

Title: **Biodiversity conservation in an uncertain world**

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First paragraph

Despite being key instruments in conservation efforts, protected areas are vulnerable to risks associated with weak enforcement and governance¹, pressure from land-use intensification², and climate change³ - any of which can reduce their effectiveness. Failure to consider such risk factors in planning diminishes the ability of protected areas to contribute to international biodiversity goals⁴. We included these three categories of risk into a multi-objective spatial optimization designed to expand the existing protected area estate, aiming to conserve at least 30% of the distribution of all the world's terrestrial vertebrates. The total global terrestrial area requiring protection was the same (3.57 million km²) across the three risk scenarios and required remarkably small (1%) increases in the amount of land protected relative to ignoring risk. Among the three risk categories, governance drove the greatest variation in the location of land prioritized. In particular, conserving wide-ranging species required countries with relatively strong governance to protect more land when bordering nations with comparatively weak governance. Our results both underscore the need for cross-jurisdictional coordination and demonstrate how risk can be efficiently incorporated into global planning efforts.

Main text

Protecting habitat is one of the best strategies for stemming the alarming decline in biodiversity⁵. As such, the cornerstone of the new global framework for biodiversity conservation is to protect at least 30% of terrestrial land area by 2030⁴. Most current approaches to identifying the most important areas to protect rely upon estimations of the conservation value of the land for biodiversity and the threats it faces^{4,6,7}. Seldom articulated in plans is the tacit assumption that protection is enforced, effective, and permanent, yet we know that many protected areas are subject to risks from weak governance, land use intensification, and climate change. For example: the quality of governance relates to investment in conservation^{8,9}; political instability and corruption can reduce protected area effectiveness¹⁰; protected

areas with more deforestation are at high risk of degazetting and failure to meet protection goals¹¹; and increased extreme weather events cause population decline or extirpation for a variety of species¹². Thus, to make effective use of limited conservation resources, planning for investment in protected areas must account for these risks^{13,14}. Here we demonstrate how accounting for governance, land-use, and climate risks can influence decisions for establishing protected areas at a global scale and may ultimately improve the resilience of protected areas and the species they support.

We defined the following three broad categories of risk, which we considered to be factors likely to diminish the long-term effectiveness of protected areas: i) governance, ii) land-use, and iii) climate. For governance risk, we used a national-scale metric that combines six governance indicators from the World Bank¹⁵: accountability, political stability, government effectiveness, regulatory quality, rule of law, and control of corruption (Figure S1). For land-use risk, we estimated the average change in biodiversity per land-use category using methods¹⁶ that model the risk of biodiversity loss for land systems due to agricultural expansion and intensification (Figure S2). For climate risk, we used the duration of extreme heat events, calculated using a probabilistic framework that estimates the novelty of temperatures relative to historical year-to-year variation from 1979 to 2019 (Figure S3), identifying areas where heat events are likely to have the most significant effects on biodiversity¹⁷. Although we used these three risk categories for illustrative purposes, the approach we propose is flexible and can easily incorporate other risk metrics with implications for biodiversity¹⁸.

We considered the influence of risk categories on allocating protection decisions at a global scale for all 30,930 known distributions of vertebrate species from the IUCN Red List of Threatened Species¹⁹ using a multi-objective optimization approach. To incorporate risk categories, we built on a classical problem formulation from the systematic conservation planning literature – the minimum set problem - where the objective is to reach species distribution protection targets, while accounting for one constraint such as land cost or area^{20–22}. We expand this approach to include multiple objectives accounting for varying risk in the problem formulation, by treating each risk layer as a separate

objective in the problem formulation²³. We use a hierarchical approach that assigns a priority to each objective, and optimizes for the objectives in decreasing priority order. At each step, the approach finds the best solution for the current objective, but only from among those that would not degrade the solution quality for higher-priority objectives.

In total 16 planning scenarios were created, such that solutions accounted for all possible combinations of risk categories within each hierarchical level (Table S1). We then compared these risk-based solutions to those produced with a null scenario that adopted the traditional area-minimizing approach to optimization without considering risk²⁴. Because our scenarios aimed to build upon the current protected area portfolio globally, we incorporated current protected areas into our solutions. For each scenario we protected 30% of the range of all vertebrate species, which is broadly analogous to the more general 30% total area Convention on Biological Diversity (CBD) target²⁴.

Surprisingly, scenarios that incorporated the three risk categories required only 0.9% more global area on average (0.34 – 1.14 %) than the null scenario to meet the target of protecting 30% of vertebrate ranges. Thus, accounting for risks cost relatively little compared to the potential gains from having a more resilient conservation network (Figure 1). Notably, the target of protecting 30% of each vertebrate species range was achieved by all 16 scenarios without exceeding the post-2020 CBD target²⁴. When only looking at scenarios that included one risk factor, land-use risk required the greatest increase in global protected area, compared to scenarios only including governance and/or climate extreme risks (Table S1).

Protected areas identified across scenarios overlapped spatially, with the same 22 million km² (6.9% of global land area) being prioritized for expansion of the current protected area system in at least eleven scenarios and 3.6 million km² (2.4% of global land area) in all fifteen risk scenarios (Figure 2). These “no regrets” areas provide examples of places that should be priorities for international agencies aiming to maximize the resilience of protected area networks, as they are robust to assumptions of the relative importance of risk factors. Example countries that have contiguous areas of

high overlap among scenarios are Canada, Egypt, Finland, Kazakhstan and Peru (Figure S4). There is some overlap among the priorities across scenarios within Conservation International's global biodiversity hotspots²⁵, but many high overlap areas lie either outside these hotspots (83.1%) or occur within small portions of the biodiversity hotspots, likely because these areas are important to protect regardless of future risk (Figure S5).

We also found variation in the locations requiring protection when risks were introduced (Figure 3; Table S2). These differences were driven largely by governance (Figure S6). Countries with relatively high governance scores had greater area requiring protection under risk scenarios relative to the null scenario, especially when species were wider ranging and when neighbouring countries had low governance scores. Thus, risk is connected across jurisdictions, where planning scenarios favour protection of species in nearby countries with low governance risk (i.e., high governance scores). For example, many vertebrate species ranges span northeastern Russia, Finland, and Sweden, with one of the most iconic being caribou (*Rangifer tarandus*), which has an IUCN conservation status of vulnerable. Because Russia suffers from low scores for 'voice and accountability, rule of law, and control of corruption' (Table S3), whereas Finland and Sweden have relatively high governance scores, the scenarios including governance pressures led to a selection of 99.4% and 48.9% of Finland and Sweden's land areas respectively compared to the null scenario with 30.8% and 15.2% (Figure 4). These results do not mean the majority of land inside Finland needs to be protected to ensure the long-term persistence of caribou, but indicate that prioritizing areas in Sweden and Finland is predicted to be far less of a risk than areas in Russia.

