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Weakly electric fish for biomonitoring water quality

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Environmental pollution is a major issue that calls for suitable monitoring systems. The number of possible pollutants of municipal and industrial water grows annually as new chemicals are developed. Technical devices for pollutant detection are constructed in a way to detect a specific and known array of pollutants. Biological systems react to lethal or non-lethal environmental changes without pre-adjustment, and a wide variety have been employed as broad-range monitors for water quality. Weakly electric fish have proven particularly useful for the purpose of biomonitoring municipal and industrial waters. The frequency of their electric organ discharges directly correlates with the quality of the surrounding water and, in this way, concentrations of toxicants down to the nanomolar range have been successfully detected by these organisms. We have reviewed the literature on biomonitoring studies to date, comparing advantages and disadvantages of this test system and summarizing the lowest concentrations of various toxicants tested. Eighteen publications were identified investigating 35 different chemical substances and using six different species of weakly electric fish.

Keywords: weakly electric fish; detection limits; electric organ discharge; biomonitoring; biosensor

1. Introduction

Population growth and higher living standards will cause ever-increasing demands for good quality municipal and industrial water. In many ways, the spectrum of pollutants is increasing. Where sewage flow is used for irrigation, the drainage water could contain disinfection by-products (DBPs) that were formed in the drinking water when it was chlorinated for public water supply and the chlorine reacted with natural dissolved organic carbon [1–3]. In that respect, there is great concern about adverse long-term health effects, for instance, cancer [4]. In addition to containing DBPs, the treated sewage effluent and, hence, the waters into which it is discharged, can also be expected to contain chemicals from pharmaceutical and other industries, hospitals and other medical facilities, households where unused medicines are flushed down the toilets, and human excreta containing incompletely metabolized medicines, which may have biological effects [5,6]. While not directly toxic or carcinogenic, these chemicals may produce adverse health effects by interfering with hormone production (endocrine disruptors), by weakening immune systems, and by other biological responses [7,8]. Other potential contaminants in the drainage water from sewage-irrigated crops and plants are substances like humic and fulvic acids. These are known precursors of DBPs when the water is chlorinated. The humic substances are formed as stable end products wherever organic matter is biodegraded [9].

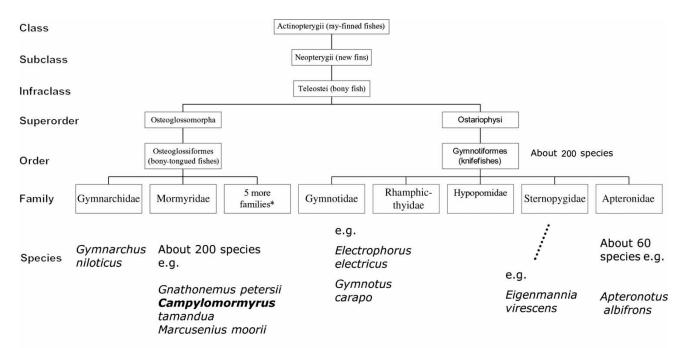
It is evident that a broad spectrum of methods is necessary for evaluations of the condition of waste bodies using direct measurements of biological organisms (macroinvertebrates, fish and plants). This article will focus on a special category: the use of weakly electric fish for the purpose of biomonitoring water quality. It summarizes the available information on detection limits in these types of fish.

2. Weakly electric fish

In the late nineteenth century, several fish with weakly electric organs were discovered. In contrast to the well known electric eel (*Electrophorus electricus*), which can deliver electric shocks of several hundred volts, the discharges of weakly electric fish are comparably low ranging from mV to a few volts [10,11]. These fish belong to the orders of Osteoglossiformes and Gymnotiformes (Figure 1) and are found in the fresh waters of Africa and South America. The order of Osteoglossiformes falls into seven families, where only the Mormyridae (*c*.200 species) and Gymnarchidae (one recent species) represent electric fish. By contrast, all fish in the five families of Gymnotiformes (*c*.200 species) posses electric organs (Gymnotidae, Rhamphichthyidae, Hypopomidae, Sternopygidae and Apteronotidae) [11].

