- Influence of boundary layer dynamics and isoprene
- ² chemistry on the organic aerosol budget in a tropical
- 3 forest
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- 4 Abstract. We study the organic aerosol (OA) budget in a tropical for-
- ₅ est by analyzing a case that is representative for the OP3 campaign at Bor-
- 6 neo. A model is designed that aims for a consistent representation of the chem-
- ₇ ical and meteorological processes that drive the diurnal evolution of reac-
- tants in the atmospheric boundary layer (BL). The model is able to repro-
- 9 duce the observed diurnal dynamics of the BL, including the evolution of most
- chemical species involved in secondary organic aerosol (SOA) formation. A
- budget analysis reveals a clear signal of the entrainment process in the di-
- urnal evolution of SOA. Further, we perform a series of sensitivity analyses
- to determine the effect of meteorological forcings and isoprene chemical path-
- ways on the OA budget. Subsidence and advection of cool air have oppos-
- ing effects on the OA concentration, although both suppress BL growth. Re-
- cycling of the OH radical in the oxidation of isoprene may affect the amount
- of SOA that is formed, but must be understood better before its impact can
- ₁₈ be definitely determined. SOA formation from isoprene is calculated for both
- 19 the low- and high-NO_x pathway, with the latter dominating the isoprene per-
- ₂₀ oxy radical chemistry. In a final analysis, we study the significance of SOA
- formation through the reactive uptake of isoprene epoxydiols (IEPOX) on
- ²² acidic sulfate aerosol. Despite the incorporation of these new pathways, the
- OA concentration is systematically underestimated by about a factor of 2.

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1. Introduction

Tropical forests are potentially an important source of biogenic secondary organic aerosol (SOA), due to high emissions of isoprene and terpenes [Langford et al., 2010; Karl et al., 2007 and potentially high concentrations of their most important oxidant, the hydroxyl radical (OH) [Lelieveld et al., 2008]. Recently, a number of measurement 27 campaigns have been conducted to gain insight in the sources and formation mechanisms of SOA in forests in Amazonia [AMAZE, Chen et al., 2009], West-Africa [AMMA, Capes et al., 2009] and South-East Asia [OP3, Robinson et al., 2011a, b, 2012]. When interpreting observations made in the atmospheric boundary layer (BL) during these campaigns, 31 it is important to realize that the evolution of chemical species in the BL is a func-32 tion of chemical conversion, emission/deposition, advection and the vertical exchange of 33 compounds between the free troposphere (FT) and the BL driven by entrainment [Vilà-Guerau de Arellano et al., 2009; Ouwersloot et al., 2012] and gas/particle partitioning in case of SOA. In addition, subsidence and advection of heat and moisture influence the growth of the BL and therefore modify its dilution capacity and the exchange of species between the BL and FT, as controlled by entrainment [Ouwersloot et al., 2012]. Here, we investigate the diurnal budget of OA by combining a model with observations from the OP3 campaign.

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- There is still a considerable gap between the understanding of ambient biogenic SOA
- and the ability of models to reproduce its observed concentration. Simulations by both

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Capes et al. [2009] and Chen et al. [2009] underpredicted SOA concentrations in isoprenedominated tropical environments but these studies could not rule out the possibility of canceling errors, due to the limited observational constraints on these estimates. These sources of uncertainty were the identification and emission rates of biogenic SOA precursors, SOA formation mechanisms, oxidant concentrations, the SOA particle mass yields, the influence of vertical mixing and advection, and the unknown contribution of the background OA concentration (OA_{BG}). On the other hand, Slowik et al. [2010] were able to reproduce observed OA concentrations in a rural environment dominated by terpene emissions. Sjostedt et al. [2011], however, underestimated OA compared to the measurements in an environment where isoprene was more abundant, both using the same model as Slowik et al. [2010] and an approach based on VOC destruction rates.

The underestimation of modeled OA in environments with high isoprene emissions may

be due to a lack of understanding of the chemical pathways that lead to the formation of

SOA from isoprene (ISOA). Current parameterizations of ISOA formation are therefore

subject to large uncertainties [Carlton et al., 2009]. More specifically, Surratt et al. [2010]

suggested that peroxy methacryloyl nitrate (MPAN) and isoprene epoxydiols (IEPOX)

are important intermediate gas-phase species under high- and low-NO_x conditions, re
spectively. The mechanisms of SOA formation through these reaction pathways have only

recently started to be explored [Paulot et al., 2009; Surratt et al., 2010; Chan et al., 2010;

Lin et al., 2012b; Kjaergaard et al., 2012].

Another issue related to isoprene chemistry in tropical forests, is the mismatch be-

tween measurements and theoretical calculations of OH concentrations, which has been

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attributed to regeneration of OH in the oxidation of isoprene [Lelieveld et al., 2008]. Several mechanistic pathways leading to OH recycling have been proposed, but none of them
has been able to explain the gap between measurements and models [Stone et al., 2011].
Taraborrelli et al. [2012] recently formulated a detailed mechanism for isoprene oxidation
by assembling and completing several previously proposed mechanisms, with which they
were able to reproduce OH concentrations under pristine tropical conditions to within the
bounds of the measurement uncertainty. Another point of view was presented by Mao
et al. [2012], who suggested that the high observed OH concentrations may be the result of
a measurement artifact. Their findings, however, are very much dependent on instrument
design and environmental conditions. The effect of OH recycling on SOA formation has
so far been subject of only one study [Lin et al., 2012a] and it is still unclear how the
formation of ISOA forming products depends on OH concentrations [Carlton et al., 2009].

The OP3 campaign provides a challenging case since the aerosol at Borneo is influenced by complex terrain and multiple sources. The measurement tower is located on top of a 200 m hill, which means that it could be influenced by anabatic flows, which are upslope flows driven by heating of the slope by insolation [e.g. Thunis and Bornstein, 1996]. The proximity of the coast means that the site may be influenced by sea breeze circulations [Robinson et al., 2012] and heterogeneities at smaller spatial scales can induce advection of heat, moisture and chemical species [Ouwersloot et al., 2011]. Moreover, the aerosol has multiple sources: Robinson et al. [2012] investigated the effect of the island on the vertical distribution of aerosol through the troposphere, based on aircraft observations

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and showed that air is enriched in OA as it passes over the island, which indicates a large

on-island source of biogenic SOA. An analysis of air mass back trajectories for the site

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based on ECMWF wind fields, however, revealed that during the campaign there was no period during which the rain forest was the only source of aerosol and that significant levels of (off-island) sulfate aerosol were transported to the site [Robinson et al., 2011b]. This is in contrast with the Amazon, which experiences periods with predominant in-basin influences during which biogenic SOA dominates the aerosol mass [Chen et al., 2009].

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Extending on the work of Robinson et al. [2011b, 2012], we focus on the interpretation of ground-based OA measurements made during OP3 and how various dynamic and chemical terms contribute to the organic aerosol budget. A schematic overview of these factors is presented in Fig. 1. We aim for an integrated approach by simultaneously accounting for atmospheric boundary layer processes, as influenced by local surface and large-scale 100 meteorological forcings, and for chemical processes, related to both gas-phase and sec-101 ondary aerosol chemistry and partitioning. In our study we overcome some of the issues in previous studies of the OA budget in tropical forests by prescribing VOC emissions as constrained by above-canopy flux measurements, by accounting for entrainment and by performing an experiment in which OH concentrations are matched with observations 105 by including OH regeneration in isoprene oxidation in the model. Since this case is well-106 constrained with data it gives the opportunity to assess which processes contribute most 107 to these large uncertainties and we shed some light on some other factors that are not that 108 well-constrained and are yet uncertain or difficult to estimate, for example subsidence. 109 To this end, we use MXLCH-SOA, a 0-D model that reproduces the essentials of the 110 dynamics of the convective boundary layer, the gas-phase chemistry leading to the for-111

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mation of semi-volatile organics and the gas/particle partitioning of these organic species

[Janssen et al., 2012]. In the design of the model, we have kept a balance between the level 113 of complexity of the representations of the different components, and its ability to repro-114 duce the observations of dynamics and chemistry. In this way, MXLCH-SOA allows us to 115 break down the budgets of OA and its precursors into the various dynamical and chemical 116 terms that contribute to them. The model is updated to include SOA from isoprene using 117 the volatility basis set (VBS) approach. After examining the complete data set, we select 118 a representative case study, discuss the performance of the model when compared to the 119 case study observations and show the contributions of several factors to the budget of 120 OA and its precursors. Then, we analyze the impact of large-scale meteorological forcings 121 on the OA concentration, discuss the potential role of OH recycling in SOA formation, 122 and analyze the formation of SOA from IEPOX. The latter has been suggested to yield 123 a specific tracer (hereinafter called 82Fac) [Lin et al., 2012b] that is present in the OA observed by an aerosol mass spectrometer (AMS) at Borneo [Robinson et al., 2011a].