Land-use and climate change also influenced variation in areas requiring protection above null. For example, large areas of Sierra Leone are experiencing high risk of biodiversity loss due to expanding intensive land-use practices (Fig. S2), whereas this same risk is lower in neighbouring Liberia. Scenarios including land-use risk selected 50.1% of the land area in Liberia compared to 21.9% in the null scenario (Figure 4). Large areas of Algeria are experiencing increasingly frequent

extreme heat events (Fig. S3), whereas neighbouring Libya is not experiencing as many extreme heat events. Scenarios including climate impact risk selected of 30.8% of Libya's land area compared to the null scenario with 20.9% (Figure 4).

These results emphasize the importance of coordinated cross-jurisdictional conservation planning initiatives²⁶ and identify countries where opportunities for collaboration would yield more resilient protected areas. To illustrate this point, we consider the Great Green Macaw (*Ara ambiguus*), with <2500 individuals remaining²⁷ and a range that stretches from southern Honduras to western Colombia. Because Great Green Macaw habitat spans several countries differing in governance, land use, and climate risk, coordinated efforts among countries will be necessary for the species to persist in the future. For countries with a predominance of wide-ranging species whose ranges will be impacted by varying climate, land-use, and governance risk across borders, conservation projects can focus on cooperative governance frameworks²⁸ (Figure 3). These governance frameworks, both within and between countries, would need to be developed in an environmentally just and equitable way to deliver benefits to biodiversity and local communities⁴⁹.

In contrast, some countries at high risk from climate change, land-use, and low governance scores, but with high endemism, for example in northern Oceania, have a low difference between scenarios that incorporate risk. Given high and endemic biodiversity, and homogeneity of risk, these countries all require high rates of protection within their borders. Moreover, some countries closer to reaching the CBD's 30% land area protection target, for example Brazil, which already has 30.3% of its land area protected, had lower differences between scenarios that incorporate risk and the null scenario that does not incorporate risk, despite having high climate, land-use, and governance risk. This outlines the importance of further considering the effectiveness of existing protected areas in planning analyses, where pressure from cropland conversion in tropical protected areas has increased to similar rates outside protected areas²⁹.

Previous work has incorporated individual risk factors analogous to those we used, including governance^{1,48}, climate change³ and land-use change^{2,30}. Yet, our results show that protected area expansion decisions can be profoundly influenced by all three risk factors combined. If data on risk alters the effectiveness of biodiversity protection, our results show that they should be used together to support decisions for resilient protected area networks. As an example, climate metrics such as disappearing climates³¹ might be relevant if the consideration is on small-ranged and threatened species. Our flexible framework and methods can allow conservation agencies looking to set priorities from the global to local scale and incorporate different metrics to explore the influence of individual parameters and metrics on decisions.

Conclusion

The conservation community has traditionally neglected to estimate how future changes in climate³², land-use³⁰, and socio-economic conditions might compromise the effectiveness of protected areas. Our results show that the spatial distribution of protected areas, rather than the land area *per se*, can be profoundly influenced by risk, particularly from governance. Surprisingly, incorporating risk into decision-making adds <1% to the total global area required to meet biodiversity targets. Accounting for risk comes at limited extra cost, but potentially large benefits to achieving global biodiversity targets. Our results also emphasize the importance of cross-jurisdictional conservation initiatives, especially in adjacent countries sharing wide-ranging species where risk varies considerably from country to country. Considering risk in conservation decision-making is more likely to result in more resilient and effective conservation plans into the future to help safeguard our planet's biodiversity in the face of the current extinction and climate crises.

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Figure legends (+ figures)

Figure 1: Spatial representation of priority areas for protection to account for governance, land use and climate risk. Accounting for these risks to protected area effectiveness to produce more resilient conservation networks would require 29.17% of land surface to protect 30% of threatened species' ranges.