Weakly electric fish are mainly nocturnal animals that live and hunt in muddy, disturbed waters [11]. The fish emit series of brief electric signals, termed electric organ

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^{*} Arapaimidae, Hiodontidae, Notopteridae, Osteoglossidae, Pantodontidae. These families contain no weakly electric species.

Figure 1. Simplified taxonomic tree of weakly electric fish. In the order of Osteoglossiformes, only members of the families Gymnarchidae and Mormyridae possess electric organs. In contrast, all species in the order of Gymnotiformes feature electric organs.

discharges (EOD), that are emitted by a specialized electric organ, which is made up of modified muscle or nerve cells [12]. In Mormyridae and Gymnarchidae, the electric organ is located in a relatively small area at the tail base [13], whereas it disperses through a large part of the body in Gymnotiformes [11]. The simplest type of an EOD is a monophasic pulse. However, several species produce multiphasic discharges of two or more phases [14].

The African Mormyridae and the South American Gymnotidae, Rhamphichthyidae and Hypopomidae (all Gymnotiformes) often produce multiphasic pulses at very irregular rates the pulses are relatively brief in duration compared with their interpulse interval (ratio usually less than 5%) resulting in a low EOD frequency (c.50–100 Hz). These fish are called 'pulse' fish [11,12,15]. The Sternopygidae and Apteronotidae (Gymnotiformes) and the one recent species of the Gymnarchidae (Gymnarchus niloticus) continuously produce very regularly spaced mono- or multiphasic pulses where the duration of the EOD is about the same as the duration of the interpulse interval, resulting in an almost sinusoidal EOD. The discharges mostly occur at high frequency (c.1–2 kHz). Because of the wavelike EOD appearance, these fish are called 'wave' fish [11,12,16].

The EODs build up stable, three-dimensional electric fields around the fish, resembling electric dipoles. The fish can detect distortions of this field caused by any nearby object (e.g. other fish) with a conductivity different from water. The distortions are detected by electroreceptors which are located all over the fish's skin [15]. Through

the pioneering work of Hans Lissmann and Ken Machin it became clear that these fish had evolved a formerly unknown electric sense that allowed them to locate objects (active electrolocation) and hunt in their turbid environment [17]. The sudden appearance and detection of a novel object cause transient increases in the frequency of the EODs of the 'pulse' fish, enabling the fish to gain additional detailed information about that object. Thereby, these fish not only are able to detect the mere presence of certain objects but also can determine their distance and discriminate between different shapes, volumes and the composition of those objects, even in complete darkness [18–20]. If the object stays close to the fish, the fish gets habituated and does not exhibit behavioural responses anymore (adaptation) [17].

Besides navigation and electrolocation, further research revealed that these fish also use their electric sense to communicate to and interact with other electric fish, intraspecifically or interspecifically [21–33]. In some species, the pulse duration increases with body size over the lifetime and might play a role in male—male competition (intrasexual selection) and female choice (intersexual selection) [33–42].

In addition, the EODs vary with changes in temperature, pH and water quality [11,43]. An increase in EOD frequency is, however, not necessarily connected to changes in locomotion. The correlation between EOD frequency and water quality has been exploited for several biomonitoring studies [44–53].

3. Detection of environmental changes by weakly electric fish

Today, environmental pollution is a major issue, calling for suitable monitoring systems. The number of possible pollutants of drinking water grows annually as new chemicals are developed. Technical devices for pollutant detection are constructed in a way so that a specific and known array of pollutants is detected. Some of the disadvantages of those devices are that (1) new pollutants may slip through the technical warning system unless readjustment takes place or new devices are installed, (2) it is technologically, financially and operationally not feasible to detect over four million chemicals known to date, and (3) samples are taken on a discontinuous basis. Because biological systems are affected by pollutants and react in distinct ways to environmental (lethal or non-lethal) changes without pre-adjustment, they have been employed as broad-range monitors for water quality. Here, the lack of specificity is advantageous. A comprehensive review of known biological test systems by Cairns and van der Schalie summarizes advantages and disadvantages [54]. In addition to bacteria and other microorganisms, fish have also been used as biosensors. Among these, the use of weakly electric fish has been proven especially useful in various experiments, because the electric discharges are particularly sensitive to water quality and are easy to detect and process (see section 4).

