

<sup>1</sup> Filtering ground noise from LiDAR returns produces  
<sup>2</sup> inferior models of forest aboveground biomass

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\*\*Running head: Filtering ground noise from forest AGB models

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**8 Abstract**

- 9       1. Airborne LiDAR has become an essential data source for large-scale, high-  
10      resolution modeling of forest biomass and carbon stocks, enabling pre-  
11      dictions with much higher resolution and accuracy than can be achieved  
12      using optical imagery alone. Ground noise filtering – that is, excluding  
13      returns from LiDAR point clouds based on simple height thresholds –  
14      is a common practice meant to improve the ‘signal’ content of LiDAR  
15      returns by preventing ground returns from masking useful information  
16      about tree size and condition contained within canopy returns. Although  
17      this procedure originated in LiDAR-based estimation of mean tree and  
18      canopy height, ground noise filtering has remained prevalent in LiDAR  
19      pre-processing, even as modelers have shifted focus to forest aboveground  
20      biomass (AGB) and related characteristics for which ground returns may  
21      contain useful information about stand density and openness. In particular,  
22      ground returns may be helpful for making accurate biomass predictions  
23      in heterogeneous landscapes that include a patchy mosaic of vegetation  
24      heights and land cover types.
- 25       2. In this paper, we applied several ground noise filtering thresholds while  
26      mapping AGB across New York State (USA), a heterogeneous landscape  
27      composed of both contiguously forested and highly fragmented areas with  
28      mixed cover types. We fit random forest models to predictor sets derived  
29      from each filtering intensity threshold and compared model accuracies,  
30      paying attention to how changes in accuracy correlated with landscape  
31      structure.
- 32       3. We observed that removing ground noise via any height threshold system-  
33      atically biases many of the LiDAR-derived variables used in AGB modeling.

34 We found that that ground noise filtering yields models of forest AGB  
35 with lower accuracy than models trained using predictors derived from  
36 unfiltered point clouds. Models fit to predictors derived from filtered point  
37 clouds performed worse as landscape heterogeneity (as measured by patch  
38 density and edge density) increased.

39 4. Our results suggest that ground filtering should be avoided when mapping  
40 biomass, particularly when mapping heterogeneous and highly fragmented  
41 landscapes, as ground returns are more likely to represent useful ‘signal’  
42 than extraneous ‘noise’ in these cases.

43 *Keywords:* aboveground biomass; ground noise; LiDAR; machine learning;  
44 random forest

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## 45 1. Introduction

46 Accurate assessment of forest carbon stocks for the purposes of greenhouse  
47 gas accounting and climate change mitigation requires high-resolution maps  
48 of above-ground biomass (AGB) across large spatial extents. The production  
49 of these maps has been aided in recent years by the proliferation of publicly  
50 available airborne LiDAR data, allowing researchers access to granular data  
51 representing the 3D profile of the earth’s surface at a landscape scale (Dubayah  
52 & Drake, 2000). By aggregating returns to a pixel or object level and computing  
53 descriptive statistics characterizing the distributions of return heights, modelers  
54 are able to convert these point clouds into tabular data formats which may then  
55 be used to fit regression models for predicting AGB (Hawbaker et al., 2010).

56 However, there exists some disagreement about precisely which returns to  
57 aggregate when computing such statistics. While some LiDAR-based AGB mod-  
58 els include all returns when calculating summary statistics (Hudak et al., 2020),

59 others first filter out returns below various height thresholds when calculating  
60 percentile heights (Ma et al., 2018), density percentiles (Huang et al., 2019),  
61 or their entire suite of predictors (García et al., 2010). Filtering is typically  
62 described as being done to remove ground noise from return data, in order  
63 to avoid having “ground” returns mask any signal contained in the remaining  
64 “canopy” returns. The height threshold used in this process varies across studies,  
65 with examples ranging from 0.3m (García et al., 2010) to 1.3m (Deo et al., 2017;  
66 Ma et al., 2018) to 2m (Anderson & Bolstad, 2013) to 2.5m (Huang et al., 2019).

67 This diversity of approaches demonstrates a lack of consensus about a  
68 preprocessing technique that produces systematically greater estimates of per-  
69 centile heights and other computed predictors. The practice itself appears to  
70 have originated with Nilsson (1996), whose early work with airborne LiDAR  
71 focused on calculating tree heights based on the maximum heights of returns,  
72 as well as stand volume as a function of the mean height of all returns. Nilsson  
73 does not appear to filter returns based on height thresholds; rather, they set  
74 the height values of all points below 2m to 0m, in effect reducing the resulting  
75 mean height values. The following year, Næsset (1997) published what may  
76 be the earliest rationale for ground noise filtering in a study calculating mean  
77 stand height from LiDAR returns, excluding returns below 2m in order to avoid  
78 interference from shrubs, rocks, and other understory features. In concert, these  
79 two studies have provided the justification for filtering out ground returns in a  
80 multitude of forest modeling studies (Anderson & Bolstad, 2013; Magnussen &  
81 Boudewyn, 1998; Wasser et al., 2013), to the extent that it appears to now be  
82 such a commonly accepted practice as to not merit discussion or citation at all  
83 (Hawbaker et al., 2010; e.g. White et al., 2015).

84 Yet this practice, initially justified so as to not include the height of stones  
85 in calculating the mean heights of trees (Næsset, 1997), may not be necessary or

desirable as modelers turn their attention to stand characteristics such as AGB. Increased density of ground returns may be associated with sparser stands, and as a result, the left-censoring of variables derived from LiDAR pulses by omitting ground noise may remove useful information about stand structure available for predictive models. This common practice may therefore result in inferior estimates of forest AGB. Filtering may particularly harm predictive accuracy in less contiguously forested and mixed-use landscapes, as we might expect filtering to exclude more returns in areas without tree canopies intercepting and reflecting pulses. As a result, these filtering procedures may adjust LiDAR-derived variables by greater amounts in these settings compared to contiguously forested regions, given their increased proportion of ground returns. It is likely that modeling such heterogeneous landscapes will be an increasing concern over time, as larger data sets and improved computing power enables modelers to map AGB over larger spatial scales; however, there has not been much discussion in the literature concerning any effects filtering may have on forest AGB predictions either in these landscapes or in more homogeneous settings.

Such a discussion is particularly timely given the current focus on producing high-resolution maps of forest AGB. Numerous studies in recent years have produced such maps using a combination of publicly-available LiDAR and field measurements collected through the United States Forest Service Forest Inventory and Analysis (FIA) program, and despite limitations in LiDAR density and FIA spatial measurement accuracy have produced admirable results. However, such studies may be limiting their success due to this common LiDAR preprocessing procedure.

In this paper, we use publicly-available LiDAR data sets representing a range of contiguously forested and mixed-use landscapes to investigate the impacts of ground noise filtering on predictive models of forest AGB. We set out to first

113 identify how filtering ground noise impacts the distribution of commonly used  
114 LiDAR-derived predictors, using multiple height thresholds as found throughout  
115 the literature. We then fit models to each of these predictor sets using the  
116 random forest algorithm (Breiman, 2001), a popular tool used in modeling AGB,  
117 to assess how the different predictor distributions affected model performance.  
118 This study sought to inform current and future efforts looking to accurately  
119 predict forest AGB using models incorporating predictors derived from airborne  
120 LiDAR data products.

121 **2. Methods**

122 *2.1. LiDAR Data Sets and Site Characteristics*

123 In order to identify the impacts of ground filtering on predictive models  
124 of AGB, we obtained leaf-off LiDAR data sets flown for sixteen regions across  
125 New York State (USA; Figure 1). This data, collected as part of a number of  
126 cross-agency federal initiatives, resembles the relatively low-density and leaf-off  
127 LiDAR relied upon in similar forest AGB modeling work (see for instance Nilsson  
128 et al. (2017), Huang et al. (2019)), and closely resembles the remote sensing data  
129 used in typical modeling practice. Data was acquired between 2014 and 2019 and  
130 had pulse densities between 1.98 and 3.24 points per square meter. Additional  
131 information about individual LiDAR data sets is included as Supplementary  
132 Materials S1.