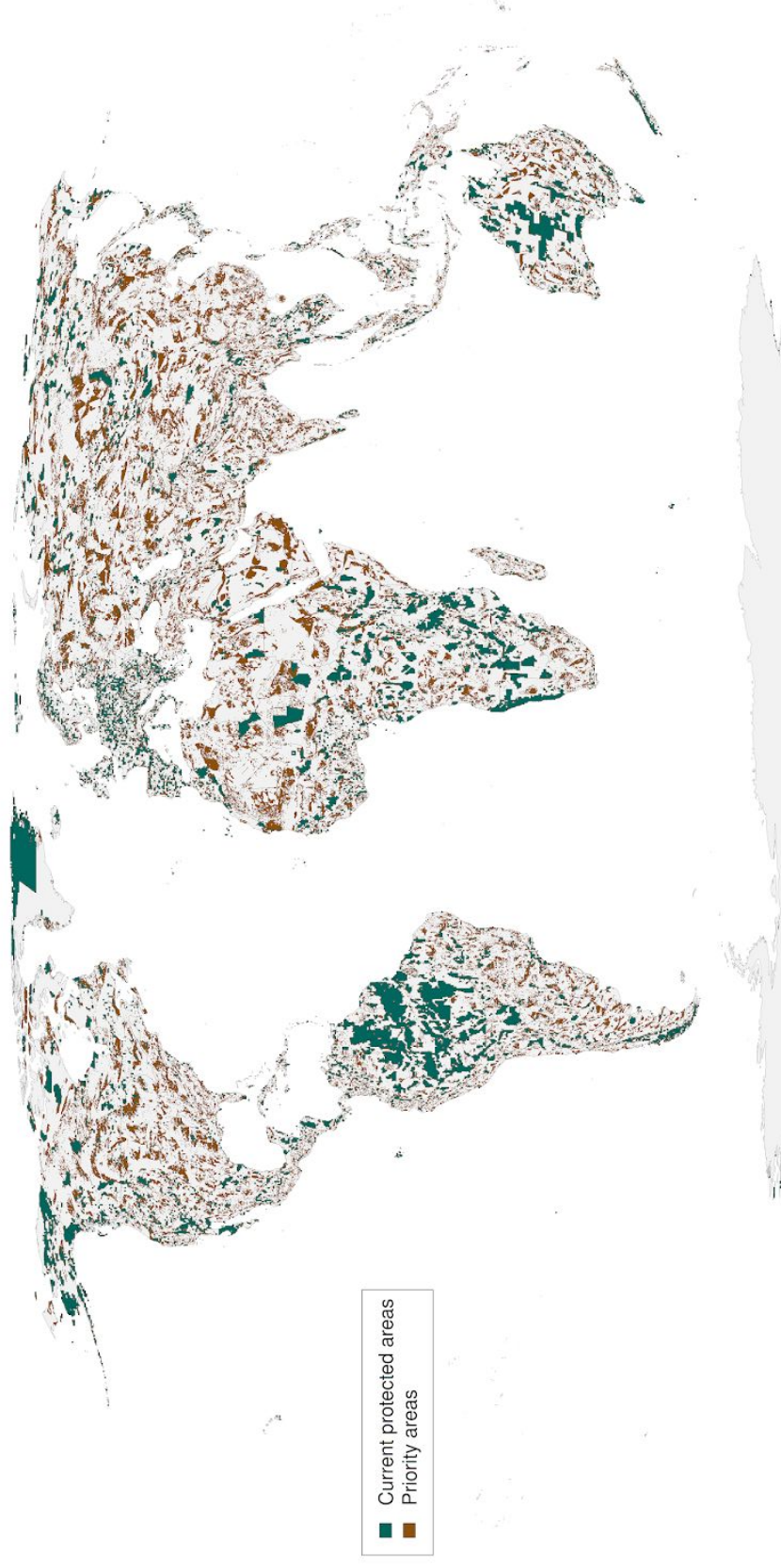


Figure 2: “No regrets” areas comprising 3.57 million km² of land that was identified as priority habitat for protection regardless of the risks included in our analysis.

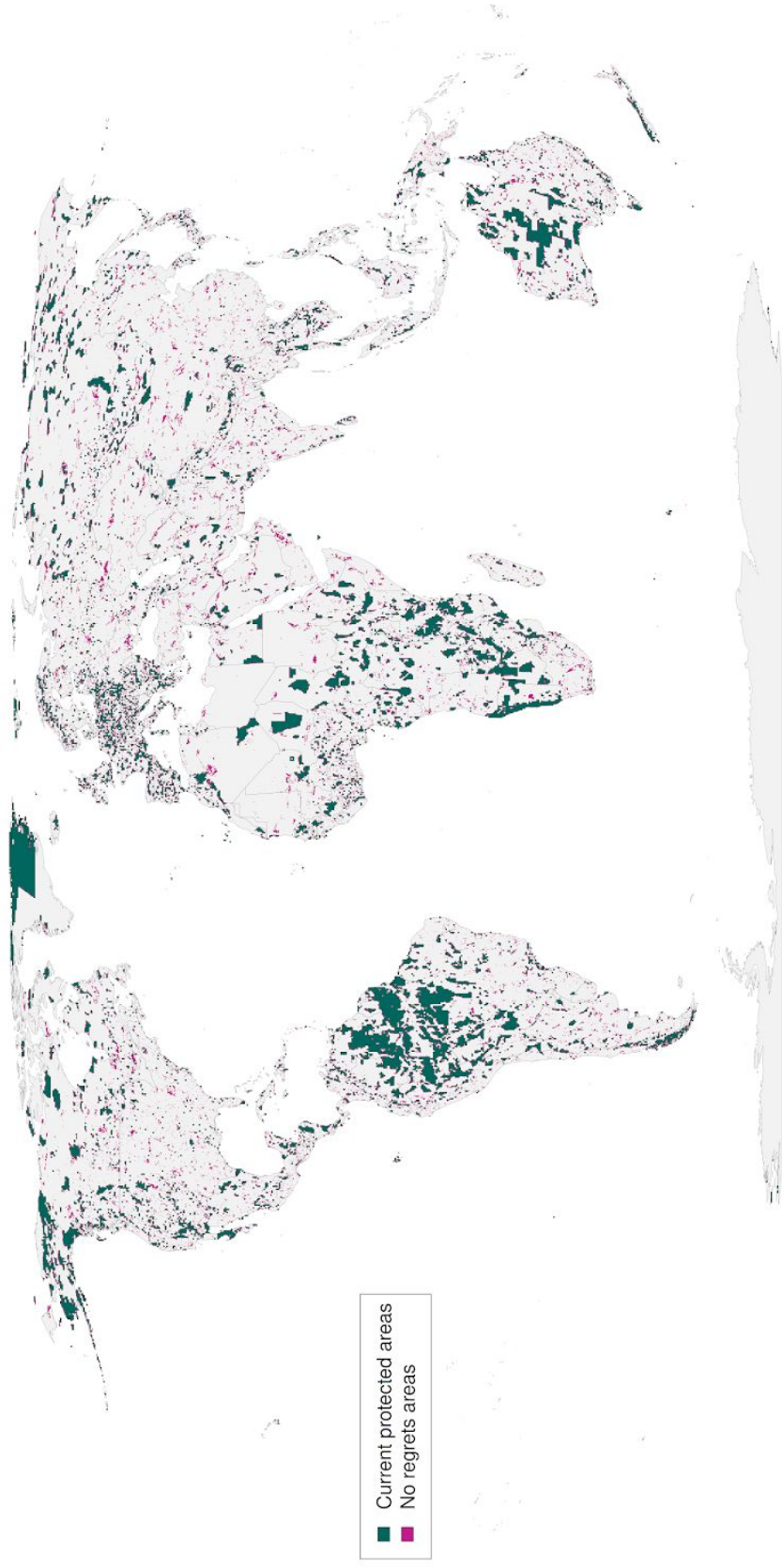


Figure 3: Percent country-level variation between the null scenario and the 15 scenarios including risk. Countries whose results are consistent across the 15 scenarios (e.g. Brazil) have low variation, while countries whose results are less consistent across the 15 scenarios have high variation (e.g. Sweden). The kmeans method⁵⁰ was used to generate class intervals for visualization.