2. Methods

2.1. Observations of the diurnal variability during OP3

Data gathered during the OP3 campaign enable us to study the diurnal variation of 126 atmospheric compounds modulated by surface and BL processes. OP3 was conducted in 127 Malaysia in 2008 [Hewitt et al., 2010] and here we use data from OP3 III (23 June to 20 128 July) at the Bukit Atur Global Atmospheric Watch station, located in the Danum Valley 129 rain forest conservation area in Sabah, Borneo (4°58′ N, 117°50′ E, 426 m a.s.l.). The 130 site was located in a clearing at the top of a 200 m hill, above most of the surrounding 131 forest. The observations were made from a measurement tower of 100 m on top of this 132 hill and measurement heights in this paper are indicated as height relative to the base of 133 the tower. Figure 1 shows a schematic overview of the measurement setting and processes 134 that potentially influence the formation and evolution of local OA. Note that we only study processes that are occurring in the well-mixed atmospheric boundary layer, i.e. the layer above the canopy. The measurements of both fluxes and concentrations that we use to constrain our model are all taken above the canopy layer.

Figure 2 shows mean diurnal cycles during OP3 III, based on half-hourly averages over a period of 4 weeks. It includes the most representative dynamic, surface and chemistry variables, as represented by potential temperature (θ), isoprene flux (F_{ISO}), isoprene concentration (ISO) and concentration of OOA2, an oxidized organic aerosol (OOA) factor, respectively. θ rises during the day due to the sensible heat flux and entrainment of warm air, F_{ISO} follows a diurnal cycle driven by temperature and light intensity [Langford et al., 2010], ISO follows F_{ISO} , but is also modified by chemical transformations and BL dynamics, and the full complexity of the behavior of OOA2 is under study here. We selected one representative day for which we initialize and evaluate the diurnal evolution of chemistry and dynamics of our model with observations: 7 July 2008. To determine the representativeness of this specific day for a typical day at the measurement site during OP3 III, we compared observations from this day to the campaign mean (Fig. 2). Similar diurnal trends are present in the data of the case study and in the campaign averages. Additionally, the observations from the case study fall within the standard deviation of these averages, except for OOA2 in the afternoon.

Furthermore, the selection of this day is based on the relatively smooth evolution of 154 the surface heat fluxes during this day, which ensures convection and turbulent mixing 155 occurring throughout the BL. Additionally, the observed diurnal cycle of the ozone pho-156 tolysis rate jO¹D followed the theoretical clear sky diurnal evolution relatively smoothly 157 compared to other days during the campaign, although there were some fluctuations, probably caused by clouds. These conditions were valid until \sim 14:00 LT. After 14:00, temperature dropped drastically and also moisture and concentrations of chemical species 160 suddenly changed. Possible explanations for such behavior are the formation of clouds or the arrival of the sea breeze at the site. To avoid the complex transport and chemistry associated with the presence of clouds on top of the BL, we finalize our analysis at 163 14:00 and focus on the period with strong boundary layer growth and OA formation. The validity of the assumption of a well-mixed layer during this day is further supported by 165 vertical profiles of O₃ and NO_x, which were obtained by measuring these species at several 166 heights between 5 and 75 m (not shown). Especially from 45 m upwards, O₃ and NO₂ 167 measurements are very similar at different heights, indicating that above this height we 168 are probing mixing ratios within the BL. On the other hand, this means that observations 169

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below 45 m could be in the surface layer and may therefore deviate from mixed-layer values.

Finally, an important reason for the selection of this day is the availability of the most complete data set of dynamics, gas-phase chemistry (most importantly VOCs and oxidants) and OA. Upper air observations of OA concentrations are potentially very useful for understanding the evolution of OA in the BL [Janssen et al., 2012]. While several vertical profiles of OA over the measurement site have been obtained, unfortunately no flight was carried out on this particular day [Robinson et al., 2012]. Therefore, we use observations from other days to get an estimate of FT OA concentrations.

On 7 July 2008, the measurement site was influenced by air masses arriving from the 179 South-East (Fig. 3), which means that the air masses were affected by substantial amounts 180 of both off-island and on-island emissions. Consequently they contained sulfate aerosol 181 from off-island sources and were affected by isoprene emissions from oil palm plantations located to the South East of the observational site (see Hewitt et al. [2010] for a detailed land-use map). The air masses traveled about 7 h over land before arriving at the measurement site. The pressure altitude during the period over land indicates that the air masses were close to the surface and within the BL. An addition meteorological factor important for our research is the presence of subsiding air motions. The site and its surroundings were influenced by subsidence, as can be inferred from ECMWF reanalysis 188 fields of vertical velocity at the 850 hPa level (Fig. 3). The downward movement of air 189 with 0.2 Pa s⁻¹ over Bukit Atur corresponds with a vertical velocity of -2 cm s⁻¹.

2.2. Description of MXLCH-SOA

The MXLCH-SOA model is based on mixed layer (MXL) theory [Lilly, 1968; Tennekes, 191 1973; Vilà-Guerau de Arellano et al., 2009, 2011], which states that under convective 192 conditions, strong turbulent mixing causes perfect mixing of quantities over the entire 193 depth of the BL. Therefore, scalars and reactants in the convective boundary layer can 194 be characterized by a single value over the whole depth of the BL. The BL dynamics 195 are driven by the surface heat fluxes that are prescribed to the model. The buoyancy 196 entrainment flux is parameterized by a zeroth-order closure assumption in which the 197 entrainment flux is a fixed fraction (β) of the surface buoyancy flux (Table 3). In addition, 198 large-scale meteorological forcings that influence the BL dynamics, like subsidence caused 199 by high pressure systems and advection of heat and moisture, can be prescribed to the 200 model. The transition between the well-mixed BL and the free troposphere is marked by an infinitesimally thin inversion layer. Typical profiles of potential temperature, specific humidity and OA are shown in Fig 1. The complete MXL equations are given by Vilà-Guerau de Arellano et al. [2009] and Ouwersloot et al. [2012].

It is coupled to a reduced chemistry scheme which contains the essentials of the O_3 - NO_x -VOC- HO_x system [Vilà-Guerau de Arellano et al., 2011] and a module for gas/particle-partitioning using the VBS approach [Donahue et al., 2006]. At each time step, the total organic aerosol concentration C_{OA} is calculated from:

$$C_{\text{OA}} = \sum_{i} (X_{p,i}C_i) + \text{OA}_{\text{BG}}; \quad X_{p,i} = \left(1 + \frac{C_i^*}{C_{\text{OA}}}\right)^{-1},$$
 (1)

where C_{OA} is the organic aerosol mass concentration ($\mu \text{g m}^{-3}$), $X_{p,i}$ is the fraction of semi-volatile compound i in the aerosol phase (dimensionless), C_i is the concentration of the semi-volatile organic compound (SVOC), originating from isoprene (IC_i) or terpene

 208 (TC_i) (μ g m⁻³), OA_{BG} the background organic aerosol concentration (μ g m⁻³), which is assumed to be non-volatile and C_i^* is the effective saturation concentration of compound i (μ g m⁻³).

SOA formation from isoprene is implemented in two ways. In the default mechanism, 211 as used in the base case, SVOCs originate directly from first-step oxidation of isoprene by 212 OH (Table 1) using yields derived from lab studies [Kroll et al., 2006], which is common 213 practice in air quality models [e.g. Slowik et al., 2010; Tsimpidi et al., 2010]. Oxidation of 214 isoprene by OH produces both the SVOC species IC_i and IRO_2 , an isoprene peroxy radical, 215 which further influences the gas-phase chemistry and therewith OH regeneration. IC_i 216 partition into the aerosol phase, together with the SVOCs formed from terpene oxidation 217 (TC_i) . The only difference between the isoprene and terpene oxidation products is their 218 molecular weight (136 and 180 g mol⁻¹, respectively). The SVOC yields strongly depend 219 on NO_x concentrations and we account for this by linearly interpolating the high and low NO_x yields (Table 2) as a function of the branching of the reaction of RO_2 from isoprene and terpenes through the NO and the HO₂ channel, respectively [Lane et al., 2008].