Changes in the frequency, the amplitude, the waveform or the pulse duration of weakly electric fish and the duration of these changes can be used to identify a response of the fish to the presence of certain chemicals in the body of water.

4. Laboratory set-up for detection of EODs

The very basic set-up to visualise the EODs is a tank including the fish and two submerged electrodes attached to an oscilloscope. The electrodes detect the voltage variations and should be placed in vicinity to the fish. It is generally useful to offer a tube, in which the fish can hide during daytime, as a concession to its natural environment (muddy, disturbed waters). Moreover, the fish will normally not leave this hiding place during daylight, facilitating the anchoring of the electrodes in permanent close vicinity to the fish. For documentation and evaluation, it is necessary to amplify and digitize the signals and store them on a computer (Figure 2). In this way, the frequency, the intervals, the amplitudes and the waveform of the signals can be evaluated. Whereas first experiments started with single measurements at a time [44,55], later improvements allowed measurements of four [45] and even eight [47,48,56] individuals at the same time (including control fish). Acclimatization of the fish to the tank is crucial to get a stable baseline [48,52,57]. For detecting the dependence of the EODs on chemical and physical influences (pH, temperature, dissolved substances) a constant water flow has

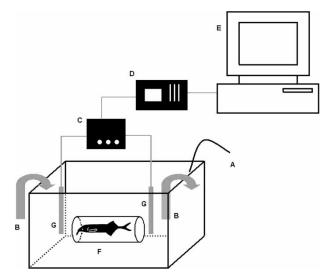


Figure 2. General scheme of the laboratory set-up for measuring the EODs of weakly electric fish. A: air supply; B: water supply (in and out); C: AC amplifier; D: converter (digitization of the EODs; E: computer (storage, analysis); F: fish in tube; G: electrodes. For clarity, control of temperature and pH has been omitted.

to be provided, as displayed in Figure 1 in [45]. Individual differences and diurnal irregularities have to be taken into consideration by statistical analysis of variance. The most advanced and sophisticated set-up so far has been employed by Thomas *et al.* [47], which even provided the opportunity to analyse, by Fourier analysis, the single harmonics of the recorded signals as a function of the toxicant, thereby granting higher reliability.

5. Detection limits of weakly electric fish

Most of the biomonitoring studies solely determined the change in the frequency of EODs. Only Thomas *et al.* additionally investigated the waveform after drug application [47,48]. The biomonitoring studies have paid little or no attention to behavioural modifications after drug exposure [44–46,49–52]. Again, only Thomas *et al.* have touched on eventual changes in locomotor behaviour, which have always been preceded by changes in the EOD characteristics [47,48].

In addition to these biomonitoring studies, further information on EOD sensitivity and detection limits comes from the field of neurophysiology, where the modes of action of stimulants and depressants of the central nervous system have been investigated [55,56,58–63]. Here again, the main parameter investigated was the modulation of the EOD frequency (all but one study – [60]); however, some studies also report on visual observations of slightly changed behaviour after drug exposure [55,56,62,63]. Only two studies directly measured the movement of the fish as a function of drug concentrations using electrodes [58,60], but here the data obtained were composite signals, not exclusively attributable to movement. In general, no strict

correlation between EOD modifications and behaviour has been found in these studies. In most cases, the EOD was modified but no changes in behaviour were seen [55,56,60,62,63]. Four of these studies also investigated the signal amplitude. In two studies, the signal amplitude was unchanged [61,62], whereas Krivoy *et al.* [58,60] found amplitude variations for three out of five substances (up to 90% amplitude reduction). However, the informative value of the estimated amplitude changes is limited because of the set-up limitations mentioned by the authors [58].