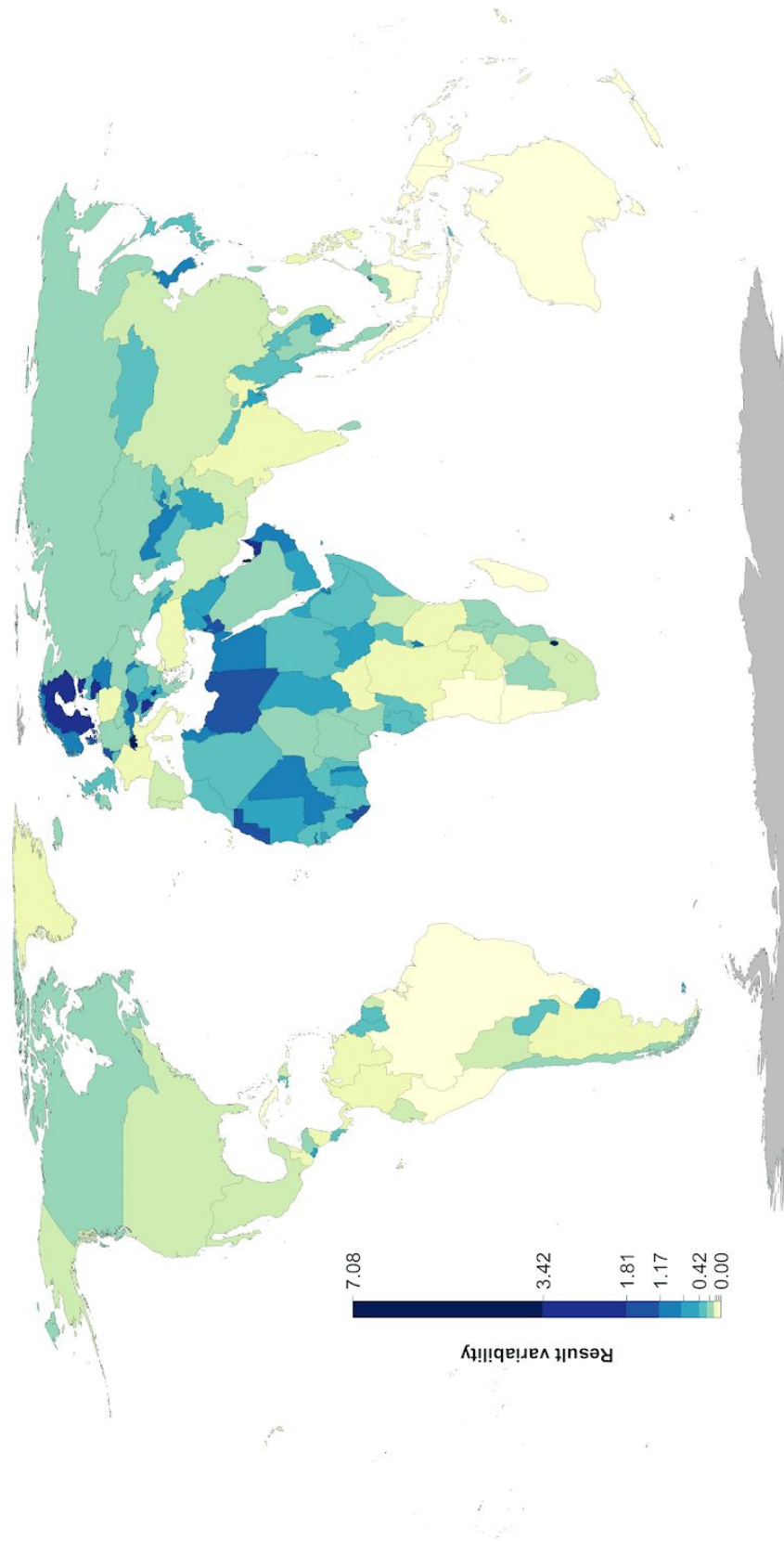
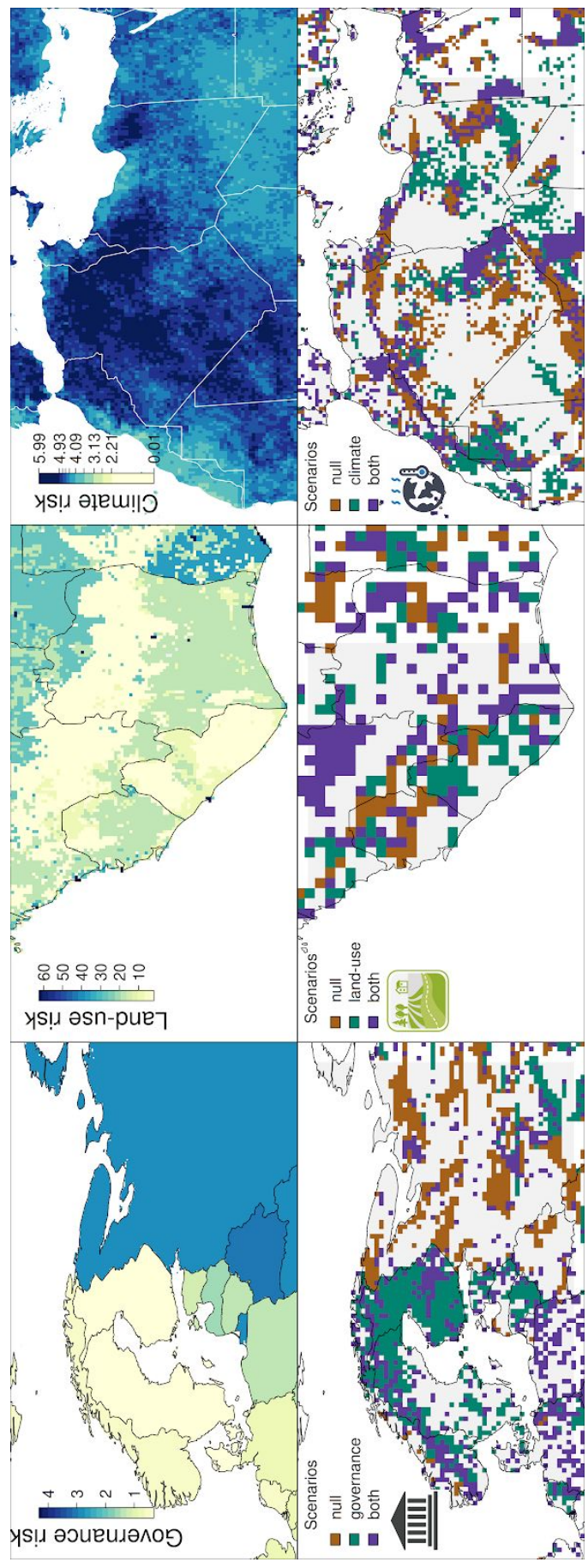


Figure 4: Contrast of using individual risk objectives (governance, land-use, climate) to the null scenario of uniform objective structure. The top panels represent the individual risk data for the focal regions. In the bottom panels brown shows null, green the specific risk objective scenario results, and purple where both scenarios agree. The figures show how the spatial configuration of the solutions changes when risk is considered in a scenario. Governance focus is on Sweden, Finland and Russia, land-use risk on Sierra Leone and Liberia, and climate risk on Algeria and Liberia.



Methods

We used a multi-objective optimization approach that incorporated governance, land use and climate constraints to prioritize the conservation of 30,930 vertebrate species. All scenarios we investigated assumed the current global protected area portfolio is locked in. We further set a target to protect 30% of the range of each species, which is broadly analogous to current CDB discussions on post-2020 biodiversity targets⁴.