In the default mechanism, we omit the formation of IEPOX of which markers are present in the OA at the site [Robinson et al., 2011a; Lin et al., 2012b]. In a sensitivity analysis we include a first-order estimate of ISOA from this new pathway. In future studies, MXLCH-SOA could be used to evaluate the performance of more detailed ISOA forming mechanisms.

In our reduced chemical mechanism (Table 1), the gas-phase oxidation of isoprene is highly simplified and all first generation products of isoprene oxidation are lumped into a single species, which combines methyl vinyl ketone and methacrolein (MVK+MACR). Together, they have a yield of 60% [e.g. Karl et al., 2009]. In the comparison with observations, it is important to note that in the PTR-MS measurements that we use here,

MVK and MACR are also observed as one lumped species, since they have the same molecular weight.

2.3. Initialization of MXLCH-SOA

Initial and boundary conditions are obtained by fitting MXLCH-SOA to the case study observations of dynamics and chemistry, thereby constituting the base case upon which further experiments are based. The initial OA_{BG} in the BL is taken as the total OA concentration, as derived from AMS measurements [Robinson et al., 2011a]. An initial BL concentration of 0.60 $\mu g \text{ m}^{-3}$ is obtained, consisting of 0.04 $\mu g \text{ m}^{-3}$ OOA2, a semi-239 volatile oxidized organic aerosol (OOA) factor, 0.06 μg m⁻³ 82Fac, a factor attributed to 240 IEPOX SOA, 0.30 $\mu g \text{ m}^{-3}$ OOA1, a low-volatile OOA factor, and 0.20 $\mu g \text{ m}^{-3}$ 91Fac, 241 an OA factor associated with biomass burning [Robinson et al., 2011a, b]. The OA_{BG} 242 in the FT is also set to 0.60 μg m⁻³, i.e. BL $OA_{BG} = FT OA_{BG}$. Vertical profiles of 243 OA obtained from aircraft observations support the assumption that OA concentrations 244 in the FT over East Borneo can be at most equal to the OA concentrations in the BL, 245 but not higher [Robinson et al., 2012]. Since entrainment does not dilute the modeled 246 OA_{BG} when concentrations in BL and FT are equal (see Eq. 5), this is the most favorable 247 assumption for calculated OA concentrations in the BL. 248 Dry deposition has been suggested to be an important sink for oxidized VOCs [Karl 249

et al., 2010] and consequently to decrease SOA production significantly [Bessagnet et al., 2010]. Therefore, we included it by applying a deposition velocity of 2.4 cm s⁻¹ for MVK+MACR and for all SVOCs, which is the above-canopy deposition velocity for

MVK+MACR as found in flux measurements in the Amazon [Karl et al., 2010]. Pugh et al. [2010] found that such a large deposition velocity is needed to reconcile modeled MVK+MACR concentrations with measurements. For SVOCs, the V_d of 2.4 cm s⁻¹ is taken as an upper limit as not every SVOC will as effectively taken up and metabolized as MVK+MACR [Karl et al., 2010]. Since the actual deposition velocity for SVOCs is uncertain, we included a simulation in which their deposition is switched off ($V_d = 0$). In this way, we obtain an upper and a lower limit for the effect of dry deposition of SVOCs.

2.4. Numerical experiments

A set of numerical experiments is designed to gain insight in the dynamical and chemical factors that drive the diurnal variability of the organic aerosol concentration as observed on 7 July 2008 during the OP3 campaign. We draw specific attention to physical and chemical processes that are not routinely taken into account by large scale models or that are often omitted in the interpretation of measurements. Figure 1 shows the factors that we account for in the interpretation of observed OA and in this section, we introduce experiments that aim to show the sensitivity of OA formation and concentration to large scale forcings and several issues related to SOA formation from isoprene.

2.4.1. Large-scale meteorological forcings

With large-scale forcings, we refer to all the meteorological phenomena not directly driven by boundary layer processes. It can encompass mesoscale flows induced by different degrees of surface spatial heterogeneities (from small spatial scale to sea breeze) [Ouwersloot et al., 2011] to phenomena like subsiding air motions driven by synoptic scale circulations. We designed two experiments to investigate the influence of large-scale meteorological forcings on the BL dynamics and their subsequent impact on OA. In the

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first experiment we analyze the sensitivity of observed OA to subsiding air motions due to divergence of the horizontal wind. In a previous study, Janssen et al. [2012] showed that OA concentrations in the BL are sensitive to subsidence, because it suppresses BL growth and enhances entrainment. The subsidence velocity w_s (m s⁻¹) is in our modeling approach represented as:

$$w_s = -\omega \cdot h,\tag{2}$$

where ω is the subsidence rate (s⁻¹) and h the BL height (m). In the sensitivity analysis we switch subsidence off by setting ω to 0 and compare it to the base case, as defined in Table 3.

Second, we analyze the sensitivity to advection of heat by the mesoscale flow. The different contributions to heat advection are combined and prescribed to the model as a single advection term (A_{θ}) . The best match with the observations is obtained when this term is negative throughout the day, meaning that relatively cool air is advected (Table 3). In our model setup, we assume that the advection of heat is uniformly distributed within the BL. In the sensitivity analysis we switch advection of heat off and compare it to the base case.

290 2.4.2. OH recycling

In a second sensitivity analysis concerning the sensitivity of OA to chemistry, we evaluate the sensitivity of simulated OA to OH recycling, since this most important oxidant may be underestimated with respect to measurements in high isoprene environments [Lelieveld et al., 2008]. The influence of OH recycling on SOA formation has been accounted for by applying a more detailed chemical mechanism in a global modeling study by Lin et al.

[2012a]. In that study it led to decreased SOA yields, because of modifications in the gas-phase oxidation of isoprene. We do not explicitly account for these reaction pathways, but we consider that it can still be useful to include OH recycling to determine 298 how OH concentrations matching the observations affect the formation of terpene SOA 299 (TSOA) for a case study that is well-constrained by observational data. In our model, 300 OH recycling is parameterized by applying a variable stoichiometric coefficient n in the 301 reaction of IRO₂ with HO₂ (R20) [Lelieveld et al., 2008; Vilà-Guerau de Arellano et al., 302 2011. We compare the base case (no recycling, n=0) with experiments in which n is set 303 to 1 and 2, respectively. 304

2.4.3. SOA formation from IEPOX

Under low-NO_x conditions, isoprene epoxydiols (IEPOX) have been found to be important reactive intermediates in the formation of isoprene SOA [Paulot et al., 2009; Surratt et al., 2010; Lin et al., 2012b]. The chemical pathways for SOA formation from IEPOX are not incorporated explicitly in MXLCH-SOA. However, we mimic the catalyzing role of 300 acidic sulfate aerosol on the formation of SOA from IEPOX by incorporating the chemical 310 mechanism suggested by Paulot et al. [2009] (see Table 5) and using a fixed aerosol yield 311 of 6.4%, which is the highest yield from experiments by Lin et al. [2012b]. In this way, we 312 neglect the complex underlying chemistry, but we obtain a first-order estimation of the 313 magnitude of its effect. In the experiment, the Paulot et al. [2009] mechanism replaces the 314 reactions R9 and R20 as used in the default mechanism. In this mechanism there is some 315 regeneration of OH, but we only evaluate its impact on SOA through IEPOX formation 316 here. 317

We are able to satisfactorily reproduce the dynamics as observed on 7 July 2008 at Bukit

3. Interpretation of observations by modeling

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Atur, see Fig. 4 with initial and boundary conditions as specified in Table 3. The BL 319 height reaching between 800 to 1000 m, as observed by over Borneo from LIDAR [Pearson 320 et al., 2010] and aircraft measurements [Robinson et al., 2012], appears to be the result of 321 the local surface forcing of the sensible (H) and latent heat flux (LE) with superimposed 322 subsidence and advection. Due to the high H (with a maximum of $\sim 400 \text{ W m}^{-2}$), this 323 low BL height can only be explained when subsidence and advection are accounted for: 324 1) subsidence directly suppresses the convective motions that are induced by the surface 325 heat flux and 2) advection of relatively cold air cools the BL and consequently increases 326 its potential temperature difference with the FT $(\Delta\theta)$, which hinders thermal plumes in 327 breaking through this potential temperature inversion to entrain warm air from the FT. In Fig. 1, a typical vertical profile of potential temperature (θ) is sketched to illustrate these effects. These large-scale meteorological forcings may explain the low BL height over Borneo compared to the Amazon, where mixed-layer heights typically exceed 1000 m [Martin et al., 1988]. Later, we explicitly investigate the effect of these meteorological 332 forcings on C_{OA} . 333 The mixed-layer potential temperature ($\langle \theta \rangle$) shows a steep increase of 3 K h⁻¹ after 334 the heat fluxes become positive, at 06:30. This is caused by direct heating of the BL by 335 the sensible heat flux and by entrainment of warm air during the rapid growth of the BL 336 between 09:00 and 11:30, which are both further enhanced by subsidence. Unfortunately, 337 two observations are missing at 07:30 and 08:00, but the fact that we are able to reproduce 338

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the strong gradient in $\langle \theta \rangle$ gives us confidence in the correct representation of its evolution.