Altogether, six different species of weakly electric fish were used in the experiments investigating modifications of the EODs after exposure to various environmental conditions. Thirty-five different substances were investigated (reported in 18 publications). The lowest detected concentrations are summarized in Tables 1 and 2 (Mormyridae) and Table 3 (Gymnotiformes).

5.1. Mormyridae

Most experiments have been performed with *Gnathonemus petersii* ('elephantnose fish'; see Figure 1a in the online supplementary material) and *Campylomormyrus tamandua* ('blunt-jaw elephantnose fish'; synonym *G. tamandua*). Both species have been exposed to various toxicants in several studies [44–46,49–53,55,56]. Lowest detected

concentrations were 20 nm bis[tri-n-butyl tin]oxide (TBTO) and 78 nm copper. Cyanide was detected down to 0.2 μm. All detected concentrations are summarized in Tables 1 and 2. The usual period under observation has been up to 60 min. The response and adaptation times after exposure vary greatly (minutes to hours) depending on the nature and concentration of the applied toxicant and the type of fish. In general, it has been found that lower concentrations result in longer periods before the fish responds [44,46,49–52]. The sensitivity appears to be almost similar for the toxicants, but slightly higher for ammonium and phosphate in G. petersii compared with C. tamandua. Besides this interspecific variation, individuals of one species also exhibit a highly variant and inconsistent pattern of response (increase or decrease or no response in frequency/change of variance prior to and after exposure) to toxicant exposure (intraspecies variation) [50,51]. Therefore, it is the change in the pulsing pattern of each fish itself that characterizes a response to a toxicant rather than the quality of the variation between individuals or species. Temperature and pH have great influence on the frequency of EODs in Mormyridae [64-66]. Moreover, the EODs greatly alter during the diurnal cycle according to the general behaviour of the fish or exogenous factors such as visual disturbances or noise [46,56,57,67], even in the absence of any chemical stimulations. Generally, the frequency of EODs is lower during the day, but

Table 1. Detection limits of Gnathonemus petersii (Mormyridae) determined by EOD output.

Substance	Lowest concentration for detecting effect	Failed to detect	Remarks	Source
Apomorphine	307 μΜ	n.d. ²		[55]
Haloperidol	109 μM	n.d.		[55]
Mercury	$0.5\mu\mathrm{M}$	n.d.	Response after 20 h	[44]
Sodium arsenite	46 µM	d.m. ³	d.m. ³	[53]
Ammonium	$28\mu\mathrm{M}$	n.d.		[52]
Phosphate	$0.21\mu M$	$0.15 \mu M$	No response within 60 min at $0.15 \mu\text{M}$	[52]
Nitrate	n.a. ⁴	n.a.	No response to 161–1.2 mM	[52]
Bis[tri-n-butyl tin]oxide; TBTO	$0.02\mu M$	n.d.		[52]
Cadmium	$0.89\mu M$	n.d.		[50]
Chromium(III) sulphate	$0.51\mu\mathrm{M}$	n.d.		[50]
Trivalent chromium	0.25 M	n.d.	Presumably chromium(III) sulphate; preliminary results	[46]
Potassium dichromate	$1.4\mu M$	$0.7 \mu M$		[50]
Potassium dichromate	$0.34\mu M$	n.d.	Preliminary results	[46]
Copper	$7.9\mu\mathrm{M}$	n.d.	After 4h	[44]
Copper	$0.078\mu\mathrm{M}$	$0.016\mu M$	Response to 0.016 μM not significant within 60 min	[50]
Cyanide	$3.8\mu M$	n.d.	After 4h	[44]
Cyanide	$0.2\mu M$	n.d.		[51]
Cyanide	$0.2\mu M$	n.d.	Preliminary results	[46]
Phenol	9.6 μM	$5.3 \mu\text{M}$	5.3 μM only tested for 30 min	[45]
Lindane	$0.17\mu\mathrm{M}$	n.d.		[49]
Atrazine	$0.15\mu M$	n.d.		[49]
Atrazine	$0.12\mu M$	n.d.		[56]

¹In general, there was no <u>systematic</u> analysis of detection thresholds. The values reported in this column just represent concentrations, where no change in the fish's <u>behaviour</u> was detected after exposure.