133 *2.2. Field Data*

134 Field measurements of AGB for all trees measuring  $\geq 12.7$  cm diameter at  
135 breast height were taken as part of the United States Department of Agriculture  
136 (USDA) Forest Inventory and Analysis (FIA) program (Gray et al., 2012), with  
137 true macroplot centroid locations obtained under agreement with the USDA. All  
138 analyses and models used data aggregated from subplots to the plot level; LiDAR

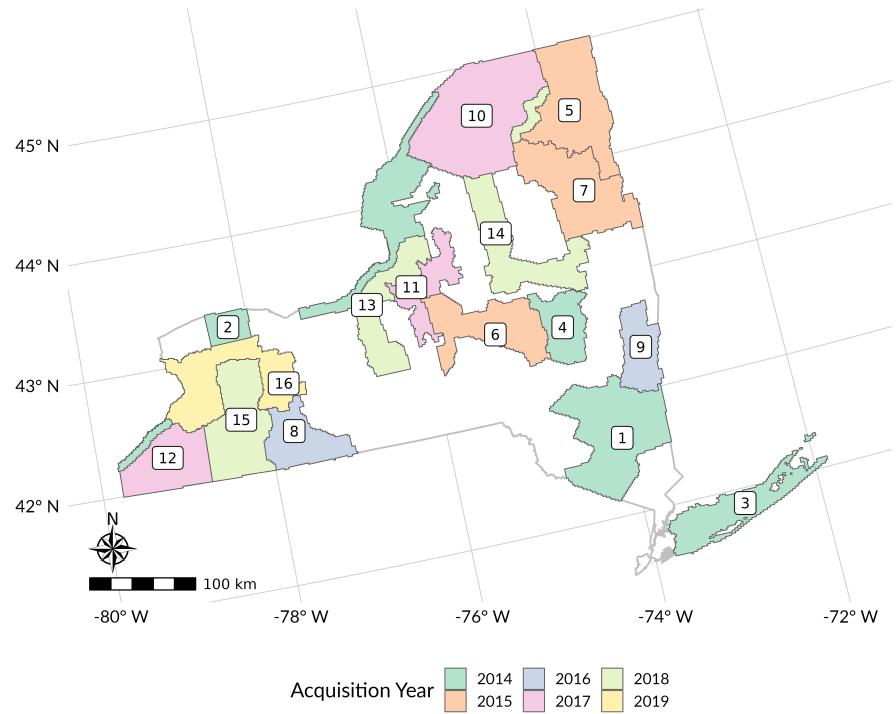


Figure 1: Locations of LiDAR regions within New York State. More information about each region and LiDAR data set is included as Supplementary Materials S1.

139 data was clipped to only the measured subplot areas, with subplot locations  
140 estimated based upon the macroplot centroid, and then pooled prior to predictor  
141 derivation. Plots entirely classified as nonforest (which are not assigned biomass  
142 by the FIA) were excluded from the dataset. Only FIA plots sampled the same  
143 year as LiDAR flights, or FIA plots with measurements both before and after the  
144 LiDAR acquisition year with a difference in AGB within  $[-5\%, \infty)$  (to allow for  
145 forest growth or small-scale disturbance) were used for training and evaluating  
146 models. In situations where FIA year did not match LiDAR acquisition year,  
147 AGB was calculated by linearly interpolating between the values measured in  
148 the temporally closest FIA samples. Plots were additionally excluded if any  
149 subplots were marked as nonsampled, if FIA measurements indicated  $0 \text{ Mg ha}^{-1}$   
150 of AGB but maximum LiDAR return heights at the plot exceeded 10 meters, or  
151 if the convex hull of all LiDAR returns for a subplot contained less than 90% of  
152 the subplot's area. This methodology was chosen to closely resemble the existing  
153 literature on forest AGB mapping (see for instance Huang et al. (2019)). AGB  
154 measurements were recorded in pounds, then converted and area-normalized to  
155 units of megagrams per hectare ( $\text{Mg ha}^{-1}$ ).

156 *2.3. LiDAR Pre-Processing*

157 A digital terrain model (DTM) was calculated for all sites using a k-nearest-  
158 neighbors inverse-distance weighting imputation algorithm (using  $k = 5$ ) as imple-  
159 mented in the lidR R package (Roussel et al., 2020), fit using the points classified  
160 as “ground” within the raw LiDAR point cloud data set. The calculated terrain  
161 was then subtracted from each point’s z value to create a height-normalized  
162 point cloud. Ground noise filtering rules were then applied to create five separate  
163 point clouds for each site, each representing a different ground noise filtering  
164 approach: one point cloud containing all points in the original file (hereafter  
165 referred to as “unfiltered”), one removing all points classified as “ground” in the

<sup>166</sup> original metadata (“ground,” equivalent to a 0m threshold), and three removing  
<sup>167</sup> all points with normalized z values below a 0.1, 1, or 2 meter threshold (“0.1m,”  
<sup>168</sup> “1m,” and “2m,” respectively).

<sup>169</sup> Separate sets of 40 predictors, chosen due to their prevalence in published  
<sup>170</sup> models of AGB and forest structure, were derived from each of these point  
<sup>171</sup> clouds using the lidR R package (Table 1) (Hawbaker et al., 2010; Huang et al.,  
<sup>172</sup> 2019; Pflugmacher et al., 2012, 2014; Roussel et al., 2020). Predictors computed  
<sup>173</sup> for FIA plot locations were derived from only the pooled returns coincident  
<sup>174</sup> with the sampled subplot locations, so as to not include any returns from the  
<sup>175</sup> unsampled regions of the macroplot. For plots where ground noise filtering  
<sup>176</sup> resulted in the removal of all points, variables were set to a default value of 0. As  
<sup>177</sup> highly correlated predictor variables may provide the random forest model less  
<sup>178</sup> information for AGB predictions, relationships between predictors were assessed  
<sup>179</sup> using Spearman’s correlation coefficient. Changes in predictor distributions  
<sup>180</sup> under different filtering methodologies were assessed using Kolmogorov-Smirnov  
<sup>181</sup> statistics (Massey, 1951).

#### <sup>182</sup> 2.4. Model Fitting

<sup>183</sup> AGB models were fit using the ranger R package’s implementation of the  
<sup>184</sup> random forest algorithm (Breiman, 2001; Wright & Ziegler, 2017), a popular  
<sup>185</sup> machine learning technique for predicting forest biomass across landscapes (see  
<sup>186</sup> for instance Huang et al., 2019; Hudak et al., 2020). Separate models were fit  
<sup>187</sup> on predictors calculated using each level of ground noise filtering (“unfiltered,”  
<sup>188</sup> “ground,” “0.1m,” “1m,” and “2m” thresholds) for each LiDAR region and a  
<sup>189</sup> combination of all LiDAR regions, for a total of 85 separate models. Each  
<sup>190</sup> model used data representing all available FIA plots within the relevant LiDAR  
<sup>191</sup> region (Section 2.2). Models were fit using only LiDAR derived predictors, as it  
<sup>192</sup> was expected that including non-LiDAR derived variables might mediate and

Table 1: Definitions of LiDAR-derived predictors used for model fitting.

Predictor	Definition
H0, H10, ... H100, H95, H99	Decile heights of returns, in meters, as well as 95th and 99th percentile return heights.
D10, D20... D90	Density of returns above a certain height, as a proportion. After return height is divided into 10 equal bins ranging from 0 to the maximum height of returns, this value reflects the proportion of returns at or above each breakpoint.
N	Number of LiDAR returns clipped to the given FIA plot or map pixel
ZMEAN, ZMEAN_C	Mean height of all returns (ZMEAN) and all returns above 2.5m (ZMEAN_C)
Z_KURT, Z_SKEW	Kurtosis and skewness of height of all returns
QUAD_MEAN, QUAD_MEAN_C	Quadratic mean height of all returns (QUAD_MEAN) and all returns above 2.5m (QUAD_MEAN_C)
CV, CV_C	Coefficient of variation for heights of all returns (CV) and all returns above 2.5m (CV_C)
L2, L3, L4, L_CV, L_SKEW, L_KURT	L-moments and their ratios as defined by Hosking (1990), calculated for heights of all returns
CANCOV	Ratio of returns above 2.5m to all returns (Pflugmacher et al. 2012)
HVOL	CANCOV * ZMEAN (Pflugmacher et al. 2012)
RPC1	Ratio of first returns to all returns (Pflugmacher et al. 2012)

193 confound the impacts of ground noise filtering.