Specific moisture $(\langle q \rangle)$, which initially increases due to the evaporation flux, decreases after 09:30 because drier air is entrained during the BL growth and increases again when BL 341 growth ceases around 11:00. 342

The evolution of the gas-phase species O_3 , NO_x and HO_x is shown in Fig. 5 with 343 initial conditions as specified in Table 4. O₃ and NO_x mixing ratios and evolution are 344 reproduced satisfactorily within the bounds set by the scatter in the observations, only 345 $\langle NO_2 \rangle$ is overestimated between 09:00 and 11:00. The exact reason for this is unknown, but it may be due to missing chemistry. The crucial point here is that NO concentrations 347 are simulated well, which is needed to calculate the branching of the low- and high-NO_x 348 SOA yields. Our model calculations show that at $\langle NO \rangle \sim 0.1$ ppb, > 80% of the terpene 349 and isoprene RO₂ reacts with NO, meaning that we are mostly under high-NO_x conditions. 350 This may, however, not be representative of the pristine tropics, where NO mixing ratios are typically in the order of 10^1 ppt. Note that since HO_2 is overestimated by roughly a factor 1.5, this is a lower limit for the branching fraction of the high-NO_x channel. Further, it should be noted that we have neglected the $RO_2 + RO_2$ reaction in these calculations, which may have an influence on the exact branching ratio, but not on the finding that the NO channel dominates over the HO₂ channel. OH is underestimated by a factor of 2-6, depending on considering the lower or the upper bound set by the scatter in the observations. We will discuss possible influences on SOA formation in Sect. 5.2. 358 Figure 6 shows the diurnal evolution of VOCs and OA. F_{ISO} and F_{TERP} are prescribed 359

and scaled to match the fluxes as observed using the virtual disjunct eddy covariance 360 technique, although it should be noted that these observed fluxes can underestimate the 361 true surface flux by 15-20% [Langford et al., 2010]. Temperature driven terpene emissions

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may continue during night time and in the early morning, but are much lower than
those during day time [Langford et al., 2010] and therefore omitted here. Reasonable
agreement is found for mixing ratios of ISO and TERP, which show a similar pattern as
specific moisture. An initial increase in their concentrations between 06:30 and 09:30 is
followed by a decrease, which is related to the rapid BL growth. In the afternoon, both
(ISO) and (TERP) increase again due to continuing emissions and weaker entrainment.
The contribution of chemical destruction is rather constant from 09:00 onwards, due to
a rather constant simulated (OH) (Fig. 5). A more thorough budget analysis is given in
Sect. 4.

(MVK+MACR) is overestimated between 09:00 and 11:00 by around 0.2 ppb. An over-372 estimation of MVK+MACR was found previously in studies of tropical regions [Ganzeveld 373 et al., 2008; Pugh et al., 2010] and several possible explanations have been proposed, in-374 cluding an underestimation of dry deposition and entrainment of MVK+MACR from the residual layer. The applied deposition velocity (V_d) for MVK+MACR of 2.4 cm s⁻¹ as suggested by Karl et al. [2010] partly resolves this issue by lowering their concentrations by 15% compared to the case without deposition. Since we set the FT concentration of MVK+MACR to zero, entrainment in this case only dilutes BL concentrations and can 379 not explain the overestimation. The simplicity of the isoprene oxidation scheme applied here may explain this overestimation. It is relevant to mention that the modeled C_{OA} 381 does not directly depend on (MVK+MACR), since we do not explicitly account for SOA 382 formation from MACR. 383

Both observed and modeled OA concentrations increase during the day, due to SOA formation (Fig. 6d). However, the modeled C_{OA} is lower than the observed OOA2 and

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as the day progresses this underestimation increases to 60% at the end of the simulation. Possible contributors to this underestimation are a misrepresentation of the pathways leading to ISOA formation or too low OH concentration. These will be subject of the sensitivity analyses in Sect. 5. Also the high deposition velocity of the SVOCs could cause an underestimation C_{OA} , due to low SOA formation. The simulation with dry deposition of the SVOCs turned off shows the maximum effect of dry deposition of SVOCs, resulting in a 22% higher C_{OA} at the end of the simulation compared to the case with deposition.

4. Budget analysis of VOCs, SVOCs and OA

To understand how dynamics and chemistry interact and how the diurnal evolution of
OA results from this interaction, it is useful to analyze the cascade of processes that lead
to SOA formation from gas-phase precursors. Here, we show the budgets of key species in
the formation of OA: VOCs from biogenic emissions, an intermediate semi-volatile species
SVOC, and OA as their end product, and how these budgets are coupled to the diurnal
variability of the boundary layer dynamics.

The budget of a primary VOC reads as follows [Janssen et al., 2012]:

$$\frac{\mathrm{d}\langle \mathrm{VOC}\rangle}{\mathrm{d}t} = \underbrace{\frac{F_{\mathrm{VOC}}}{h} \sin\left(\frac{\pi t}{t_d}\right)}_{\text{emission}} + \underbrace{\frac{w_{\mathrm{e}}\Delta VOC}}{h} - \underbrace{\sum_{j} k_{j} \langle \mathrm{VOC}\rangle \langle OX_{j}\rangle}_{\text{chemistry}}$$
(3)

where F_{VOC} is the maximum daily VOC emission flux (ppb m s⁻¹), assuming a sinusoidal diurnal emission profile; h is the BL height (m); t is the time since the start of the simulation (s); t_d is the length of the period during which the heat fluxes are positive (s); w_e is the entrainment velocity (m s⁻¹); ΔVOC is the VOC mixing ratio difference (jump) between the BL and the FT (ppb) (with the jump of a scalar or reactant C defined as $\Delta C = C_{\text{FT}} - \langle C \rangle$); k_j is the reaction rate of VOC with oxidant OX_j (either O_3 or OH); and $\langle OX_j \rangle$ is the mixed layer mixing ratio of oxidant OX_j (ppb).

A similar equation holds for SVOCs, but they do not have an emission term and are removed from the atmosphere by dry deposition, so their budget equation is:

$$\frac{\mathrm{d}\langle C_{\mathrm{i}}\rangle}{\mathrm{d}t} = \underbrace{\frac{w_{\mathrm{e}}\Delta C_{\mathrm{i}}}{h}}_{entrainment} + \underbrace{\sum_{j} \alpha_{i}k_{j}\langle VOC\rangle\langle \mathrm{OX}_{j}\rangle}_{j} - \underbrace{\frac{\mathrm{deposition}}{V_{dCi}\langle C_{\mathrm{i}}\rangle}}_{h}$$
(4)

where ΔC_i is the concentration jump of the SVOC C_i (ppb); α_i the stoichiometric coefficient for C_i ; and Vd_{C_i} its deposition velocity (m s⁻¹).

And finally the budget of OA reads:

$$\frac{\mathrm{d}\langle OA \rangle}{\mathrm{d}t} = \underbrace{\frac{W_{\mathrm{e}}\Delta \mathrm{OA_{BG}}}{h}}^{\mathrm{OA_{BG}-entrainment}} + \underbrace{\sum_{i} \left[X_{p,i} \frac{dC_{i}}{dt} + C_{i} \frac{dX_{p,i}}{dt} \right]}_{\mathrm{GA}} - \underbrace{\frac{\mathrm{deposition}}{V_{dOA}\langle OA \rangle}}_{h} \tag{5}$$

where ΔOA_{BG} is the jump in the background organic aerosol concentration between the
BL and the FT ($\mu g m^{-3}$); and V_{dOA} the deposition velocity of OA.

In Eqs. 3-5, the BL height h modulates the contributions of the emission, entrainment and deposition terms and the entrainment velocity w_e appears in the entrainment term. Through the dependence on h and w_e , the evolution of the chemical species is coupled to the dynamics of the boundary layer, which in turn are affected by large-scale meteorological forcings, as will be shown in Sect. 5.1.

