2004; Chen et al., 2005) or climate, mostly temperature (Bale et al., 2002; Salinari et al., 2006). Importantly, increased climate extremes may promote plant disease and pest outbreaks (Alig et al., 2002; Gan et al., 2004) and such interactions are not yet included in modelling approaches.

The development of linked disease-crop models is still an important objective within the overall goal of developing a predictive capability for agricultural impact assessment (IPCC, 2007a). In recent years only, models of plant disease have been developed to incorporate more sophisticated climate predictions. Several approaches have been used for modelling the effect of climate change on disease. Climate matching is applied by quantifying the climatic features of locations so that the success of an organism in a reference climate can be used to predict the success of that organism in other locations with similar climates. Empirical models, such as regression models with climatic variables as predictors and epidemic parameters as response variables, can be used to predict the success of organisms across the range of conditions studied (Garrett et al., 2006). Simulation models are based on theoretical relationships and can predict outcomes under a range of scenarios. However, there are continuing problems with the application of models for predicting climate change effects on disease, including the lack of data on the geographic distribution of disease, non-linear relationships and thresholds in the relationship between climatic variables and epidemiological responses, and the potential for adaptation by plants and pathogens that is often ignored in models (Scherm, 2004).

Weeds

Increases in high temperatures (above 32 °C) have been reported to enhance weed competitiveness in soybean crops (Tungate et al., 2007). Recent research has also highlighted the key role of competition between C₃ crop and C₄ weed species under different climate and CO₂ concentrations. Ziska and Bunce (2006) reviewed the role of weeds under elevated CO₂, showing only a handful of weed/crop competition studies with respect to soybean (Ziska, 2000; Ziska and Goins, 2006), one study with respect to rice (Alberto et al., 1996) and no studies with respect to wheat, where the effects of projected changes in CO2/climate on seed yield have been quantified. To our knowledge, there has been no research specifically to examine the potential for agricultural weeds to adapt rapidly to climate change and elevated CO₂. Phenological changes in agricultural weeds could significantly alter crop-weed interactions and recent work by Franks and Weis (2007) has shown the potential for rapid life-history evolution in response to climate change in annual weedy plants. Demo-genetic models that incorporate demographic and environmental stochasticity with quantitative genetics may have utility for predicting population level responses of weed species under changing management and climatic conditions (Neve et al., 2010).

Multi-species grasslands and intercropping

Mechanistic bases for establishing the differential responses of plant species to elevated CO₂ have been thought of. However, the relative growth responses to elevated atmospheric CO₂ concentrations obtained for isolated plants cannot be used to predict the response in multi-species mixtures although the CO₂ response of monocultures may be a better predictor (Poorter and Navas, 2003). Studies of between-species competition among three grasses, showed that grasses that capture relatively more light per unit leaf area in mixtures than their competitors become increasingly dominant under elevated atmospheric CO₂ concentrations (Teyssonneyre et al., 2002b). Moreover, a high N-use efficiency can confer a competitive advantage under elevated atmospheric CO2 concentrations to mixed grasses (Soussana et al., 2005). Such experiments suggest that the inclusion in models of mechanisms of resource capture and use among plant functional types may help simulate plant species dynamics in multi-species grasslands under elevated CO₂ (Soussana and Oliveira Machado, 2000; Lazzarotto et al., 2009) and climate change. In the same way, intercrop models (Corre-Hellou et al., 2009) could be applied to future climatic conditions.

Modelling adaptation to climate change

Adaptation to climate change will require changes in genotype (G) and management (M) to match the climateinduced changes in environment (E). For rainfed crops in dry areas, evidence shows that variation due to E far outweighs the variation of grain yield that is due to M or

G, or the interactions between these factors, and between these factors and E (Anderson, 2010). As shown by a review of modelling studies, there is a tendency for most of the benefits of adapting the existing systems to be gained under moderate warming (<2 °C) then to level off with increasing temperature changes (Howden et al., 2007). In addition, the yield benefits tend to be greater under scenarios of increased than decreased rainfall, reflecting that there are many ways of more effectively using more abundant resources, whereas there are fewer and less-effective options for significantly ameliorating risks when conditions become more limiting (Howden et al., 2007).

Most field-scale crop and pasture simulation models can test the effects of changes in farming practices, such as changes in planting and harvest dates, in the timing and amount of fertilizer and/or water supply. By contrast, there are only few examples of modelling studies (but see Challinor et al., 2007), which have tried to alter varieties/ species to those with more appropriate thermal time and vernalization requirements and/or with increased resistance to heat shock and drought to maintain crop productivity and quality consistent with the prevailing climate. In the same way, improving the effectiveness of pest, disease, and weed management practices through wider use of integrated pest and pathogen management, development, and use of varieties and species resistant to pests and diseases are effective adaptation measures (Howden et al., 2007) which, however, have not yet been tested in the context of climate change modelling studies.

Climate change alters breeding targets. Increasing emphasis needs to be placed on the identification of the most CO₂-responsive genotypes (Ainsworth *et al.*, 2008*a*) and of drought- and salinity-tolerant genotypes (Tester and Langridge, 2010) in order to provide starting lines for breeding programmes. Modelling can be used to facilitate plant breeding by dissecting and integrating the role of traits in an altered climate (Hammer *et al.*, 2002).

Conclusions

There is no single approach to the complex task of modelling the impacts of climate change on agriculture and food security. In climate sciences, with the development of computer capacities, simpler models have not disappeared; on the contrary, a stronger emphasis has been given to the concept of a 'hierarchy of models' as the only way to provide a linkage between theoretical understanding and the complexity of realistic models (Held, 2005). In the same way, a hierarchy of models ranging from simple 'theoretical' models to complex models is useful for crop and pasture modelling, which will have a continuing central, heuristic role to support scientific investigation of climate change impacts on agriculture and to facilitate adaptative decision making by farmers and land managers.

To approach the vulnerability to climate change of crops and pastures better, we envision that future modelling studies will need to develop a risk assessment approach by combining ensembles of:

- downscaled regional climate models and GHG emissions (or stabilization) scenarios,
- improved crop models tested against available experimental evidence (e.g. in FACE studies) in a large range of climatic conditions (including extreme climatic events),
- adaptation options, including both changes in management practices and in varieties/species.

Moreover, uncertainties in projections will need to be better assessed, separating climate projection-related, crop model-related, and adaptation-related uncertainties. While such approaches would help make better use of the available knowledge in simulation studies, there is also a need to reduce epistemic uncertainties by further developing research to reduce some of the knowledge gaps in terms of agronomy, ecology, and ecophysiology which have been highlighted in this review. Compared with past studies, targeted model developments will be required based on experimental data designed to reduce uncertainties in areas such as: (i) the role of extreme climatic events, (ii) the interactions between abiotic factors and elevated CO₂, (iii) the genetic variability in CO₂ and temperature responses, (iv) the interactions with biotic factors, and (v) the effects on harvest quality.

Acknowledgements

AI Graux acknowledges financial support for her PhD thesis from the Environment project of the Auvergne region. This review benefited from scientific discussions within the 'Climator', 'Validate' and 'Adage' projects funded by the French ANR (Agence Nationale de la Recherche).

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