 $^{^{2}}$ n.d. = not determined.

 $^{^{3}}$ d.m. = data missing; original literature not available – data taken from [44].

 $^{^4}$ n.a. = not applicable.

Table 2. Detection limits of Campylomormyrus tamandua and Marcusenius moorii (both Mormyridae) determined by EOD output.

Species	Substance	Lowest detected concentration	Failed to detect 1	Remarks	Source
Campylomormyrus	Ammonium	55 μΜ	28 μΜ		[52]
tamandua	Phosphate	0.53 μM	0.21 μM		[52]
	Nitrate	n.a. ²	n.a.	No response to 161–1.2 mM	[52]
	Cadmium	$0.89\mu\mathrm{M}$	$n.d.^3$	•	[50]
	Chromium(III) sulphate	1 μΜ	$0.5\mu M$	Response to 0.5 μM not significant within 60 min	[50]
	Potassium dichromate	n.a.	0.7 and $1.4\mu M$		[50]
	Copper	$0.078\mu\mathrm{M}$	$0.016\mu M$	Response to 0.016 μM not significant within 60 min	[50]
	Cyanide	$0.2\mu\mathrm{M}$	n.d.		[51]
	Bis[tri-n-butyl tin]oxide; TBTO	$0.02\mu M$	n.d.		[52]
	Atrazine	$0.15 \mu\mathrm{M}$	n.d.		[49]
	Lindane	0.17	n.d.		[49]
Marcusenius	Benzocaine and isomers	0.1 mM	n.d.		[61,62]
moorii	MS-222 Sandoz®	0.1 mM	n.d.	= Methanesulphonate of ethyl-m-aminobenzoate	[62]
	Methane sulphonic acid	n.a.	0.1 mM	•	[62]
	Ethyl benzoate	0.1 mM	n.d.		[62]
	Aniline	n.a.	$0.1 \mathrm{mM}$		[62]
	Benzaldehyde	0.1 mM	n.d.	Weak signal	[62]
	p-Aminobenzoic acid	n.a.	0.1 mM		[62]
	Propiophenone	n.a.	0.1 mM		[62]
	n-Butyl acetate	n.a.	0.1 mM		[62]

¹In general, there was no <u>systematic</u> analysis of detection thresholds. The values reported in this column just represent concentrations, where no change in the fish's <u>behaviour</u> was detected after exposure.

Table 3. Detection limits of Gymnotiformes determined by EOD output.

Species	Substance	Lowest detected concentration	Failed to detect ¹	Remarks	Source
Apteronotus albifrons	Cyanide Chlorpromazine	0.5 μM 0.39 μM	n.d. ² n.d.		[47,48] [63]
Gymnotus carapo	Chlorpromazine Chlorpromazine Methylphenidate	$2\mu M \ 0.039\mu M \ 0.11\mu M$	0.39 μM n.d. n.d.	Set-up limitations	[58] [59] [59]
Eigenmannia virescens	Methylphenidate Nalorphine Phenobarbital Morphine Pentylenetetrazole Benzocaine Aniline Ethyl benzoate Caffeine Mephenesin Mescaline	$0.02\mu M$ $0.16\mu M$ Slightly higher than $0.11\mu M$ $0.91\mu M$ $0.91\mu M$ $0.40\mu M$ $0.40\mu M$ $0.40\mu M$ $0.40\mu M$ $0.40\mu M$ $0.50\mu M$	n.d. n.d. 0.11 µM n.d. 18 µM n.d. 0.1 mM n.d. 0.1 mM n.d. n.d.	Set-up limitations Set-up limitations Set-up limitations Set-up limitations Set-up limitations Preliminary results Preliminary results Preliminary results	[60] [60] [60] [60] [61] [61] [61] [61] [61]

¹In general, there was no <u>systematic</u> analysis of detection thresholds. The values reported in this column just represent concentrations, where no change in the fish's <u>behaviour</u> was detected after exposure.