Species selection

Our species list included all terrestrial vertebrate species from the IUCN Red List of threatened species, following Pouzols et al.². For mammal, amphibian and reptile species ranges, we used the IUCN Red List website (<http://www.iucnredlist.org/>, accessed 2019-11-14) and for birds we used the BirdLife International data zone webpage (<http://www.birdlife.org/datazone/home>, accessed 2019-11-14). We used these taxa because no analogous data are available for a high proportion of species in other taxonomic groups such as insects³³. These data have certain limitations, including possible underestimation of the extent of occurrence and overestimation of the true area of occupancy², but they have been shown to be robust to commission errors as long as the focus is on species assemblages rather than single species⁷. They are currently the most frequently used and updated source for vertebrate species distributions³⁴.

For each taxonomic group, we restricted our analysis to species that fell into the presence category of ‘Extant’, the origin categories of ‘Native’ or ‘Reintroduced’ and the seasonality categories ‘Resident’, ‘Breeding Season’ or ‘Non-breeding Season’, thus only focusing on stationary periods of the life cycle of migratory species. This resulted in the following final numbers of amphibian, bird, mammal and reptile species ranges: 5660, 13375, 5442, and 6153, respectively.

Basic administrative delineations

National boundaries were derived from the Global Administrative Areas database (<http://gadm.org/>, accessed 2019-10-31). We obtained protected area boundaries from the World Database on Protected Areas (WDPA, <https://www.protectedplanet.net>). Following standard procedures for cleaning the protected area dataset^{35,36}, we (i) projected the data to an equal-area coordinate system (World Behrman), (ii) excluded reserves with unknown or proposed designations, (iii) excluded UNESCO Biosphere Reserves³⁷, (iv) buffered sites represented as point localities to their reported area, (v) dissolved boundaries to prevent issues with overlapping areas, and (vi) removed slivers (code available at <https://github.com/jeffreyhanson/global-protected-areas>). After the protected area data were modified as described above, we overlaid the protected area boundaries with a 10 x 10 km grid covering the Earth. These spatial data procedures were implemented using ArcMap (version 10.3.1) and python (version 2.7.8).

Governance risk

Conservation risk due to governance can affect the outcomes of strategies, and effective governance can promote the resilience of conservation in the face of sociopolitical and economic shocks. We used worldwide governance indicators from the World Bank¹⁵ to capture these pressures. The indicators include six scaled measures: voice and accountability; political stability and absence of violence; government effectiveness; regulatory quality; rule of law; and control of corruption (see Table S4 for definitions). We chose these indicators because evidence suggests that they reliably predict protected area effectiveness³⁸ and state investment and efforts for biodiversity conservation⁸. For each country, we used a mean of annual averages of all six measures⁸ (Figure S1).

Land use risk

We used a recently developed global land systems map produced by Kehoe et al.¹⁶ to incorporate the risk of land-use change. This map is based on a global land systems map for the year

2000³⁹ at a 9.25 km² spatial resolution, but is refined based on recent land-cover and land-use datasets to a spatial resolution of 1 km². Kehoe et al.¹⁶ further estimated the impact of land use and land use intensity on biodiversity, with data originating from the PREDICTS project⁴⁰. They first matched their land-systems classes to varying intensity levels for each land-use type (for detailed conversion table, see ref⁴¹). This allowed Kehoe et al.¹⁶ to calculate average biodiversity loss per land system (relative to an unimpacted baseline) by taking the mean model estimates of biodiversity loss per land-use intensity class from previous work⁴¹. The result gives average relative biodiversity gain or loss per land-system class. Here, we used their modelled mean estimates (following Newbold et al.⁴¹) of relative percent biodiversity change for each land-system class for species abundance as a measure of the land-use pressure (Figure S2).

Climate risk

Anthropogenic climate change is affecting the frequency and duration of extreme heat events (AghaKouchak et al. 2020; Diffenbaugh et al. 2017). Exposure to these events can adversely affect human populations (Anderson and Bell Michelle 2011; Battisti and Naylor 2009; Guo et al. 2017; Mitchell et al. 2016) and natural systems (Harris et al. 2018; Maxwell et al. 2019). For species in natural systems, these events can further the decline and extirpation of populations, increasing the chances of extinction (Maron et al. 2015; Maxwell et al. 2019). EHE and ECE can also promote the formation of novel ecosystems (Harris et al. 2018), generate enhanced selection pressures (Grant et al. 2017; Gutschick and BassiriRad 2003), and change the phenology of life history events (Cremonese et al. 2017; La Sorte et al. 2016). There are a number of climate indices that have been used to estimate the occurrence of these events (Fenner et al. 2019; Smith et al. 2013). These indices are often context specific and there is little consensus on the most appropriate technique (McPhillips et al. 2018).

We estimated climatic risk based on the estimated trend in the annual proportion of days containing extreme heat events from 1979 to 2019¹⁷. Extreme heat events were estimated using hourly

air temperature at 2 m above the surface and gridded at a 31 km (0.28125° at the equator) spatial resolution (DOI: 10.24381/cds.adbb2d47). The temperature data was acquired from the European Centre for Medium-Range Weather Forecasts (ECMWF) fifth generation atmospheric reanalysis of the global climate (ERA5)^{42,43}. The approach first extracted daily minimum and maximum temperature for each grid cell over the 41-year period. To reduce the influence of warming trends, the daily minimum and maximum temperature was then detrended across years for each day and grid cell using empirical mode decomposition (EMD)^{44,45}. The occurrence of extreme heat events was estimated using the following approach: The detrended minimum and maximum temperature data was treated as normally distributed across years for each day and grid cell. The probability density function for the detrended minimum and maximum temperature was then estimated using the mean and standard deviation calculated across years for each day and grid cell. Extreme heat events occurred when the probabilities for both minimum and maximum temperature on a given day and grid cell were within the 0.95-1.00 quartile of the probability density function. The trend in the annual proportion of days containing extreme heat events for each year was calculated for each grid cell using beta regression with a logit link function and an identity function in the precision model^{46,47}. (Figure S3). See La Sorte et al.¹⁷ for additional details.

Multi-objective optimization of pressure reduction

We processed all data described previously to a 10 x 10 km resolution and clipped data to the extent of land based on the global administrative areas database. We then developed an extension on the minimum set problem, which has the goal to identify a set of sites within a planning area that represents all conservation targets in the fewest number of sites²¹. Instead of including one objective we expanded the formulation to include multiple objectives in the problem formulation. We used a hierarchical or lexicographic approach that assigns a priority to each objective, and optimizes for the objectives in decreasing priority order. At each step, it finds the best solution for the current objective,

but only from among those that would not degrade the solution quality for higher-priority objectives. We considered up to three objectives in our prioritization scenarios, i) governance risk, ii) land-use risk, and iii) climate risk. To compare different scenarios, we calculated solutions for each unique objective combination ($n = 15$), as well as one where we use a constant objective function as the null scenario, as the order of the hierarchy can influence the results.