194 Each of these models were tuned separately using a standard uniform grid  
195 search, with each model evaluated using the same 8,892 combinations of hy-  
196 perparameters detailed in Supplementary Materials S2. Models from this set  
197 were ranked on the basis of mean root mean squared error (RMSE) from 5-fold  
198 cross validation (Stone, 1974) (Equation (1)), with 5 folds chosen to reduce  
199 computational demands. In order to ensure the best model was chosen for each  
200 combination, the top 100 models as determined from 5-fold cross validation  
201 were then evaluated again using leave-one-out cross validation (Lachenbruch  
202 & Mickey, 1968), with the final model fit using the hyperparameter set with  
203 the lowest RMSE. This method ensured that each random forest compared is  
204 the best version of the model that could be fit to these predictors, with the  
205 intention that any difference in model performance will be due to ground noise  
206 filtering and not stochastic differences between models or effort spent in tuning  
207 hyperparameters. Recent work has suggested cross validation assessments of  
208 model accuracy are likely overoptimistic compared to actual predictive accuracy  
209 (Bates et al., 2021), which does not impact our aim of comparing ground noise  
210 filtering approaches within a single study, but should be kept in mind when  
211 assessing these models as AGB estimators in their own right.

212 All modeling work was done using R version 4.0.5 (R Core Team, 2021).

213 *2.5. Model Assessment*

214 Given the scarcity of field data available for some LiDAR regions, models were  
215 evaluated using multiple metrics calculated via leave-one-out cross validation  
216 (Lachenbruch & Mickey, 1968). Performance metrics calculated included root-  
217 mean-squared error both as a value in Mg ha<sup>-1</sup> (RMSE, Equation (1)) and as a  
218 percentage of mean plot AGB (RMSE %, Equation (2)), mean absolute error  
219 (MAE, Equation (3)), and the coefficient of determination ( $R^2$ , Equation (4)).

$$\text{RMSE} = \sqrt{\left(\frac{1}{n}\right) \sum_{i=1}^n (y_i - \hat{y}_i)^2} \quad (1)$$

$$\text{RMSE \%} = 100 \cdot \frac{\text{RMSE}}{\bar{y}} \quad (2)$$

$$\text{MAE} = \left(\frac{1}{n}\right) \sum_{i=1}^n |y_i - \hat{y}_i| \quad (3)$$

$$R^2 = 1 - \frac{\sum_{i=1}^n (y_i - \hat{y}_i)^2}{\sum_{i=1}^n (y_i - \bar{y})^2} \quad (4)$$

220 Where  $n$  is the number of FIA plots included in the data set,  $\hat{y}_i$  is the  
 221 predicted value of AGB,  $y_i$  the AGB value measured at the corresponding  
 222 location, and  $\bar{y}$  the mean AGB value from FIA field measurements.

223 Given that these regions represent a diversity of landscapes, including both  
 224 highly developed regions and large swaths of contiguous forest (Figure 2), we  
 225 investigated how changes in model accuracy due to ground noise filtering varied  
 226 with differences in landscape structure. Landscape structure was quantified for  
 227 each LiDAR region using temporally matching land use/land cover classifications  
 228 from USGS LCMAP (Brown et al., 2020). We computed the proportion of  
 229 pixels classified as forest (Equation (5)), as well as edge density (Equation  
 230 (6)) in units of meter per hectare and patch density (Equation (7)) in units  
 231 of number of patches per 100 hectares for each individual LiDAR region using  
 232 the landscapemetrics R package (Hesselbarth et al., 2019; McGarigal & Marks,  
 233 1995).

$$\text{Forest Cover \%} = \frac{F}{A} \quad (5)$$

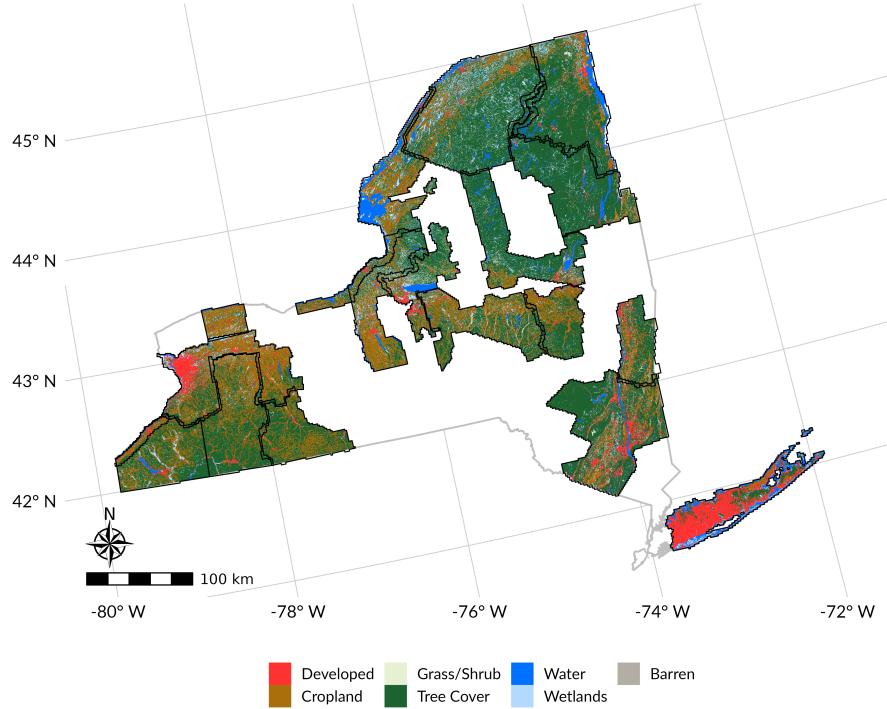


Figure 2: Land cover classifications across LiDAR regions, using land cover classifications from LCMAP (Brown et al., 2020). Lines represent LiDAR data set boundaries.

$$\text{Edge Density} = \frac{E}{A} \cdot 10000 \quad (6)$$

$$\text{Patch Density} = \frac{N}{A} \cdot 10000 \cdot 100 \quad (7)$$

234 Where  $F$  is the area classified as forest in square meters,  $A$  the total landscape  
 235 area in square meters,  $E$  the total landscape edge in meters, and  $N$  the number  
 236 of patches in the region.

237 The relationship between changes in model accuracy due to ground noise  
 238 filtering and landscape structure was measured using Spearman's correlation

<sup>239</sup> coefficient ( $\rho$ ).

<sup>240</sup> **3. Results**

<sup>241</sup> *3.1. Landscape Structure*

<sup>242</sup> Edge density ranged from 38.73 to 100.17 meters per hectare, patch density  
<sup>243</sup> from 8.63 to 23.70 patches per 100 hectares, and forest coverage from 15.38% to  
<sup>244</sup> 83.29% of each LiDAR region (Figure 3). LiDAR regions had between 9 and  
<sup>245</sup> 126 FIA plots available for models after applying plot inclusion rules, for a total  
<sup>246</sup> of 874 plots in the combined data set (Table 3).

<sup>247</sup> *3.2. Variable Distribution*

<sup>248</sup> Filtering out ground noise resulted in shifts in predictor distributions (Figure  
<sup>249</sup> 4). Filtering returns based upon z-thresholds or ground classifications resulted  
<sup>250</sup> in systematically elevated height percentile and return density predictors (the H  
<sup>251</sup> and D prefixed predictors in Table 1; Figure 4), with differences persisting into  
<sup>252</sup> the highest percentiles. Notable differences in distributions also existed for all  
<sup>253</sup> L-moment based predictors, with increasing height thresholds associated with  
<sup>254</sup> increased magnitude of difference. Mean predictor values for each ground noise  
<sup>255</sup> filtering method, alongside Kolmogorov-Smirnov test statistic values comparing  
<sup>256</sup> the distributions of filtered predictors to that of the unfiltered predictors, are  
<sup>257</sup> presented in Supplementary Materials S3.