Figure 7 shows how the evolution of C_{OA} depends on the behavior of its precursors. 415 The isoprene tendency (Fig. 7a), which is shown as an example VOC here, has a positive 416 contribution from the emission term, especially between 08:00 and 10:00 when the emission 417 increases (Fig. 6a) and the BL is still shallow as the morning ground inversion is not yet 418 broken. This results in the peak in the isoprene mixing ratio seen in both the observations 419 and model results just before 10:00 (Fig. 6b). Entrainment is important during the period 420 of fast BL growth, between 09:00 and 11:00. As shown by the negative value, entrainment 421 contributes to the decrease of isoprene by introducing residual layer/free tropospheric air 422 characterized by lower ISO mixing ratio. Our findings are corroborated by the observed 423 ISO concentration, which decreases between 09:30 and 11:00 (Fig. 6b). The chemical 424 destruction term, only by OH in this case, is rather constant from 09:00 onwards, but becomes the most important loss term after 11:00. 426

The chemical destruction of isoprene is mirrored in the chemical production of IC₁ (Fig. 427 7b). Due to the low yield of IC₁ (see Table 1), the production of IC₁ is a factor 10^3 smaller 428 than the destruction of ISO. This low yields results in small concentrations changes of IC_1 429 compared to those of ISO. Since the FT concentration is set to zero, the concentration 430 jump is equal to the BL concentration. Therefore, the ratio between entrainment and 431 deposition terms depends solely on the ratio of the entrainment and deposition velocities. 432 w_e peaks at 10:00 at 13 cm s⁻¹ and in our case, the entrainment contribution is larger 433 that of the dry deposition process with $V_d=2.4 \text{ cm s}^{-1}$. 434

The contribution of the entrainment term to the SOA precursors is clearly visible in 435 the evolution of OA (Fig. 7c). The budget of OA includes the entrainment of OA_{BG} , 436 but here the concentrations are equal in BL and FT and therefore the OA_{BG} entrainment 437 term is zero (first term on the right hand side (RHS) of Eq. 5). This is because the OA consists of OA_{BG} and fresh SOA. The former has no concentration gradient between BL and FT and the latter is calculated at each time step as the fraction of the SVOCs that enter the aerosol phase, in which the effect of entrainment is already accounted for. In case of a large concentration jump of OA_{BG} between BL and FT, the entrainment term is important for the evolution of OA [Janssen et al., 2012]. The minimum in the gas/particle-partitioning (second term on the RHS of Eq. 5) is therefore caused by the dilution of the SVOCs due to entrainment. The OA tendency has its peak in the afternoon 445 at 12:30, similarly as the SVOC tendency. This is caused not only by the fact that more 446 SVOC is present, but also by the presence of a larger available organic mass for SVOCs 447 to partition in. Consequently, partitioning of the SVOCs into the aerosol phase will be 448 efficient, i.e. in Eq. 1, $X_{p,i}$ increases with increasing C_{OA} . TSOA and ISOA contribute in 449

similar quantities to the calculated SOA formation. ISOA has lower yields than TSOA,
but the emission of ISO is a factor 5 larger than that of TERP, which compensates for
this. However, the ISOA shown here should be regarded as a lower limit and the formation
of ISOA through reactive intermediates will be discussed in Sect. 5.3.

We omit advection of any of these species and therefore implicitly assume the footprint
area of the site to be horizontally homogeneous for the emission of VOCs, the formation of
SVOCs and the OA concentration. Especially for long-lived species as OA, this assumption
may not hold and we can not rule out a possible contribution of advection to the OA
budget. On the other hand, the satisfactory agreement of the measurements on this
particular day and the campaign averages, makes it unlikely that the source of the air
masses arriving at the site is of major importance for the diurnal variability.

Another process that it has not been taken into account is the dry deposition of OA, as dry deposition of sub-micron aerosols over forests is not well constrained by observations. Recently, Farmer et al. [2013] observed a V_d for the sub-micron mode mass of 0.02 cm s⁻¹ over temperate and tropical forests, which indicates that the contribution of dry deposition would be small. If it would be included it would act to further increase the discrepancy between measured and modeled OA.

5. Sensitivity analyses

5.1. Large scale meteorological forcings

Large-scale meteorological forcings influence the coupled system of BL height (h), mixed 467 layer potential temperature $(\langle \theta \rangle)$ and specific moisture $(\langle q \rangle)$. Consequently, they affect the 468 OA concentration through their influence on h and w_e , as shown in the previous section. 469 In case of subsidence, the BL is compressed by large-scale downward vertical motions, 470 that lead to a stronger heating of the BL because the same amount of heat is distributed 471 over a shallower layer. As a consequence, the potential temperature jump at the BL-FT 472 interface decreases, which enhances the entrainment velocity [Janssen et al., 2012]. In our 473 case study, not taking subsidence into account would have led to an overestimation of h 474 of 250 m at the end of the numerical experiment (Fig. 8), an underestimation of $\langle \theta \rangle$ by 475 1.5 K and a BL, which is too moist. Because of the increased entrainment velocity, OA is diluted more under the presence of subsidence, resulting in 27% lower OA at the end of the simulation. Advection of θ in this case acts to cool the BL. Therefore the potential temperature jump across the BL-FT interface increases, which weakens entrainment (so opposite to the effect of subsidence) and BL growth. Not taking advection of cool air into account 481 results in a BL which is 200 m higher at the end of the run than in the case with advection 482 (similar as for subsidence), an overestimation of $\langle \theta \rangle$ by 4 K and an underestimation of $\langle q \rangle$ 483 by 0.9 g kg⁻¹. Due to the weaker entrainment when advection of cool air is considered,

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the OA precursors are diluted less, leading to more SOA formation and a 16% higher C_{OA}

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Subsidence and advection of cold air, while both decreasing BL height, therefore have opposing effects on C_{OA} .

5.2. OH recycling

The evolution of OA has so far been simulated in this paper without considering OH 489 recycling from isoprene (n = 0 in R20). In this section we explore n=1 and n=2 in R20. This corresponds to daily average recycling rates of OH with respect to the OH consumed by the first-step oxidation of isoprene of 19 and 54\%, respectively, which is below the range of 75 to 120%, as estimated recently by Taraborrelli et al. [2012]. Nevertheless, including OH recycling leads to a better agreement with the observations of $\langle OH \rangle$, especially for n=2 (Fig. 9). This enhancement of $\langle OH \rangle$ leads to an increase of the calculated C_{OA} at the end of the simulation by 25% and 75% for n=1 and n=2, respectively. This enhancement 496 of the calculated C_{OA} , however, is not enough to explain the observed OOA2. On the 497 other hand and as expected from reactions R9 and R30, ISO and TERP are depleted at a 498 faster rate when OH is recycled, which leads to an underestimation of their concentrations 499 as n is increased, especially in the afternoon when isoprene oxidation and OH recycling 500 have a maximum. This is similar to the findings of Pugh et al. [2010] for observations 501 made during OP3 I. They questioned the validity of the assumption of a well-mixed layer 502 and suggested that the segregation of isoprene into distinct plumes could deplete OH in 503 those plumes, which may have affected the measurements. However, the high degree of 504 segregation assumed in their simulations of this effect (50%), was later dispelled by Pugh 505 et al. [2011] and Ouwersloot et al. [2011], who found a reduction of the rate constant of 506 the isoprene and OH reaction due to incomplete mixing of less than 15%.

The main point here is that we are able to reproduce the evolution of ISO and TERP 508 satisfactorily for n=0 as a function of emission, entrainment and chemistry. Hereby, the 509 emissions are constrained by the flux measurements and the correct representation of BL 510 dynamics gives us confidence in the representation of entrainment. Seen in this way, the 511 first-step oxidation by OH (and O_3 in case of TERP), is apparently represented reasonably 512 well in the base case. So while increasing n leads to an improved representation of $\langle OH \rangle$ 513 and C_{OA} , it serves to worsen the match with the observed $\langle TERP \rangle$ and $\langle ISO \rangle$. OH 514 recycling thus has the potential to influence modeled SOA formation in high isoprene 515 environments, but current knowledge is not sufficient to constrain its effects. 516

5.3. SOA formation from IEPOX

Our last sensitivity study focuses on the formation of IEPOX SOA, catalyzed by acidic sulfate aerosol (Table 5). We find concentrations of IEPOX in the order of $\sim 10^{-1}$ ppb after midday (Fig. 10a), which is of the same order-of-magnitude as calculated by a global model for Borneo [Paulot et al., 2009]. The branching between the IRO₂ + HO₂ and the IRO₂ + NO reactions determines the efficiency with which IEPOX is formed. As in the previous experiments, the NO channel dominates, with the HO₂ channel contributing only $\sim 13\%$ to the destruction of IRO₂.