In systematic conservation planning, conservation features describe the biodiversity units (e.g. species, communities, habitat types) that are used to inform protected area establishment. Planning units describe the candidate areas for protected area establishment (e.g. cadastral units). Each planning unit contains an amount of each feature (e.g. presence/absence, number of individuals). A prioritization describes a candidate set of planning units selected for protected establishment. Each feature has a representation target indicating the minimum amount of each feature that ideally should be held in the prioritization (e.g. 50 presences, 200 individuals). To minimize risk, we have a set of datasets describing the relative risk associated with selecting each planning unit for protected area establishment. Thus we wish to identify a prioritization that meets the representation targets for all of the conservation features, with minimal risk.

Let I denote the set of conservation features (indexed by i), and J denote the set of planning units (indexed by j). To describe existing conservation efforts, let p_j indicate (i.e., using zeros and ones) if each planning unit $j \in J$ is already part of the global protected area system. To describe the spatial distribution of the features, let A_{ij} denote (i.e., using zeros and ones) if each feature is present or absent from each planning unit. To ensure the features are adequately represented by the solution, let T_i denote the conservation target for each feature $i \in I$. Next, let D denote the set of risk datasets (indexed by d). To describe the relative risk associated with each planning unit, let R_{dj} denote the risk for planning units $j \in J$ according to risk datasets $d \in D$.

The problem contains the binary decision variables x_j for planning units $j \in J$.

$$x_j = \begin{cases} 1, & \text{if } j \text{ selected for prioritisation,} \\ 0, & \text{else} \end{cases} \quad (\text{eqn 1a})$$

The reserve selection problem is formulated following:

$$\text{lexmin } f_1(x), f_2(x), \dots f_D(x) \quad (\text{eqn 2a})$$

$$\text{subject to } f_d(x) = \sum_{j \in J} R_{dj} X_j \quad \forall d \in D \quad (\text{eqn 2b})$$

$$\sum_{j \in J} A_{ij} \geq T_i \quad \forall i \in I \quad (\text{eqn 2c})$$

$$x_j \geq p_j \quad \forall j \in J \quad (\text{eqn2d})$$

$$x_j \in \{0, 1\} \quad \forall j \in J \quad (\text{eqn 2e})$$

The objective function (eqn 2a) is to lexicographically (hierarchically) minimize multiple functions. Constraints (eqn 2b) define each of these functions as the total risk encompassed by selected planning units given each risk dataset. Constraints (eqn 2c) ensure that the representation targets (T_i) are met for all features. Constraints (eqn 2d) ensure that the existing protected areas are selected in the solution. Finally, constraints (eqns 2e) ensure that the decision variables x_j contain zeros or ones.

For all scenarios we locked in current protected areas and used the same feature set of 30,930 vertebrates. The target for each feature was set to 30% of their range. The optimality gap, which specifies how far from numerical optimality we would allow the solution to be, was 10% for each

objective in the hierarchy. We chose a 10% optimality gap to allow for some flexibility in the result of each step in the hierarchy to avoid getting too restricted in the solution space.

Methods references

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Author contributions

Competing interest declaration

No competing interests to declare

Data availability

All data, scripts and full results are available on Open Science Framework (OSF) and will be assigned a DOI once the manuscript is in print:

https://osf.io/e2fuw/?view_only=46eb2e525daf42d29df318a92762d885

Table S1. Scenarios explored and global protection results. The risk factor order represents the order risk factors were included in the hierarchical prioritization. (G = governance, L = land use, C = Climate).

| Scenario | Risk factors included | Global land area protected [%] |
|-------------|-----------------------|--------------------------------|
| null | - | 28.08 |
| 1 | G | 28.44 |
| 2 | L | 29.23 |
| 3 | C | 28.45 |
| 4 | G > L | 29.2 |
| 5 | L > G | 29.23 |
| 6 | G > C | 28.43 |
| 7 | C > G | 28.45 |
| 8 | L > C | 29.19 |
| 9 | C > L | 29.19 |
| 10 | G > L > C | 29.17 |
| 11 | G > C > L | 29.2 |
| 12 | L > G > C | 29.19 |
| 13 | L > C > G | 29.19 |
| 14 | C > G > L | 29.2 |
| 15 | C > L > G | 29.19 |

Table S2. Country specific results for the 15 scenarios investigated. Numbers represent % of land area of a country selected.

(As an example 5 countries included here, full list in csv. N = null, G = governance, L = land use, C = Climate)

https://drive.google.com/file/d/1eD4y4K8XG4nrxnRL5fNtiTqzuqfIJ_DfB/view?usp=sharing

| | Afghanistan | Akrotiri and Dhekelia | Åland | Albania | Algeria |
|-----|-------------|--------------------------|-------|---------|---------|
| N | 22.94 | 50 | 0 | 18.75 | 32.66 |
| G | 15.76 | 50 | 91.67 | 17.71 | 20.52 |
| L | 26.58 | 50 | 33.33 | 21.88 | 30.02 |
| C | 24.82 | 50 | 0 | 22.92 | 22.53 |
| GL | 26.54 | 50 | 0 | 22.22 | 29.87 |
| LG | 26.65 | 50 | 25 | 21.88 | 29.97 |
| GC | 15.93 | 50 | 75 | 17.36 | 20.66 |
| CG | 24.68 | 50 | 0 | 22.92 | 22.52 |
| LC | 26.72 | 50 | 16.67 | 21.18 | 30.02 |
| CL | 26.52 | 50 | 0 | 21.88 | 30.44 |
| GLC | 26.51 | 50 | 0 | 21.18 | 29.83 |
| GCL | 26.52 | 50 | 0 | 21.88 | 30.49 |
| LGC | 26.71 | 50 | 25 | 20.83 | 29.99 |
| LCG | 26.72 | 50 | 16.67 | 21.18 | 30.02 |
| CGL | 26.54 | 50 | 0 | 21.88 | 30.19 |
| CLG | 26.52 | 50 | 0 | 21.88 | 30.44 |

Table S3. Governance risk score table (see csv)

https://drive.google.com/file/d/1g_LePBfCbphXzTiCOXCzQtNLSSYoV6me/view?usp=sharing

Table S4.