<sup>258</sup> Shifts in predictor distributions resulted in changes to covariance among  
<sup>259</sup> variables, as measured via Spearman correlation coefficients. More aggressive  
<sup>260</sup> filtering approaches were generally associated with stronger positive correlations  
<sup>261</sup> and collinearity between all variables (Figure 5).

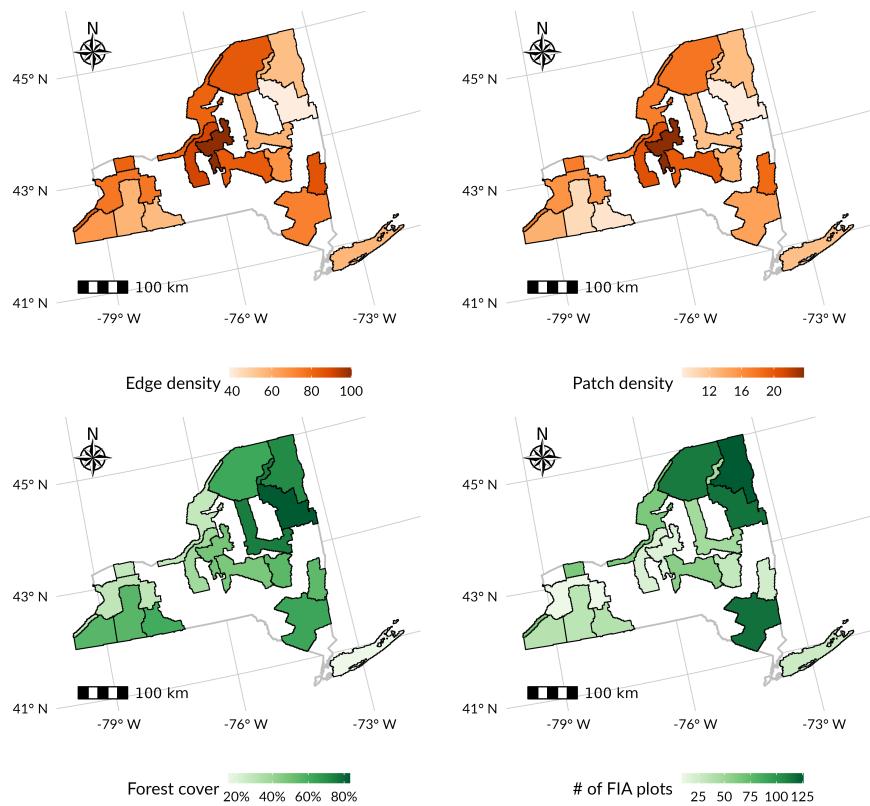


Figure 3: Landscape fragmentation metrics, derived from LCMAP LULC classifications for all LiDAR regions used in this project at year of LiDAR acquisition, and number of FIA plots available for modeling after inclusion rules within each coverage.

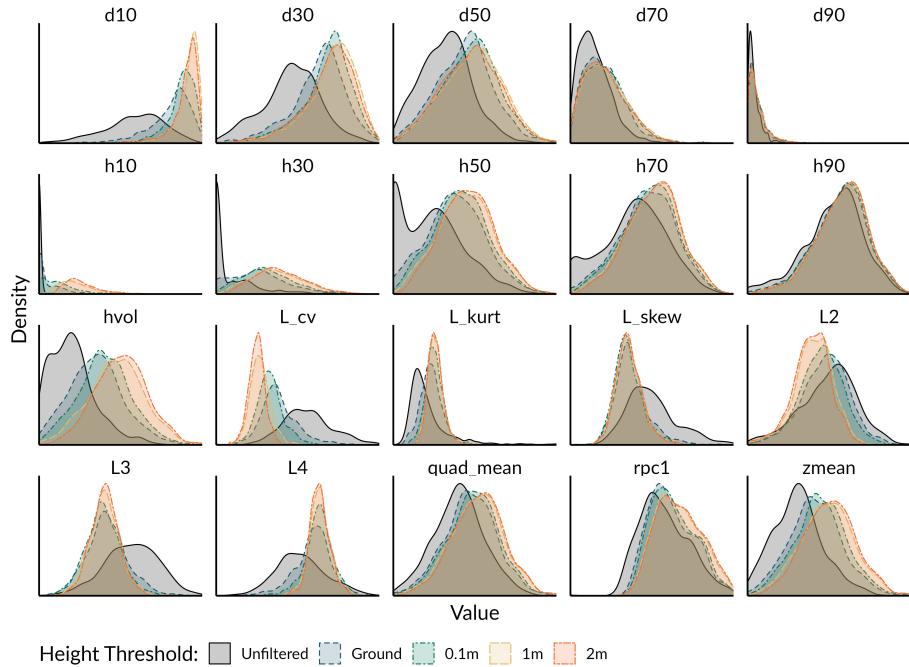


Figure 4: Selected LiDAR-derived predictor distributions for five ground noise filtering approaches, using all LiDAR regions combined. Each subplot is scaled independently so that the X-axis represents the full range of that predictor and the Y-axis represents the full range of the kernel density estimate of that predictor.

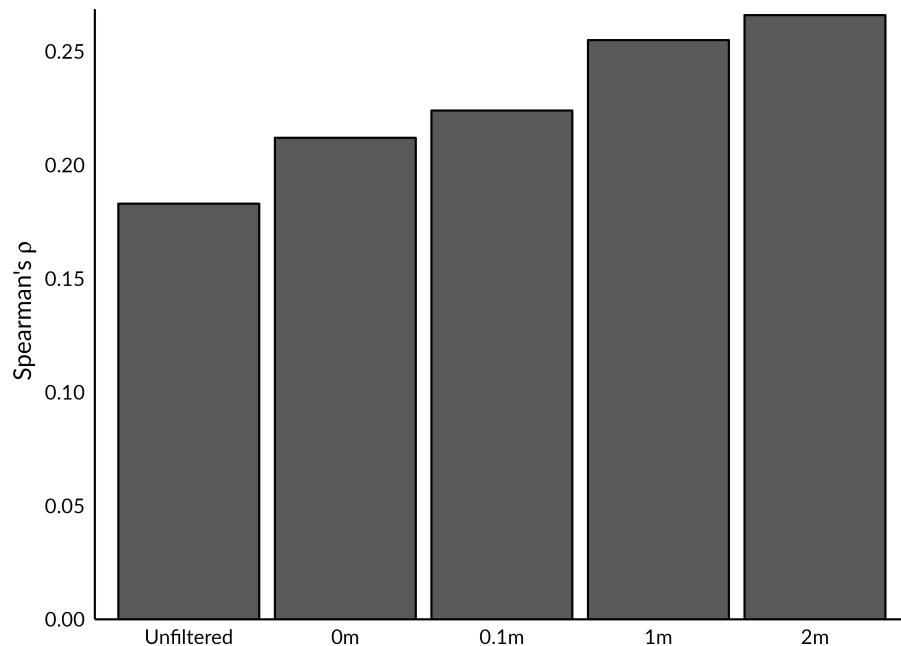


Figure 5: Mean Spearman correlation coefficients between LiDAR-derived variables calculated from point clouds processed with five different ground noise filtering methodologies across the combined data set. Variables with standard deviations of 0 after filtering (such as when minimum return height at all plots became 0 due to ground noise filtering) were excluded from calculations.