IEPOX SOA is modeled here with a fixed yield of 6.4%, which is the largest yield found in the experiments of Lin et al. [2012b]. 82Fac shows a decrease, possibly due to entrainment, from 08:00 to 09:00 and after that time increases rapidly to reach a concentration which is one order-of-magnitude greater at 14:00 than its minimum at 09:00 (Fig. 10b). We are not able to match the strong increase of 82Fac as observed between

09:00 and 14:00. An IEPOX SOA yield of 35% would be required to explain this rapid

increase.

There are several possible causes for the underestimation of the concentration of 82Fac 531 by the model. First, the concentration of gas-phase IEPOX may be underestimated, but 532 unfortunately there is no data available to validate this. Further, we use a fixed yield 533 of IEPOX SOA, which implies that effects of the OA_{BG} on gas/particle-partitioning are 534 not accounted for, which may affect the aerosol yield of semi-volatile species (see Eq. 535 1). Then, the yield of IEPOX SOA found in chamber studies [Lin et al., 2012b] may 536 be too low, possibly due to yet unknown chemical pathways. Finally, we consider a 537 meteorological factor, long range horizontal transport by advection. Advection may be 538 important because OA has, in contrast to VOCs, a long atmospheric lifetime of days to a 539 week. This means that observed OA could reflect integrated VOC oxidation over a larger area and period rather than in situ production of OA from VOCs. As a consequence, it is plausible that IEPOX SOA is advected to the measurement site that is formed from isoprene emitted at the oil palm plantations located 50 km and further to the South-East, which emit 4-7 times more isoprene than the forest [Hewitt et al., 2010]. However, it is not known what the 82Fac looks like after aging and it could be transformed into OOA2 or OOA1 by the time it reaches the site. Although we can not rule out the contribution of advection, there are two reasons to expect that the fast increase of 82Fac is likely 547 not due to advection. First, the fact that 82Fac was not observed in morning air plane 548 profiles, but that it was observed in the afternoon flights, may indicate that the 82Fac is 549 formed rapidly and mainly from local sources. Second, the time at which the fast increase $_{551}$ of 82Fac begins (after 10:00, see Fig. 10b) corresponds well with the modeled increase $_{552}$ IEPOX and IEPOX SOA.

6. Conclusions

We studied the diurnal evolution of organic aerosol, its gas-phase precursors and their 553 oxidants, coupled to the dynamics of a convective boundary layer for a characteristic 554 situation observed above the tropical forest during the OP3 field campaign. Observations 555 of BL dynamics and chemical species combined with a boundary layer model of physics 556 and chemistry are used to determine the dominant processes in the SOA formation driven 557 by terpene and isoprene emissions. We are able to satisfactorily reproduce the diurnal 558 variability of the BL in terms of its height, potential temperature and specific humidity 559 as driven by land surface and large-scale meteorological forcings. Advection of cooler air 560 and subsidence are important contributions to the characterization of the BL as observed 561 over Borneo and complicate the characterization of a tropical BL climatology. Because 562 of their influence on BL height and entrainment, subsidence and advection of heat affect 563 the diurnal evolution of chemical species in the BL and should be taken into account when interpreting observations of OA. Subsidence and advection of cool air, while both decreasing BL height, have opposing effects on the diurnal trend of C_{OA} .

An analysis of the budgets of VOCs, SVOCs and OA shows the importance of studying dynamical and chemical processes simultaneously in order to understand the diurnal
variability of reactants. Specifically, it shows how the OA budget is strongly modified by
the various processes that shape the diurnal cycle of its gas-phase precursors, in which
the signal of entrainment is clearly visible.

By confronting our model with a rather complete set of data of gas-phase chemistry and organic aerosol, we are able to exclude the influence of some factors that in other studies have been suggested to explain underestimation by models of biogenic OA concentrations. Nevertheless, as in previous studies we underestimate OA concentrations by about a factor of 2, even though we are able to reproduce the diurnal evolution of isoprene and terpene concentrations with observed and prescribed fluxes and we explicitly take the role of entrainment on VOCs and their oxidants into account.

In our investigation of the role of isoprene chemistry in SOA formation, we find that 579 OH recycling decreases the model-measurement discrepancy of OA concentrations, but at 580 the cost of a worse comparison with VOC concentrations. Before isoprene SOA formation 581 can be quantified, OH recycling must be understood. In a final sensitivity analysis, we 582 underestimate the concentration of 82Fac, an OA component thought to be specifically 583 related to SOA from isoprene epoxides (IEPOX), although we incorporate a parametriza-584 tion based on the current knowledge on the formation of IEPOX SOA in our model. There 585 are several factors which may explain this underestimation, and further insights in the formation and evolution in the atmosphere of IEPOX SOA are needed to get a definitive answer. We find that the low-NO_x pathway leading to IEPOX formation is only a minor one under observed NO concentrations. Nevertheless, the concentration of 82Fac is of comparable magnitude as that of OOA2, suggesting that a minor pathway in the gas-phase chemistry of isoprene can still lead to substantial SOA formation. 591

The strong dependence of isoprene SOA formation on NO_x chemistry implies that if NO concentrations increase in Borneo by increased anthropogenic activities, the type and amount of isoprene SOA has the potential to change significantly.

Although incorporating these new pathways does not yet explain the discrepancies between modeled and observed biogenic OA, we propose that models need to account for the different pathways by which isoprene chemistry drives SOA formation, both through

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formation of its second-generation products following the low- and high- NO_x pathways and through its effect on gas-phase chemistry by OH recycling.

Since our model includes only SOA forming species resulting from the first-step oxidation of isoprene and terpenes, there may be additional sources of SOA, for instance higher
generation oxidation products or unmeasured very reactive species that contribute to the
OA budget at the studied site and that should be taken into account in future studies in
order to close the budget.

Finally, we advocate the use of conceptual but realistic models similar to the one presented here to bridge the gap between observations made during chamber studies and field campaigns, on one hand, and the gap between local observations and large-scale forcings, on the other hand, to gain further understanding of the organic aerosol budget in tropical forests.

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Figure captions

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Figure 1. Conceptual representation of the main dynamic and chemical contributions to the organic aerosol budget during the OP3-campaign and sketches of typical vertical profiles of potential temperature (θ) , specific humidity (q) and organic aerosol concentration (C_{OA}) .

Figure 2. Campaign mean and case study diurnal evolution of measured a) potential temperature (θ) , b) isoprene emission flux (F_{ISO}) , c) isoprene concentration (ISO) and d) concentration of the SV-OOA factor (OOA2). The error bars indicate the standard deviation.

Figure 3. Vertical wind speed w at 850 hPa at 14:00 LT on 7 July 2008 and 120 (dashed lines) and 36 h (solid lines) air mass back-trajectories arriving at Bukit Atur at 08:00 (black) and 14:00 (green), respectively, at a pressure altitude of 925 hPa. Note that positive values of vertical wind speed reflect subsidence, since it is expressed in units of pressure. Both are based on ECMWF wind fields with a resolution of 1.125°×1.125°. The inset shows the pressure level of the back-trajectories with the dotted line indicating their approximate arrival time over land.

Figure 4. Diurnal evolution of (a) surface sensible (H) and latent (LE) heat flux, which are both prescribed, (b) boundary layer height (h), (c) mixed layer potential temperature $(\langle \theta \rangle)$ and (d) mixed layer specific moisture $(\langle q \rangle)$ for the case study. Dots indicate tower measurements at 45 m and model results are indicated by lines.