| Indicator | Definition |
|---|---|
| | Source: World Bank, 2020 https://datacatalog.worldbank.org/dataset/worldwide-governance-indicators) |
| Voice and accountability | “Voice and accountability captures perceptions of the extent to which a country's citizens are able to participate in selecting their government, as well as freedom of expression, freedom of association, and a free media.” |
| Political stability and absence of violence | “Political Stability and Absence of Violence/Terrorism measures perceptions of the likelihood of political instability and/or politically-motivated violence, including terrorism.” |
| Government effectiveness | “Government effectiveness captures perceptions of the quality of public services, the quality of the civil service and the degree of its independence from political pressures, the quality of policy formulation and implementation, and the credibility of the government's commitment to such policies.” |
| Regulatory quality | “Regulatory quality captures perceptions of the ability of the government to formulate and implement sound policies and regulations that permit and promote private sector development.” |
| Rule of law | “Rule of law captures perceptions of the extent to which agents have confidence in and abide by the rules of society, and in particular the |

quality of contract enforcement, property rights, the police, and the courts, as well as the likelihood of crime and violence.”

Control of
corruption

“Control of corruption captures perceptions of the extent to which public power is exercised for private gain, including both petty and grand forms of corruption, as well as "capture" of the state by elites and private interests.”

Figure S1. Governance risk (yellow = low, blue= high)

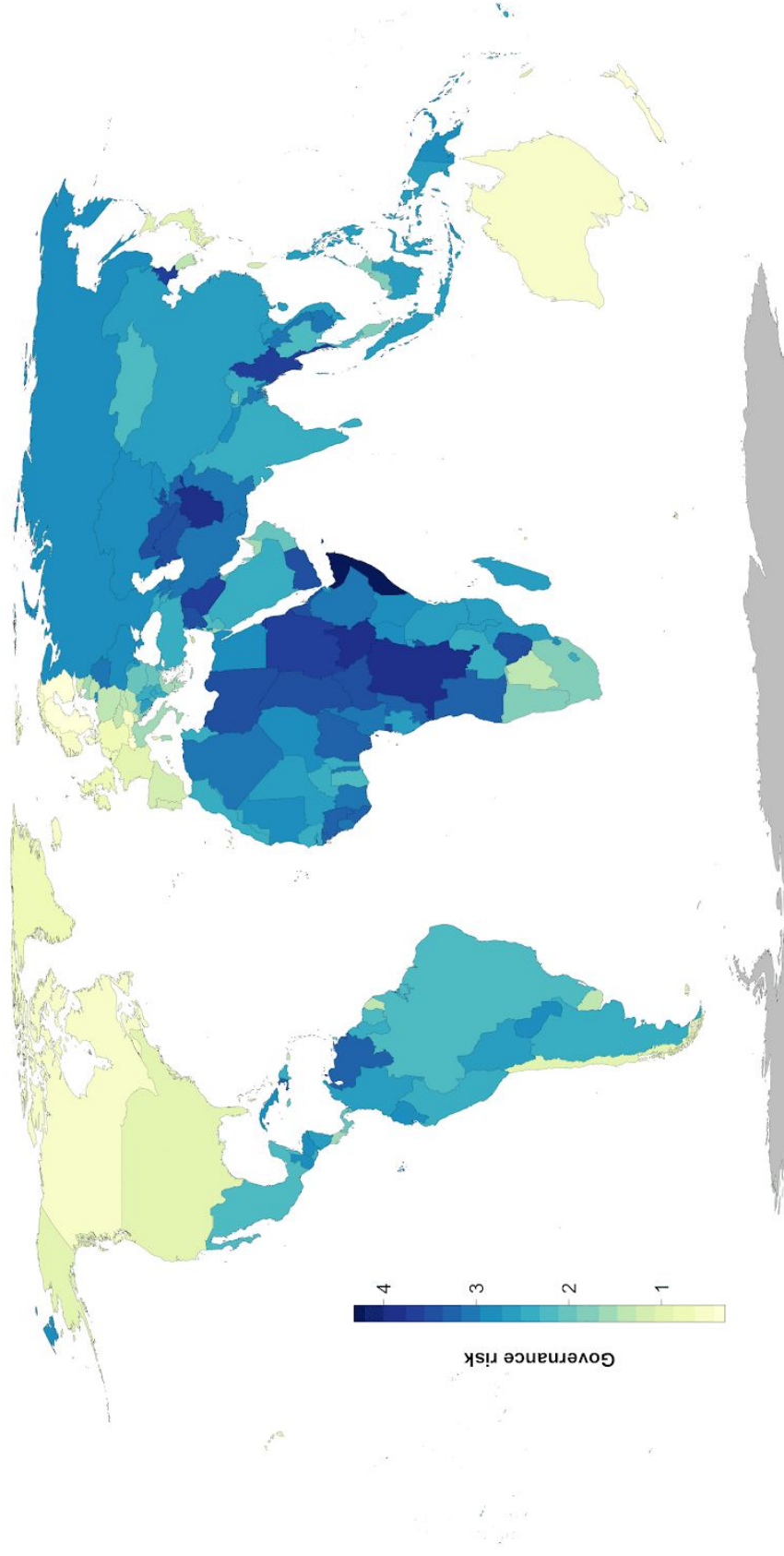


Figure S2. Land systems risk (yellow = low, blue= high)