 $^{^{2}}$ n.a. = not applicable.

 $^{^{3}}$ n.d. = not determined.

 $^{^{2}}$ n.d. = not determined.

 $^{^{3}}$ n.a. = not applicable.

higher during night-time [53,67]. However, in the absence of exogenous stress factors, at constant temperature and pH and after acclimatization of the fish, the EOD frequency remains stable enough for reliable measurements [57,64].

5.2. Gymnotiformes

Thomas et al. have employed Apteronotus albifrons (Apteronotidae; 'black ghost knifefish'; see Figure 1b in the online supplementary material) for real-time detection of cyanide [47,48]. Concomitant analysis of EOD frequency and EOD waveform results in a higher reliability of the detection system [48]. Apteronotus albifrons responds to concentrations of potassium cyanide in the range of 0.5– 2.8 µm, usually within the first 30 min (Table 3) [47,48]. The order of magnitude of Δf was about 5–15 Hz (information sparse). At 2.8 µm the fish already showed signs of suffocation [48]. The response times to detect cyanide pollution present a global decreasing trend as cyanide concentrations increase. The sensitivity of both parameters was identical, but waveform changes usually appeared somewhat later (20-560 min). The modality of response (increase/decrease of frequency) to increasing cyanide concentrations and the response time vary individually, but, in the single individual, the time variations of the EOD frequency and the waveform are remarkably stable under constant ambient conditions in this species [63,68–70], and no diurnal oscillations have been observed by Thomas et al. [47], though there are inconsistent statements on this matter [71]. Some other gymnotiform fish display significant day-night changes in their EODs [72–75]. Variations in the pH significantly decrease the EOD frequency (but have no effect on the EOD waveform) [76,77]. In addition, temperature is known to highly influence the EOD frequency in Gymnotiformes [17,68,70,78–80] with a Q_{10} around 1.5 in different species [68,70,78,80] ($\Delta f \approx 40$ –65 Hz [70,80]). Again, no temperature effect on the EOD waveform has been observed in the range of 22–32°C in A. albifrons [70]. Adaptation to chemicals occurs within 60 min in Gymnotiformes [48,59,60], though data are partly conflicting [47]. Further experiments have been performed with Eigenmannia virescens (Sternopygidae) and Gymnotus carapo (Gymnotidae), though not in the context of biomonitoring [58-61] but for evaluation of the mode of action of stimulants and depressants of the central nervous system (Table 3).

6. Discussion and perspective

6.1. Characteristics of experiments employing weakly electric fish

Various aspects of the behaviour or physiological responses of fish have been studied for monitoring water quality [81], e.g. avoidance or rheotactic reactions [82] or ventilation frequency [83–85]. In all these systems either complex

mechanical sensors or highly sensitive electrodes had to be employed to preprocess the behavioural signal to the data analysis system. Moreover, the relatively high rate of false alarms in these systems diminishes the reliability. Weakly electric fish, on the contrary, emit rather strong electric signals that are directly generated by the fish and comparatively easy to detect and process.

Several studies have shown the direct correlation between the EOD frequency and the quality of the surrounding water in weakly electric fish. The biological relevance of these modifications of the EODs caused by toxicants is unknown. Because these fish communicate with each other via their EOD output [21–31,33–37], it is possible that the EOD sensitivity to changes in water quality represents some kind of intraspecific early-warning system to inform other fish of the detected threat. However, such a kind of social behaviour has not been reported in the literature so far, and most experiments do not suggest a correlation between modified EODs and changed locomotion [55,56,60,62,63]. Therefore, the biological relevance has, to date, remained unsolved.