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Figure 5. Diurnal evolution of mixed layer concentrations of (a) O₃, (b) OH, and HO₂ (c) 808 NO and (d) NO₂ for the case study. Markers indicate measurements from the tower at 75 m $(O_3, NO \text{ and } NO_2) \text{ or } 5 \text{ m } (OH \text{ and } HO_2).$ 810

Figure 6. Diurnal evolution of (a) ISO- and TERP-flux, which are both prescribed, and 813 mixed layer concentrations of (b) ISO and TERP, (c) MVK+MACR and (d) OA for the case 814 study. Markers indicate measurements from the tower at 75 (F_{ISO} and F_{TERP} , ISO, TERP,

MVK+MACR) or 33 m (OA). 816

Figure 7. Contribution of the individual processes to the budgets of a) ISO (Eq. 3), b) the 819 SVOC IC₁, which results from oxidation of ISO (Eq. 4) and c) SOA (Eq. 5), split up in the terpene SOA (TSOA) and the isoprene SOA (ISOA) fraction. Note that the contribution of 821 entrainment to the $\langle OA \rangle$ -tendency is zero since the concentrations of OA_{BG} in BL and FT are equal. Therefore, the partitioning of SVOCs from terpenes and isoprene into the aerosol phase forms the only contribution. 824

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Figure 8. Sensitivity of modeled a) organic aerosol concentration C_{OA} , b) BL height h, c) 827 mixed layer potential temperature $\langle \theta \rangle$ and d) mixed layer specific humidity $\langle q \rangle$ to subsidence 828 and advection of heat. Shown are the base case (see Table 3), an experiment without subsidence 829 and an experiment with no advection of θ . 830

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Figure 9. Sensitivity of a) organic aerosol C_{OA} , b) OH, c) ISO and d) TERP to OH recycling.

Shown are experiments for n = 0, 1 and 2 in R20, respectively.

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Figure 10. Concentration of a) IEPOX and b) IEPOX SOA, assuming a fixed yield for the

latter of 6.4%, which is the upper limit under acidic conditions found by Lin et al. [2012b].

Chemical reaction scheme used in the numerical experiments with MXLCH-SOA^a Table 1.

Number			Reaction	Reaction Rate
R1	$O_3 + h\nu$	\rightarrow	$O^1D + O_2$	$3.00 \cdot 10^{-5} \cdot e^{\frac{-0.575}{\cos(\chi)}}$
R2	$\mathrm{O}^1D + \mathrm{H}_2\mathrm{O}$	\rightarrow	2OH	$1.63 \cdot 10^{-10} \cdot e^{\frac{60}{T}}$
R3	$\mathrm{O}^1D + \mathrm{N}_2$	\rightarrow	O_3	$2.15 \cdot 10^{-11} \cdot e^{\frac{110}{T}}$
R4	$\mathrm{O}^1D + \mathrm{O}_2$	\rightarrow	O_3	$3.30 \cdot 10^{-11} \cdot e^{\frac{55}{T}}$
R5	$NO_2 + h\nu$	\rightarrow	$NO + O_3$	$1.67 \cdot 10^{-2} \cdot e^{\frac{-0.575}{\cos(\chi)}}$
R6	$\mathrm{CH_2O} + h\nu$	\rightarrow	HO_2	$1.47 \cdot 10^{-4} \cdot e^{\frac{-0.575}{\cos(\chi)}}$
R7	OH + CO	\rightarrow	$\mathrm{HO}_2^- + \mathrm{CO}_2$	$2.40 \cdot 10^{-13}$
R8	$OH + CH_4$	\rightarrow	$\mathrm{CH_{3}O_{2}}$	$2.45 \cdot 10^{-12} \cdot e^{\frac{-1775}{T}}$
R9	OH + ISO	\rightarrow	$IRO_2 + \alpha_1^I \mathbf{IC}_1 + \alpha_2^I \mathbf{IC}_2 + \alpha_3^I \mathbf{IC}_3$	$2.70 \cdot 10^{-11} \cdot e^{\frac{390}{T}}$
R10	OH + [MVK+MACR]	\rightarrow	$\mathrm{HO_2} + \mathrm{CH_2O}$	$2.40 \cdot 10^{-11}$
R11	$OH + HO_2$	\rightarrow	$H_2O + O_2$	$4.80 \cdot 10^{-11} \cdot e^{\frac{250}{T}}$
R12	$OH + H_2O_2$	\rightarrow	$\mathrm{H_2O} + \mathrm{HO_2}$	$2.90 \cdot 10^{-12} \cdot e^{\frac{-160}{T}}$
R13	$\mathrm{HO}_2 + \mathrm{O}_3$	\rightarrow	$OH + 2O_2$	$2.03 \cdot 10^{-16} \cdot \left(\frac{T}{300}\right)^{4.57} e^{\frac{693}{T}}$
R14	$HO_2 + NO$	\rightarrow	$OH + NO_2$	$3.50 \cdot 10^{-12} \cdot e^{\frac{250}{T}}$
R15	$CH_3O_2 + NO$	\rightarrow	$HO_2 + NO_2 + CH_2O$	$2.80 \cdot 10^{-12} \cdot e^{\frac{300}{T}}$
R16	$IRO_2 + NO$	\rightarrow	$HO_2+NO_2 +0.6[MVK+MACR]+CH_2O$	$1.00 \cdot 10^{-11}$
R17	$OH + CH_2O$	\rightarrow	$HO_2 + CO$	$5.50 \cdot 10^{-12} \cdot e^{\frac{125}{T}}$
R18	$2\mathrm{HO}_2$	\rightarrow	$H_2O_2 + O_2$	*
R19	$\mathrm{CH_3O_2} + \mathrm{HO_2}$			$4.10 \cdot 10^{-13} \cdot e^{\frac{750}{T}}$
R20	$IRO_2 + HO_2$	\rightarrow	nOH + PRODUCTS	$1.50 \cdot 10^{-11}$
R21	$OH + NO_2$	\rightarrow	HNO_3	$3.50 \cdot 10^{-12} \cdot e^{\frac{340}{T}}$
R22	$NO + O_3$	\rightarrow	$NO_2 + O_2$	$3.00 \cdot 10^{-12} \cdot e^{\frac{-1500}{T}}$
R23	$NO + NO_3$	\rightarrow	$2NO_2$	$1.80 \cdot 10^{-11} \cdot e^{\frac{110}{T}}$
R24	$NO_2 + O_3$	\rightarrow	$NO_3 + O_2$	$1.40 \cdot 10^{-13} \cdot e^{\frac{-2470}{T}}$
R25	$NO_2 + NO_3$	\rightarrow	$ m N_2O_5$	**
R26	$ m N_2O_5$	\rightarrow	$NO_3 + NO_2$	***
R27	$N_2O_5 + H_2O$	\rightarrow	$2HNO_3$	$2.50 \cdot 10^{-22}$
R28			$2HNO_3 + H_2O$	$1.80 \cdot 10^{-39}$
R29	$TERP + O_3$	\rightarrow	$\alpha_1^T \mathbf{T} \mathbf{C}_1 + \alpha_2^T \mathbf{T} \mathbf{C}_2 + \alpha_3^T \mathbf{T} \mathbf{C}_3 + \alpha_4^T \mathbf{T} \mathbf{C}_4$	$5.00 \cdot 10^{-16} \cdot e^{\frac{-530}{T}}$
R30			$\alpha_1^T \mathbf{T} \mathbf{C}_1 + \alpha_2^T \mathbf{T} \mathbf{C}_2 + \alpha_3^T \mathbf{T} \mathbf{C}_3 + \alpha_4^T \mathbf{T} \mathbf{C}_4$	$5.00 \cdot 10^{-16} \cdot e^{\frac{-530}{T}}$

^{*} $k = (k1 + k2)/k3; k1 = 2.21 \cdot 10^{-13} \cdot e^{\frac{600}{T}}; k2 = 1.91 \cdot 10^{-33} \cdot e^{\frac{9T}{T}} \cdot c_{air}; k3 = 1 + 1.4 \cdot 10^{-21} \cdot e^{\frac{2200}{T}} \cdot c_{H_2O}$ ** $k = 0.35 \cdot (k_0 k_\infty)/(k_0 + k_\infty); k_0 = 3.61 \cdot 10^{-30} \left(\frac{T}{300}\right)^{-4.1} \cdot c_{N_2}; k_\infty = 1.91 \cdot 10^{-12} \left(\frac{T}{300}\right)^{0.2}$ *** $k = 0.35 \cdot (k_0 k_\infty)/(k_0 + k_\infty); k_0 = 1.31 \cdot 10^{-3} \left(\frac{T}{300}\right)^{-3.5} \cdot e^{\frac{-11000}{T}} \cdot c_{N_2}; k_\infty = 9.71 \cdot 10^{14} \left(\frac{T}{300}\right)^{0.1} \cdot e^{\frac{-11080}{T}}$ a In the reaction rates, T the absolute temperature in Kelvin and χ the solar zenith angle. First-

order reaction rates are in s⁻¹, second-order reaction rates in cm³ molecule⁻¹s⁻¹. Aerosol-forming reactions and products are printed in bold font. $\alpha_1^I - \alpha_4^I$ and $\alpha_1^T - \alpha_4^T$ are stoichiometric coefficients for ISO and TERP, respectively, see Table 2. In R20 n is the rate of OH-recycling, PRODUCTS are species which are not further evaluated in this chemical reaction scheme. Reaction of isoprene O_{3}^{R} O_{3}^{F} O_{3}^{T} is not considered. April 12, 2013, 4:11pm DRAFT

Table 2. Stoichiometric coefficients at $T = 298 \, K$ for the different volatility bins of the SOA precursor categories TERP and ISO, with saturation concentration C_i^* in $\mu \text{g m}^{-3}$ from Tsimpidi et al. [2010].