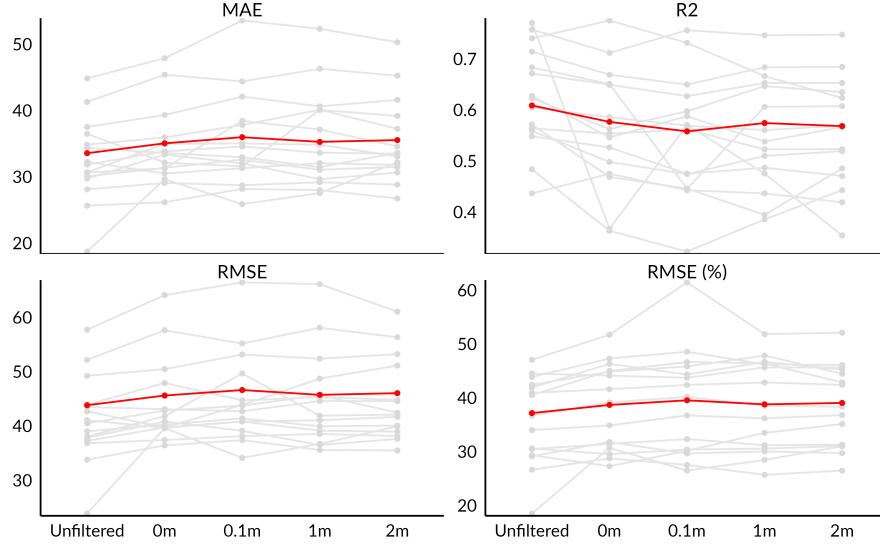


Figure 6: Model accuracy metrics at each ground noise filtering height threshold. Red line indicates models fit to all LiDAR regions (874 FIA plots), while grey lines represent each individual LiDAR region model with more than 10 FIA plots. Metrics are defined in Section 2.5.

262     3.3. *Model Performance*

263     Models fit to the unfiltered set of predictors were almost always equally or  
 264     more accurate than those fit to predictors derived from filtered point clouds  
 265     (Figure 6, Table 2, Table 3). Model accuracy generally decreased as filtering  
 266     thresholds increased, with RMSE % for models fit to all regions combined  
 267     increasing from 37.18% when using the unfiltered data set to 39.06% when using  
 268     a threshold of 2 meters (Figure 6). An exception to this pattern was the Erie,  
 269     Genesee, & Livingston LiDAR region, which saw improvements in accuracy with  
 270     filtering procedures; this is likely related to the small sample size available for  
 271     this region (with only 9 FIA plots available for models) making this region highly  
 272     susceptible to small changes in the predictor space or hyperparameter space.

273     Model accuracy was impacted most by filtering when the area mapped  
 274     was highly fragmented or contained large tracts of non-forested land (Table

Table 2: Model accuracy metrics for the model fit to the combined data set at various ground filtering height thresholds. The complete set of model accuracy metrics for all LiDAR regions is included as Supplementary Materials S4.

	Unfiltered	0m	0.1m	1m	2m
RMSE	43.826	45.608	46.622	45.734	46.044
RMSE (%)	37.177	38.689	39.548	38.795	39.058
MAE	33.560	35.048	35.974	35.271	35.540
R2	0.609	0.577	0.558	0.574	0.568

275 4). Increasing edge and patch densities were both positively correlated with  $\Delta$   
 276 RMSE following ground noise filtering, indicating greater increases in RMSE  
 277 after filtering in more heterogenous landscapes, while increasing forest cover was  
 278 negatively correlated with  $\Delta$  RMSE (Table 4).

#### 279 4. Discussion

280 This study set out to evaluate empirical support for threshold-based ground  
 281 noise filtering for models of forest AGB. We found that this common practice  
 282 results in worse models of AGB, with lower predictive accuracy across multiple  
 283 combinations of LiDAR regions and filtering thresholds representing a broad  
 284 spectrum of landscape structures. While filtering had minimal impact on predic-  
 285 tive accuracy in the most contiguously forested regions, the increasing research  
 286 focus on large-scale “wall-to-wall” biomass mapping and potential for decreased  
 287 accuracy following filtering procedures should encourage future modeling studies  
 288 to use unfiltered point clouds when deriving variables for AGB models.

##### 289 4.1. Ground noise filtering produces inferior predictive models

290 Our study demonstrates that the ground noise filtering approaches commonly  
 291 used in preprocessing data for models of AGB systematically biases LiDAR-  
 292 derived variables, with an end result being inferior models that produce less

Table 3: RMSE for each LiDAR region at various ground filtering height thresholds. The complete set of model accuracy metrics for all LiDAR regions is included as Supplementary Materials S4.

Region	# Plots	RMSE				
		Unfiltered	0m	0.1m	1m	2m
All Regions	874	43.826	45.608	46.622	45.734	46.044
Allegany & Steuben	38	43.478	43.102	42.702	44.589	44.577
3 County	117	49.238	50.479	53.164	52.394	53.238
Cayuga & Oswego	19	23.873	39.584	34.126	36.687	39.947
Clinton, Essex & Franklin	126	37.255	39.742	40.821	39.135	38.952
Columbia & Rensselaer	23	42.689	39.721	43.885	48.731	51.126
Erie, Genesee & Livingston	9	56.942	51.461	30.960	32.279	49.731
Franklin & St. Lawrence	113	36.818	37.411	38.121	38.538	38.143
Fulton, Saratoga, Herkimer & Franklin	47	37.840	40.823	39.105	36.496	37.610
Great Lakes	64	33.790	36.419	37.395	35.569	35.497
Long Island	26	38.047	41.796	49.667	41.893	42.107
Madison & Otsego	58	39.014	40.252	41.412	39.937	40.072
Oneida Subbasin	17	40.490	42.839	43.741	45.677	42.455
Schoharie	30	52.186	57.639	55.185	58.110	56.344
Southwest (spring)	37	43.921	47.921	44.715	45.297	44.806
Southwest (fall)	34	57.744	64.114	66.464	66.126	61.060
Warren, Washington & Essex	116	41.072	39.656	40.816	41.054	41.678

Table 4: Correlation (Spearman’s  $\rho$ ) between  $\Delta$  RMSE (%) and landscape structural metrics at various filtering thresholds.  $\Delta$  RMSE (%) represents the difference between RMSE (%) for the filtered scenario compared to RMSE (%) without filtering; positive correlations represent error increasing as the landscape metric increases. The negative correlation with increasing forest cover implies that areas with less forest are more negatively impacted by filtering; it is not generally the case that contiguously forested landscapes are positively impacted.

Filtering threshold	Edge density	Patch density	% Forest cover
0m	0.026	0.056	-0.368
0.1m	0.141	0.218	-0.382
1m	0.365	0.332	-0.388
2m	0.321	0.326	-0.388

293 accurate predictions than models fit on unfiltered data sets (Figure 4, Figure  
 294 6, Table 2). Increasing intensity of ground noise filtering was generally, but  
 295 not universally, associated with worse model performance (Table 2, 3). These  
 296 patterns were generally stronger as landscape fragmentation increased, with  
 297 the correlation between model errors and landscape fragmentation increasing as  
 298 filtering intensity increased.

299 These results are intuitive when thinking about the actual stand character-  
 300 istics that may lead to an abundance or lack of ground returns. Dense forest  
 301 stands making full use of the available light should be expected to have fewer  
 302 returns reaching below the uppermost branches, while regions with many gaps  
 303 in the canopy will have more such returns. If we conceive of our returns as  
 304 providing information about the height structure of the stand as a whole, rather  
 305 than about individual trees, it stands to reason that variables calculated using  
 306 all returns are more informative about stand metrics such as AGB than those  
 307 using filtered point clouds which may sacrifice information about stand openness.  
 308 This could explain the impact of ground noise filtering seen in this study using  
 309 leaf-off LiDAR; we might expect this impact to be even more pronounced were  
 310 we to use leaf-on LiDAR in its place.

Our results also make sense mechanistically given the properties of the random forest algorithm used to construct AGB models in this study. Random forests excel at predicting outcomes based upon the consensus of weak learners (Breiman, 2001), individual decision trees which themselves rely upon small and ephemeral correlations between predictor variables and the outcome of interest. As shown in Figure 5, ground noise filtering approaches increase positive correlations between predictor variables, with the resulting increased collinearities shrinking the number and magnitude of possible weak correlations between individual variables and AGB (Langford et al., 2001). While the decision trees comprising the random forest may be able to take advantage of the correlations between predictor variables and the outcome to achieve similar accuracy as when trained on unfiltered data sets, we would not expect that a process that uniformly increases the positive linear correlation between variables would be associated with improved predictions.