Generally, sensitivity of weakly electric fish is very high and has so far not been tested to its limits in most studies (Tables 1, 2 and 3). Where limits have been determined, longer observation times might lower the given limits in future experiments. On the other hand, adaptation to environmental changes (attenuation) has been observed and will limit the validity of longer observations (sections 2, 5.1 and 5.2). The differences in the reported response times of G. petersii towards copper (4 h [44] vs 15 min [50]) might be due to the improved set-up of Lewis et al. [50]. The most appropriate laboratory set-up has been introduced by Thomas et al. [47], measuring eight individuals at the same time. More importantly, their advanced data analysis (the increase in the number of measured parameters) provides a better way for coping with intraspecific sensitivity variations. In addition, selection of an appropriate temperature also reduces the variance ratio of the EODs [46].

For comparing sensitivity of species, criteria for identifying a response are also important. It has to be noted that the criteria for a positive response were more strict in Thomas *et al.* [47,48] compared with the experiments performed by Lewis *et al.* (e.g. [49–51]). Thomas *et al.* considered a change in the EOD frequency or waveform below or above a certain threshold as a positive response if at least 50% of *A. albifrons* reacted significantly. In contrast, a response was considered positive by Lewis and co-workers, if at least one of three individuals exhibited a significantly modified reaction pattern in a set of experiments after Student t-test and variance analysis.

It remains open as to whether the combination of the highly developed set-up (eight individuals measured at the same time, multiple biological responses evaluated) of Thomas *et al.* [47,48] with *G. petersii* (drawback: interindividual and intra-individual variability; see section 5.1)

would end up with better results, i.e. a higher sensitivity, than the presented combination of this set-up with *A. albifrons* (drawback: 0.2 µm cyanide not tested – no direct comparison possible) [47,48], whose natural advantage is the greater (compared with the tested Mormyridae) interindividual response uniformity (increase/decrease of frequency) to increasing cyanide concentrations (see section 5.2) and temporal stability of the EODs. This has to be elucidated.

In addition, there are a huge number of species in these two orders of fish that have not yet been employed in monitoring measurements to date. Therefore, fish with even higher sensitivity and stability might be discovered in expanded screening experiments.

6.2. Stimulus detection and response

Two aspects in the electrosensory system have to be distinguished:

- 1) The detection of distortions of the self-generated electric field via changes in the potential distribution over the skin surface. The electroreceptor cells in the skin surface constantly measure the local electric voltage between the outside and the inside of the fish, which is proportional to the transepidermal currents passing through the electroreceptor cells. An object with a resistance different from water distorts the self generated electric field and thereby changes the voltage pattern at certain skin areas during discharges of the electric organ. This local modulation of the electric field is called the 'electric image' of an object [19,20,86–88]. The electroreceptor cells synapse on afferent fibres, and the afferent nerve fibres terminate centrally in the electrosensory line lobe in the hindbrain, where the signals are processed [89,90]. The synaptic transmission is mediated either chemically or by means of electrotonic synapses, depending on the kind of receptor cell [90,91]. Local modulations of the electric field are also produced when the fish is exposed to EODs of other fish or externally applied electrical currents [17,92]. Under these conditions, the perceived stimulus is of electro-physical origin and modifies the discharges of the electric organ. The fish might react with, for example, an attack (prey, intruding fish), evasive manoeuvres (obstacles, predators) or mere variation (frequency, duration etc.) of the EODs (communication) [17,28].
- 2) The reaction of the fish upon environmental changes in water quality, by modification of their EODs (e.g. frequency, waveform, duration). As mentioned above, the biological relevance of this phenomenon is unknown. The few authors that have commented on the mode of action assumed that the stimulus is