\overline{i}	1	2	3	4
$\overline{C_i^*}$	1	10	100	1000
TERP, low-NO _x	0.107	0.092	0.359	0.600
TERP, high- NO_x	0.012	0.122	0.201	0.500
ISO, low- NO_x	0.009	0.030	0.015	_
ISO, high- NO_x	0.001	0.023	0.015	_

Figure 1. Conceptual representation of the main dynamic and chemical contributions to the organic aerosol budget during the OP3-campaign and sketches of typical vertical profiles of potential temperature (θ) , specific humidity (q) and organic aerosol concentration (C_{OA}) .

Table 3. The initial and boundary conditions in boundary layer (BL) and free troposphere (FT) as obtained from fitting MXLCH-SOA to the case study observations. All initial conditions are imposed at 06:30 LT. Heat fluxes are applied from 06:30 to 14:00 with $H = \rho c_p \overline{w'\theta'}_s$ and $LE = \rho L_v \overline{w'q'}_s$. t is the time (s) and t_d the length of the day (s). The subscripts s and e indicate values at the surface and the entrainment zone, respectively.

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Property	Value
Initial BL height	300
h (m)	
Subsidence rate	3.10^{-5}
$\omega \ (\mathrm{s}^{-1})$	
Surface sensible heat flux	$0.30\sin(\pi t/t_d)$
$\overline{w'\theta'}_{\rm s} \; ({\rm Kms^{-1}})$	
Entrainment/surface heat-flux ratio	0.2
$\beta = -\overline{w'\theta'}_{\rm e}/\overline{w'\theta'}_{\rm s}$ (dimensionless)	
Initial BL potential temperature	298
$\langle \theta \rangle$ (K)	
Initial FT potential temperature	303.5
$ heta_{\mathrm{FT}} \; (\mathrm{K})$	
Potential temperature lapse rate FT	$0.0030_{h < 800m}$
$\gamma_{ heta}~({ m Km^{-1}})$	$0.0095_{h \ge 800m}$
Advection of potential temperature	-3.10^{-4}
$A_{\theta} (\mathrm{K} \mathrm{m}^{-1}))$	
Surface latent heat flux	$0.16\sin(\pi t/t_d)$
$\overline{w'q'}_{\rm s} \; ({\rm g kg^{-1} m s^{-1}})$	
Initial BL specific humidity	11.5
$\langle q \rangle \; (\mathrm{gkg^{-1}})$	
Initial FT specific humidity	11.4
$q_{\rm FT}~({ m gkg^{-1}})$	
Specific humidity lapse rate FT	-0.0026
$\gamma_q \; ({ m g kg^{-1} m^{-1}})$	

Table 4. Initial mixing ratio in BL and FT and surface emission fluxes of the reactants as obtained from fitting MXLCH-SOA to the case study observations. Species in the reaction mechanism that are not included in this table have zero initial concentrations and zero surface

emissions.					
Species	Initial mixing ratio (ppb)		Surface emission/deposition flux (ppb m s ⁻¹)		
	BL	FT			
O_3	17.5	19.0	$-0.45\sin\left(\frac{\pi t}{t_d}\right)$		
NO	0.05	0.0	$9 \cdot 10^{-3}$ $1.5**$		
NO_2	0.15	0.10	1.5**		
ISO	0.40	0.0	$0.35\sin\left(\frac{\pi t}{t_d}\right)$		
TERP	0.08	0.0	$0.04\sin\left(\frac{\pi t}{t_d}\right)$		
OA_{BG}	0.60^{*}	0.60^{*}	0.0		
MVK+MACR	0.0	0.0	2.4**		
TC_i	0.0	0.0	2.4**		
IC_i	0.0	0.0	2.4**		
CH_4	1800.	1800.	0.0		
CO	100.	100.	0.0		
O_2	2.10^{8}	2.10^{8}	0.0		
N_2	8.10^{8}	8.10^{8}	0.0		
$^*\mu g \text{ m}^{-3}, ^{**}V_d \text{ (cm s}^{-1})$					

Table 5. Paulot et al. [2009] mechanism for oxidation of isoprene under low- NO_x conditions, which replaces R9 and R20 in Table 1 in the sensitivity analysis for the chemical mechanism of SOA formation from isoprene (Sect. 5.3).

	1		,	
R29	ISO + OH	\rightarrow	IRO_2	$2.70 \cdot 10^{-11} \cdot e^{\frac{390}{T}}$
R30	$IRO_2 + HO_2$	\rightarrow	0.880 ISOPOOH + 0.120 OH + 0.047 MACR +	$7.40 \cdot 10^{-13} \cdot e^{\frac{700}{T}}$
			$0.073 \text{ MVK} + 0.120 \text{ HO}_2 + 0.120 \text{ CH}_2\text{O}$	
R31	ISOPOOH + OH	\rightarrow	IEPOX + OH	$1.90 \cdot 10^{-11} \cdot e^{\frac{390}{T}}$
R32	ISOPOOH + OH	\rightarrow	$0.70 \text{ ISOPOO} + 0.300 \text{ HC}_5 + 0.300 \text{ OH}$	$3.80 \cdot 10^{-12} \cdot e^{\frac{200}{T}}$
R33	IEPOX + OH	\rightarrow	IEPOXOO	$5.78 \cdot 10^{-11} \cdot e^{\frac{-400}{T}}$
R34	$IEPOXOO + HO_2$	\rightarrow	$0.725~{\rm HAC} + 0.275~{\rm GLYC} + 0.275~{\rm GLYX} +$	$7.40 \cdot 10^{-13} \cdot e^{\frac{700}{T}}$
			$0.275 \text{ MGLY} + 1.125 \text{ OH} + 0.825 \text{ HO}_2 +$	
			$0.200 \text{ CO}_2 + 0.375 \text{ CH}_2\text{O} + 0.074 \text{ HCOOH} +$	
			0.251 CO	



Figure 2. Campaign mean and case study diurnal evolution of measured a) potential temperature (θ) , b) isoprene emission flux (F_{ISO}) , c) isoprene concentration (ISO) and d) concentration of the SV-OOA factor (OOA2). The error bars indicate the standard deviation.

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Figure 5. Diurnal evolution of mixed layer concentrations of (a) O₃, (b) OH, and HO₂ (c) NO and (d) NO₂ for the case study. Markers indicate measurements from the tower at 75 m (O₃, NO and NO₂) or 5 m (OH and HO₂).

Figure 6. Diurnal evolution of (a) ISO- and TERP-flux, which are both prescribed, and mixed layer concentrations of (b) ISO and TERP, (c) MVK+MACR and (d) OA for the case study. Markers indicate measurements from the tower at 75 (F_{ISO} and F_{TERP} , ISO, TERP, MVK+MACR) or 33 m (OA).

Figure 7. Contribution of the individual processes to the budgets of a) ISO (Eq. 3), b) the SVOC IC₁, which results from oxidation of ISO (Eq. 4) and c) SOA (Eq. 5), split up in the terpene SOA (TSOA) and the isoprene SOA (ISOA) fraction. Note that the contribution of entrainment to the $\langle OA \rangle$ -tendency is zero since the concentrations of OA_{BG} in BL and FT are equal. Therefore, the partitioning of SVOCs from terpenes and isoprene into the aerosol phase forms the only contribution.

Figure 8. Sensitivity of modeled a) organic aerosol concentration C_{OA} , b) BL height h, c) mixed layer potential temperature $\langle \theta \rangle$ and d) mixed layer specific humidity $\langle q \rangle$ to subsidence and advection of heat. Shown are the base case (see Table 3), an experiment without subsidence and an experiment with no advection of θ .

Figure 9. Sensitivity of a) organic aerosol C_{OA} , b) OH, c) ISO and d) TERP to OH recycling. Shown are experiments for n = 0, 1 and 2 in R20, respectively.

Figure 10. Concentration of a) IEPOX and b) IEPOX SOA, assuming a fixed yield for the latter of 6.4%, which is the upper limit under acidic conditions found by Lin et al. [2012b].