Insights drawn from these results may not be limited to only machine learning based models. Anderson and Bolstad (2013) briefly note that, when fitting linear models to predict AGB, models based on unfiltered point clouds always provided better results than those fit to predictors calculated using only returns above 2 meters. However, few other AGB modeling studies have performed similar investigations, necessitating our current study. Our conclusions may not apply to AGB models of non-forest systems; investigations of ground noise filtering as a preprocessing step for models of corn AGB found improvements in predictive accuracy with relatively low height thresholds (Luo et al., 2016), emphasizing that commonly accepted data processing practices cannot be assumed to transfer across systems or domains to new questions of interest.

336 *4.2. Differences between regional models*

337 Although we found that models fit using predictors derived from unfiltered  
338 point clouds to be the most consistently accurate, the degree to which ground  
339 noise filtering damaged predictive accuracy and the relationship between filtering  
340 intensity and accuracy varied between regions. More fragmented landscapes  
341 tended to be more impacted by ground noise filtering, with model error increasing  
342 the most in landscapes with greater patch and edge densities and less forest  
343 cover (Table 4). These regions are characterized by large amounts of marginal  
344 forestland, resulting in a higher proportion of plots with low AGB and lower  
345 mean AGB values compared to more contiguously forested regions. As a result,  
346 it stands to reason that more returns in these highly fragmented landscapes are  
347 affected by the filtering procedure, removing more information from the model  
348 and resulting in inferior predictions.

349 *4.3. Limitations as AGB models*

350 The models discussed in this study were fit using only LiDAR-derived pre-  
351 dictors so as to maximize the potential effect of ground noise filtering on model  
352 performance, as predictors obtained from additional data sources may be corre-  
353 lated with unfiltered predictors and as such used in their place by the random  
354 forest algorithm (Efron, 2020), mediating the impact of filtering. Additionally,  
355 these models were fit using extensive hyperparameter tuning performed via  
356 an automated process so as to avoid unintentionally biasing results by giving  
357 different models differing levels of attention or time in tuning. This process  
358 ensures that our models can be directly compared without worrying about a  
359 human “thumb on the scale,” but might result in models which fail to generalize  
360 beyond the training data due to the extensive tuning process. Further, model  
361 assessment was done using leave-one-out cross validation, which is sufficient  
362 for comparison between individual models but lacking as a way to characterize

363 model AGB predictions spatially and across multiple scales (Riemann et al.,  
364 2010). While none of these limitations impact the comparison of ground noise  
365 filtering approaches at the center of this study, in combination they prevent us  
366 from using these models to make fine-scale estimates about AGB stocks across  
367 these regions and how model predictions compare to regional FIA estimates.

368 *4.4. Recommendations for future models*

369 Our results and examination of the literature suggest that ground noise  
370 filtering procedures are not well justified for studies modeling AGB, given  
371 both the potential information lost about stand density and structure, and the  
372 empirical inferiority of models fit using predictors derived from filtered point  
373 clouds. We make no such claim about researchers modeling other variables using  
374 LiDAR-derived predictors. For instance, the procedure likely makes sense when  
375 modeling mean tree heights similar to Næsset's (1997) study which originated  
376 the practice of ground noise filtering. The best data preprocessing procedure  
377 will necessarily depend on the purpose of the model (Sambasivan et al., 2021).

378 More generally, this study offers a reminder that all data preprocessing steps  
379 should be well justified in the context of any new analytical workflow. While  
380 tracing methodological details to their origins in the literature may not always be  
381 fruitful, researchers should ideally have the ability to separate out small sections  
382 of their data to evaluate model performance with and without the proposed  
383 procedure. The results of these small tests may justify including the procedure  
384 in the data preprocessing workflow for the full data set, or alternately lead a  
385 team to remove a processing step to save data cleaning time without damaging  
386 predictive accuracy. In these early days of big data in environmental science,  
387 we remain wanting for a cohesive theory of optimal prediction (Efron, 2020);  
388 as a result, beliefs about methodological improvements are still best tested by  
389 experiment.

390 **5. Conclusion**

391 Our study demonstrates that preprocessing LiDAR point clouds to filter  
392 out ground noise may be detrimental when making predictions of above-ground  
393 biomass using machine learning methods. This impact is particularly notable  
394 within mixed-use and otherwise heterogeneous landscapes, given the increased  
395 proportion of ground returns recorded when mapping these areas compared  
396 to contiguously forested regions. Although well-justified in its original context  
397 of modeling mean stand heights, the persistence of ground noise filtering in  
398 LiDAR-based AGB modeling appears to produce less accurate predictions than  
399 could be achieved using currently available data.

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405 **7. Conflict of Interest**

406 The authors declare that they have no conflicting interests.

407 **8. Authors' Contributions**

408 MM, LJ, and CB conceived the ideas and designed methodology; MM, LJ, EB,  
409 and CB analysed the data; MM led the writing of the manuscript. All authors  
410 contributed critically to the drafts and gave final approval for publication.

<sup>411</sup> **9. Data Availability Statement**

<sup>412</sup> Data available from the Dryad Digital Repository <https://doi.org/10.5061/dryad.t1g1jwt47>  
<sup>413</sup> (Mahoney et al., 2022).

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# Supplementary Materials

## Contents

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## 1 S1: LiDAR data sets

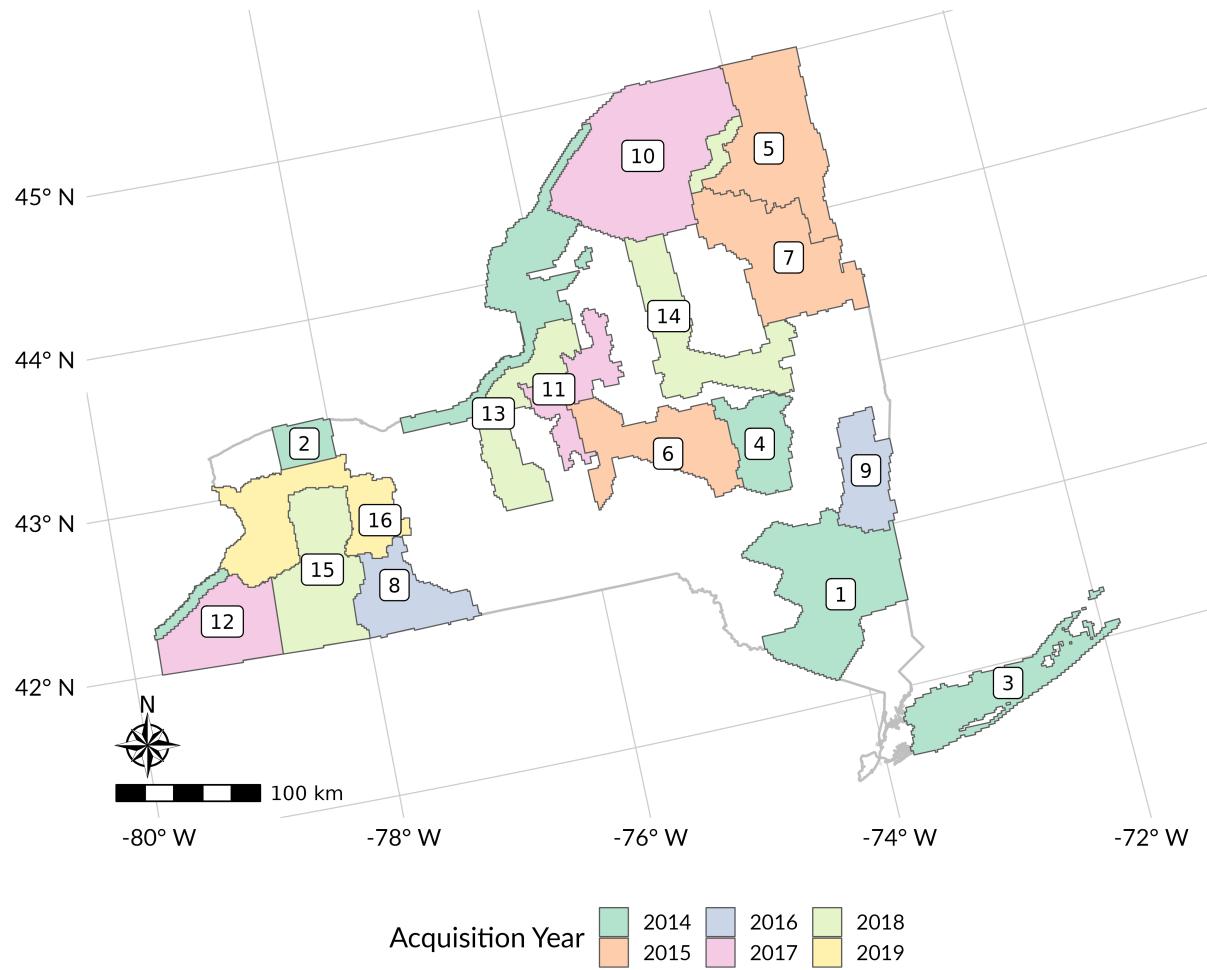


Figure 1: Boundaries for all LiDAR coverages used in this project, colored by year of data acquisition. Numbers on each coverage represent the "index" value of that coverage in table Supplementary Materials S1.