of **chemical origin** and they predict a direct chemical interaction between the dissolved substances and cells. However, the mechanism by which the various chemical substances (Tables 1-3) might cause the EOD variations has not been investigated in these experiments. Walsh and Schopp [61,62] propose a mechanism of depressive action on the electromotor system of Eigenmannia virescens and Marcusenius moorii. The electromotor system consists of pacemaker and relay neurons in the medulla and electromotor neurons in the spinal cord. The axons of the electromotor neurons form the electric organ that generates the EODs. Benzocaine (a local anaesthetic) and its isomers are thought to depress the pacemaker cells directly and to block the synaptic transmissions of the downstream components, resulting in the observed decrease in the discharge frequency [61,62]. Similarly, Thomas et al. propose a depressing action of cyanide on the central nervous system due to anoxia and subsequent deterioration of the pacemaker cells, but an endocrine action is also conceivable according to the authors [47]. Kunze interprets the EOD variations after atrazine exposure as a response to general stress caused by physiological deteriorations [56]. The proposed modification of EODs as a response to toxin stress, mediated by the endocrine system, is generally plausible, because the influence of the endocrine system on the regulation and modulation of the EODs in weakly electric fish has been shown for cortisol [93], androgens [13,94-106] and estrogens [95,96,105]. However, these modulations of the EODs in living fish occurred within days, not within minutes and hours as reported in the biomonitoring experiments. Short-term modifications of the EODs (within minutes and hours) have, so far, only been reported for the melanocortins, a group of peptide hormones [107] and the monoamine neurotransmitter serotonin [108]. Various mechanisms for the interaction of the endocrine system and the EOD modulation in gymnotiform [15,32,93,102] and mormyrid [13,15,101] fish have been proposed (e.g. modification of the ion channel properties in the cells of the electric organ or manipulation of the medullary pacemaker nucleus).

Another target of chemical interactions may be the receptor cells themselves. *In vitro* experiments with excised skin of *G. petersii* have shown that divalent ions and L-glutamate affect the impulse generation and transmission in artificially electro-stimulated electroreceptors [91].

As mentioned above, the mode of action has not been investigated in detail in the presented biomonitoring studies. The proposed mechanism of direct pacemaker depression/stimulation is plausible for analgesics, depressants and stimulants, but has also not been proven.

In addition, it is more than questionable that this mechanism applies to the whole array of substances tested (Tables 1–3). As outlined above, there is good reason to assume that some of the EOD modifications observed were a response to toxin-induced stress mediated by the endocrine system. It also cannot be ruled out that the modified EODs, at least in some cases, directly reflect electro-physical changes in the medium (altered water conductivity [43,76]).

7. Summary

Weakly electric fish comprise a substantial set of unique and interesting characteristics: (1) the signals are, depending on the experimental conditions, rather large and easy to detect and process; (2) the observed parameter does not represent a passive, eventually long-lasting reaction, which is not under active control of the fish (like wound-healing, response to infections/intoxications), but an immediate active reply to a stimulus (section 5); (3) as long as high toxic concentrations are avoided, an individual fish can be tested more than once (sections 5.1 and 5.2) and each experiment is repeatable with one fish (keeping adaptation times in mind), avoiding intraspecific differences; finally, (4), the repeated investigation of several fish at the same time and the evaluation of several parameters from the recorded EOD transients (section 4) provide a stable system with a good signal-to-noise ratio. All these points render weakly electric fish an especially valuable tool for water monitoring. In addition, these characteristics suggest that weakly electric fish might be a helpful tool in basic research experiments not exclusively related to water monitoring (e.g. pharmacology).

Altogether, 18 publications have been identified, which employed six different species of weakly electric fish and measured the change in their EOD after exposure to dissolved chemicals. Thirty-five chemical substances were investigated. The lowest concentration detected was 20 nm bis[tri-n-butyl tin]oxide (TBTO) using *Gnathonemus petersii* and *Campylomormyrus tamandua*.

The most reliable data have been collected using a set-up that measured eight individuals at the same time and evaluated multiple biological parameters—a set-up mandatory for future experiments. In addition, higher sensitivities might be retrieved by screening the numerous types of weakly electric fish not investigated to date.

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