Table 1: Lidar region characteristics. "Index" numbers reflect identifier numbers as used in Supplementary Materials Figure 1. Region names reflect the naming conventions used by the NYSGPO; this often, but not always, reflects included counties. Area values are approximate and in square kilometers. Density values are in points per square meter ( $\text{ppm}^2$ ). Edge density is in units of meters per hectare, and patch density in number of patches per 100 hectares.

Index	Region Name	Acquisition Year	Area	Density	Edge Density	Patch Density	Forest Cover	Citation
1	3 County	2014	7,370	2.04	73.20	14.63	62.57%	United States Geological Survey (2015a)
2	Great Lakes	2014	5,780	2.04	81.24	17.23	27.77%	United States Geological Survey (2015c)
3	Long Island	2014	3,170	2.04	56.57	12.34	15.38%	Woolpert, Inc (2014)
4	Schoharie	2014	2,500	2.04	67.48	13.60	54.95%	United States Geological Survey (2015b)
5	Clinton, Essex & Franklin	2015	1,110	2.04	54.03	12.49	71.94%	Quantum Spatial (2016)
6	Madison & Otsego	2015	4,780	2.18	83.99	19.76	47.73%	Axis GeoSpatial, LLC (2016a)
7	Warren, Washington & Essex	2015	6,280	3.24	38.73	8.63	83.29%	Atlantic Inc (2015)
8	Allegany & Steuben	2016	3,410	2.04	55.41	9.39	60.50%	New York Office of Information Technology Services (2016)
9	Columbia & Rensselaer	2016	2,600	2.60	86.36	18.62	54.18%	Axis GeoSpatial, LLC (2016b)
10	Franklin & St. Lawrence	2017	9,880	2.04	84.66	17.84	62.14%	Quantum Spatial (2017b)
11	Oneida Subbasin	2017	2,550	2.04	100.17	23.70	49.09%	Quantum Spatial (2017a)
12	Southwest (spring)	2017	4,460	2.04	65.44	13.40	55.76%	New York Office of Information Technology Services (2017)
13	Cayuga & Oswego	2018	4,450	2.04	90.75	20.55	37.86%	New York Office of Information Technology Services (2018a)
14	Fulton, Saratoga, Herkimer & Franklin	2018	5,010	1.98	57.74	12.31	75.21%	Quantum Spatial (2018)

Table 1: Lidar region characteristics. "Index" numbers reflect identifier numbers as used in Supplementary Materials Figure 1. Region names reflect the naming conventions used by the NYSGPO; this often, but not always, reflects included counties. Area values are approximate and in square kilometers. Density values are in points per square meter ( $\text{ppm}^2$ ). Edge density is in units of meters per hectare, and patch density in number of patches per 100 hectares.  
*(continued)*

Index	Region Name	Acquisition Year	Area	Density	Edge Density	Patch Density	Forest Cover	Citation
15	Southwest (fall)	2018	5,660	2.04	57.52	10.20	55.74%	New York Office of Information Technology Services (2018b)
16	Erie, Genesee & Livingston	2019	5,670	2.04	76.19	15.69	29.22%	New York Office of Information Technology Services (2019)

## 2 S2: List of hyperparameters used in tuning random forests and evaluated ranges

Table 2: List of hyperparameters used in tuning random forests and evaluated ranges. Models were tuned using a uniform grid containing all combinations of all values of all variables save num.trees, which was tuned separately.

Hyperparameter	Definition	Evaluated Range
mtry	Number of variables to include in each node	Integers between 3 and 40
min.node.size	The minimum number of observations per terminal node	Integers between 3 and 15
sample.fraction	Fraction of observations to sample	0.2 to 1.0 in increments of 0.1
replace	Whether or not to sample with replacement	TRUE and FALSE
num.trees	Number of trees to aggregate	100 to 2000 in increments of 100. Tuned separately from other hyperparameters for only the best performing model given that performance tends to improve with additional trees, independent of other parameter values.

### 3 S3: Mean values of LiDAR-derived variables and Kolmogorov-Smirnov test statistic values (in parentheses).

Table 3: Mean values of LiDAR-derived variables and Kolmogorov-Smirnov test statistic values (in parentheses) for the combined data set. Variables calculated for plots with no returns following filtering were set to 0, which may result in counterintuitive comparisons.

	0.1m	Ground	1m	2m	Unfiltered
cancov	0.868 (0.722)	0.799 (0.538)	0.951 (0.886)	0.997 (0.994)	0.586
cv	0.575 (0.716)	0.683 (0.541)	0.464 (0.847)	0.428 (0.897)	1.090
cv_c	0.407 (0.000)	0.407 (0.000)	0.407 (0.000)	0.407 (0.000)	0.407
d10	0.865 (0.732)	0.798 (0.565)	0.905 (0.810)	0.893 (0.799)	0.582
d20	0.770 (0.644)	0.712 (0.498)	0.802 (0.700)	0.789 (0.681)	0.522
d30	0.670 (0.568)	0.619 (0.442)	0.696 (0.600)	0.682 (0.574)	0.456
d40	0.564 (0.477)	0.521 (0.379)	0.582 (0.506)	0.568 (0.473)	0.384
d50	0.450 (0.397)	0.416 (0.317)	0.462 (0.414)	0.448 (0.389)	0.307
d60	0.329 (0.320)	0.304 (0.261)	0.335 (0.333)	0.322 (0.291)	0.224
d70	0.206 (0.260)	0.190 (0.209)	0.208 (0.260)	0.199 (0.232)	0.140
d80	0.098 (0.229)	0.090 (0.192)	0.099 (0.222)	0.094 (0.189)	0.066
d90	0.026 (0.199)	0.023 (0.165)	0.026 (0.195)	0.025 (0.176)	0.016
h10	2.409 (0.912)	1.361 (0.876)	4.448 (0.969)	5.040 (0.974)	0.156
h20	5.002 (0.783)	3.555 (0.652)	6.650 (0.864)	7.156 (0.891)	0.715
h30	7.224 (0.634)	5.867 (0.526)	8.537 (0.700)	8.968 (0.740)	1.885
h40	9.204 (0.514)	8.042 (0.411)	10.232 (0.574)	10.596 (0.606)	3.801
h50	10.998 (0.405)	10.107 (0.341)	11.801 (0.461)	12.114 (0.479)	6.164
h60	12.672 (0.301)	11.995 (0.246)	13.302 (0.346)	13.566 (0.360)	8.830
h70	14.334 (0.221)	13.803 (0.184)	14.813 (0.252)	15.031 (0.264)	11.383
h80	16.062 (0.166)	15.700 (0.135)	16.419 (0.191)	16.584 (0.205)	13.918
h90	18.095 (0.105)	17.842 (0.090)	18.324 (0.118)	18.430 (0.129)	16.722
h95	19.490 (0.082)	19.321 (0.068)	19.654 (0.090)	19.734 (0.095)	18.536
h99	21.510 (0.047)	21.422 (0.040)	21.596 (0.050)	21.634 (0.054)	21.036
hvol	9.526 (0.548)	8.281 (0.416)	11.229 (0.648)	11.938 (0.701)	4.875
L_cv	0.323 (0.723)	0.378 (0.546)	0.262 (0.854)	0.242 (0.903)	0.546
L_kurt	0.051 (0.457)	0.031 (0.325)	0.060 (0.526)	0.059 (0.539)	0.012
L_skew	-0.014 (0.466)	0.020 (0.406)	-0.005 (0.460)	-0.002 (0.457)	0.201
L2	3.271 (0.191)	3.448 (0.109)	2.930 (0.347)	2.830 (0.392)	3.585
L3	-0.086 (0.518)	0.006 (0.447)	-0.051 (0.532)	-0.043 (0.537)	0.542
L4	0.137 (0.457)	0.060 (0.324)	0.152 (0.526)	0.147 (0.540)	-0.090
max	23.413 (0.000)	23.413 (0.000)	23.412 (0.001)	23.410 (0.002)	23.413
min	0.104 (1.000)	0.005 (0.349)	1.077 (0.999)	2.030 (0.998)	0.000
n	3,872.788 (0.267)	4,178.778 (0.219)	3,543.318 (0.316)	3,423.683 (0.339)	5,471.683
quad_mean	12.197 (0.253)	11.711 (0.193)	12.754 (0.304)	12.976 (0.327)	10.027
quad_mean_c	13.122 (0.000)	13.122 (0.000)	13.122 (0.000)	13.122 (0.000)	13.122
rpc1	0.633 (0.189)	0.620 (0.152)	0.661 (0.255)	0.667 (0.272)	0.578
z_kurt	-0.583 (0.294)	-0.615 (0.193)	-0.513 (0.343)	-0.515 (0.352)	0.080
z_skew	-0.057 (0.408)	0.050 (0.352)	-0.015 (0.375)	0.002 (0.366)	0.589
zmean	10.667 (0.403)	9.879 (0.305)	11.608 (0.474)	11.942 (0.501)	7.374
zmean_c	12.151 (0.000)	12.151 (0.000)	12.151 (0.000)	12.151 (0.000)	12.151

## 4 S4: Model accuracy by LiDAR region

Table 4: RMSE for each LiDAR region at various ground filtering height thresholds.

Region	# Plots	RMSE				
		Unfiltered	0m	0.1m	1m	2m
All Regions	874	43.826	45.608	46.622	45.734	46.044
Allegany & Steuben	38	43.478	43.102	42.702	44.589	44.577
3 County	117	49.238	50.479	53.164	52.394	53.238
Cayuga & Oswego	19	23.873	39.584	34.126	36.687	39.947
Clinton, Essex & Franklin	126	37.255	39.742	40.821	39.135	38.952
Columbia & Rensselaer	23	42.689	39.721	43.885	48.731	51.126
Erie, Genesee & Livingston	9	56.942	51.461	30.960	32.279	49.731
Franklin & St. Lawrence	113	36.818	37.411	38.121	38.538	38.143
Fulton, Saratoga, Herkimer & Franklin	47	37.840	40.823	39.105	36.496	37.610
Great Lakes	64	33.790	36.419	37.395	35.569	35.497
Long Island	26	38.047	41.796	49.667	41.893	42.107
Madison & Otsego	58	39.014	40.252	41.412	39.937	40.072
Oneida Subbasin	17	40.490	42.839	43.741	45.677	42.455
Schoharie	30	52.186	57.639	55.185	58.110	56.344
Southwest (spring)	37	43.921	47.921	44.715	45.297	44.806
Southwest (fall)	34	57.744	64.114	66.464	66.126	61.060
Warren, Washington & Essex	116	41.072	39.656	40.816	41.054	41.678

Table 5: RMSE (%) for each LiDAR region at various ground filtering height thresholds.

Region	# Plots	RMSE (%)				
		Unfiltered	0m	0.1m	1m	2m
All Regions	874	37.18	38.69	39.55	38.80	39.06
Allegany & Steuben	38	44.52	44.13	43.73	45.66	45.64
3 County	117	34.02	34.88	36.74	36.21	36.79
Cayuga & Oswego	19	18.53	30.73	26.49	28.48	31.01
Clinton, Essex & Franklin	126	36.69	39.14	40.20	38.54	38.36
Columbia & Rensselaer	23	29.35	27.31	30.17	33.50	35.15
Erie, Genesee & Livingston	9	43.61	39.41	23.71	24.72	38.09
Franklin & St. Lawrence	113	40.95	41.61	42.40	42.86	42.43
Fulton, Saratoga, Herkimer & Franklin	47	26.66	28.76	27.55	25.71	26.49
Great Lakes	64	43.89	47.30	48.57	46.20	46.11
Long Island	26	47.08	51.71	61.45	51.83	52.10
Madison & Otsego	58	30.49	31.45	32.36	31.21	31.31
Oneida Subbasin	17	42.42	44.88	45.83	47.86	44.48
Schoharie	30	41.93	46.31	44.34	46.69	45.27
Southwest (spring)	37	29.15	31.81	29.68	30.06	29.74
Southwest (fall)	34	40.53	45.00	46.65	46.41	42.86
Warren, Washington & Essex	116	30.60	29.55	30.41	30.59	31.05

Table 6: MAE for each LiDAR region at various ground filtering height thresholds.

Region	# Plots	MAE				
		Unfiltered	0m	0.1m	1m	2m
All Regions	874	33.560	35.048	35.974	35.271	35.540
Allegany & Steuben	38	30.681	34.978	35.034	34.838	33.073
3 County	117	37.520	39.326	42.094	40.619	41.600
Cayuga & Oswego	19	18.721	29.625	25.873	27.545	32.179
Clinton, Essex & Franklin	126	30.068	31.358	32.594	31.056	31.495
Columbia & Rensselaer	23	36.475	32.143	31.581	40.088	39.186
Erie, Genesee & Livingston	9	47.617	46.599	24.948	27.124	41.119
Franklin & St. Lawrence	113	28.114	29.075	28.728	29.180	28.822
Fulton, Saratoga, Herkimer & Franklin	47	29.724	33.365	32.020	29.640	30.655
Great Lakes	64	25.647	26.167	28.175	27.999	26.735
Long Island	26	30.655	31.290	38.446	37.132	34.594
Madison & Otsego	58	31.842	33.841	34.575	33.633	33.480
Oneida Subbasin	17	34.843	35.944	37.813	40.118	37.229
Schoharie	30	41.306	45.392	44.386	46.290	45.248
Southwest (spring)	37	34.352	33.594	32.959	31.450	33.620
Southwest (fall)	34	44.847	47.884	53.584	52.324	50.291
Warren, Washington & Essex	116	32.257	30.497	31.295	32.056	31.800

Table 7: R2 for each LiDAR region at various ground filtering height thresholds.

Region	# Plots	R2				
		Unfiltered	0m	0.1m	1m	2m
All Regions	874	0.609	0.577	0.558	0.574	0.568
Allegany & Steuben	38	0.564	0.554	0.562	0.523	0.523
3 County	117	0.548	0.527	0.475	0.487	0.471
Cayuga & Oswego	19	0.771	0.367	0.571	0.475	0.354
Clinton, Essex & Franklin	126	0.559	0.498	0.474	0.510	0.519
Columbia & Rensselaer	23	0.740	0.775	0.731	0.666	0.624
Erie, Genesee & Livingston	9	0.544	0.657	0.883	0.871	0.650
Franklin & St. Lawrence	113	0.605	0.586	0.569	0.561	0.569
Fulton, Saratoga, Herkimer & Franklin	47	0.623	0.563	0.598	0.647	0.635
Great Lakes	64	0.714	0.669	0.650	0.684	0.684
Long Island	26	0.683	0.651	0.446	0.606	0.607
Madison & Otsego	58	0.672	0.649	0.627	0.653	0.653
Oneida Subbasin	17	0.571	0.469	0.446	0.394	0.485
Schoharie	30	0.628	0.546	0.588	0.538	0.567
Southwest (spring)	37	0.758	0.712	0.756	0.746	0.747
Southwest (fall)	34	0.484	0.363	0.323	0.385	0.443
Warren, Washington & Essex	116	0.437	0.475	0.443	0.437	0.419

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