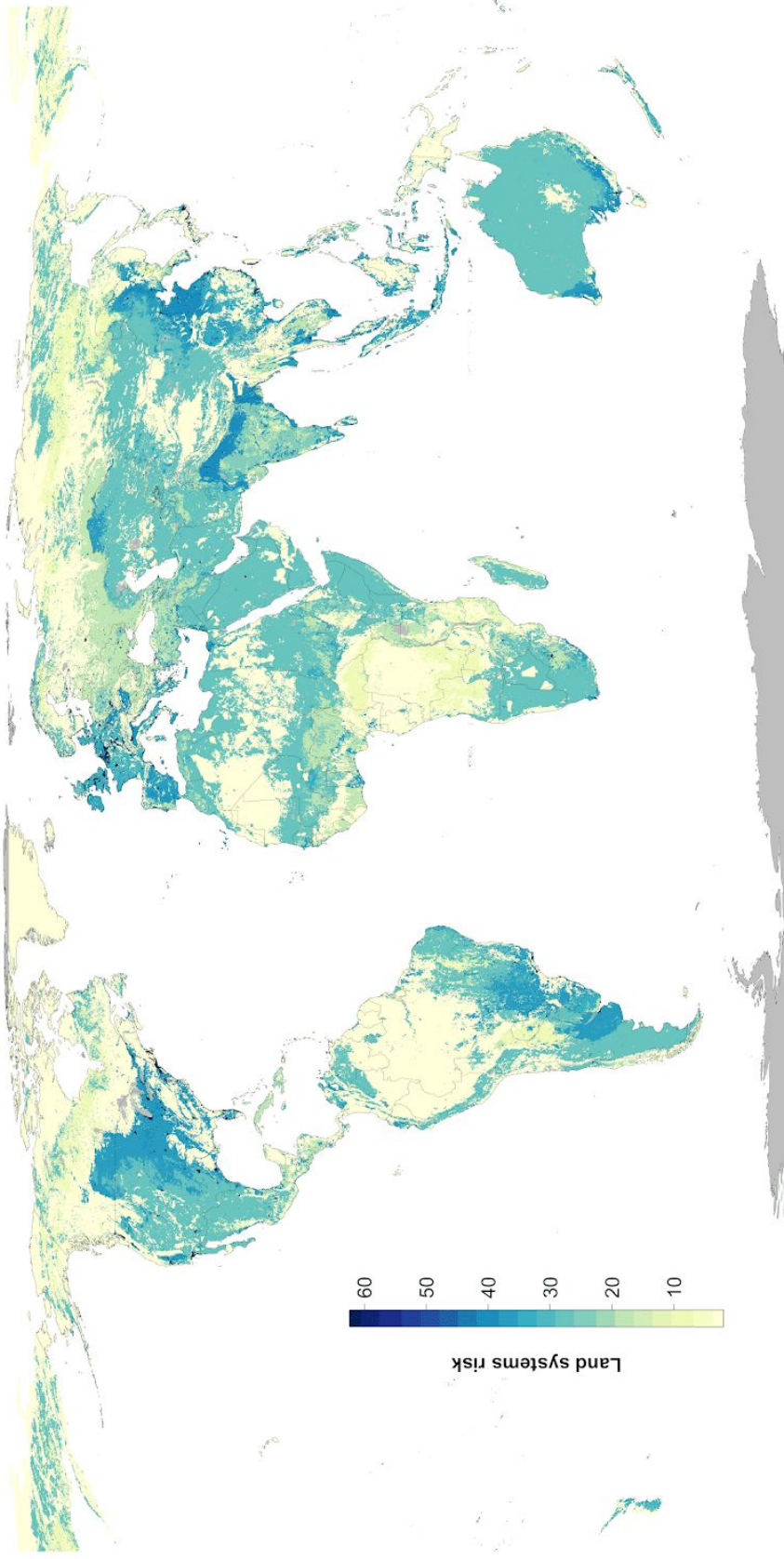


Figure S3. Climate risk (extreme heat events) (yellow = low, blue= high)

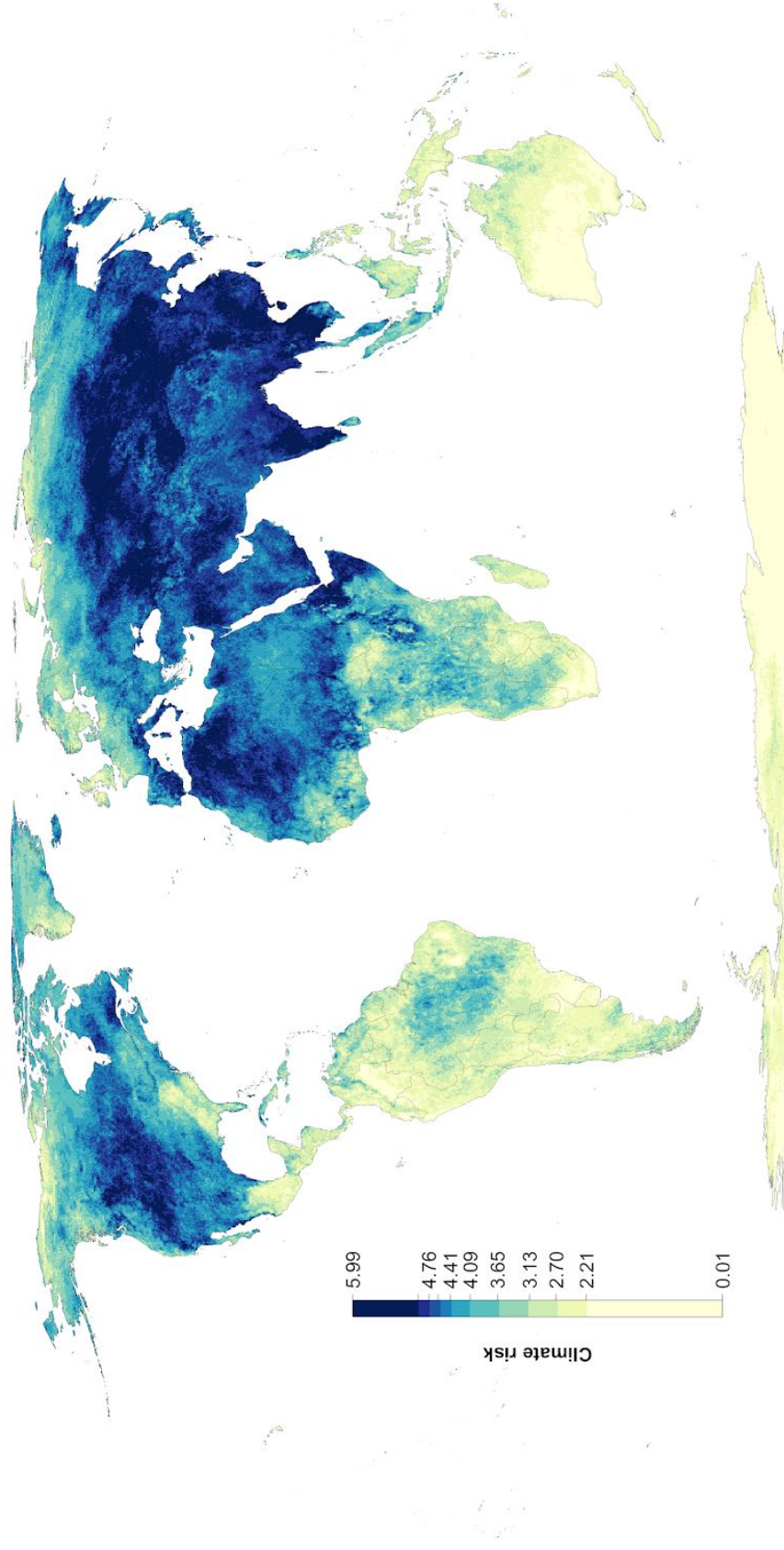


Figure S4: Scenario overlap. green = protected areas. Color gradient from yellow (one scenario) to red (15 scenarios) = overlap.

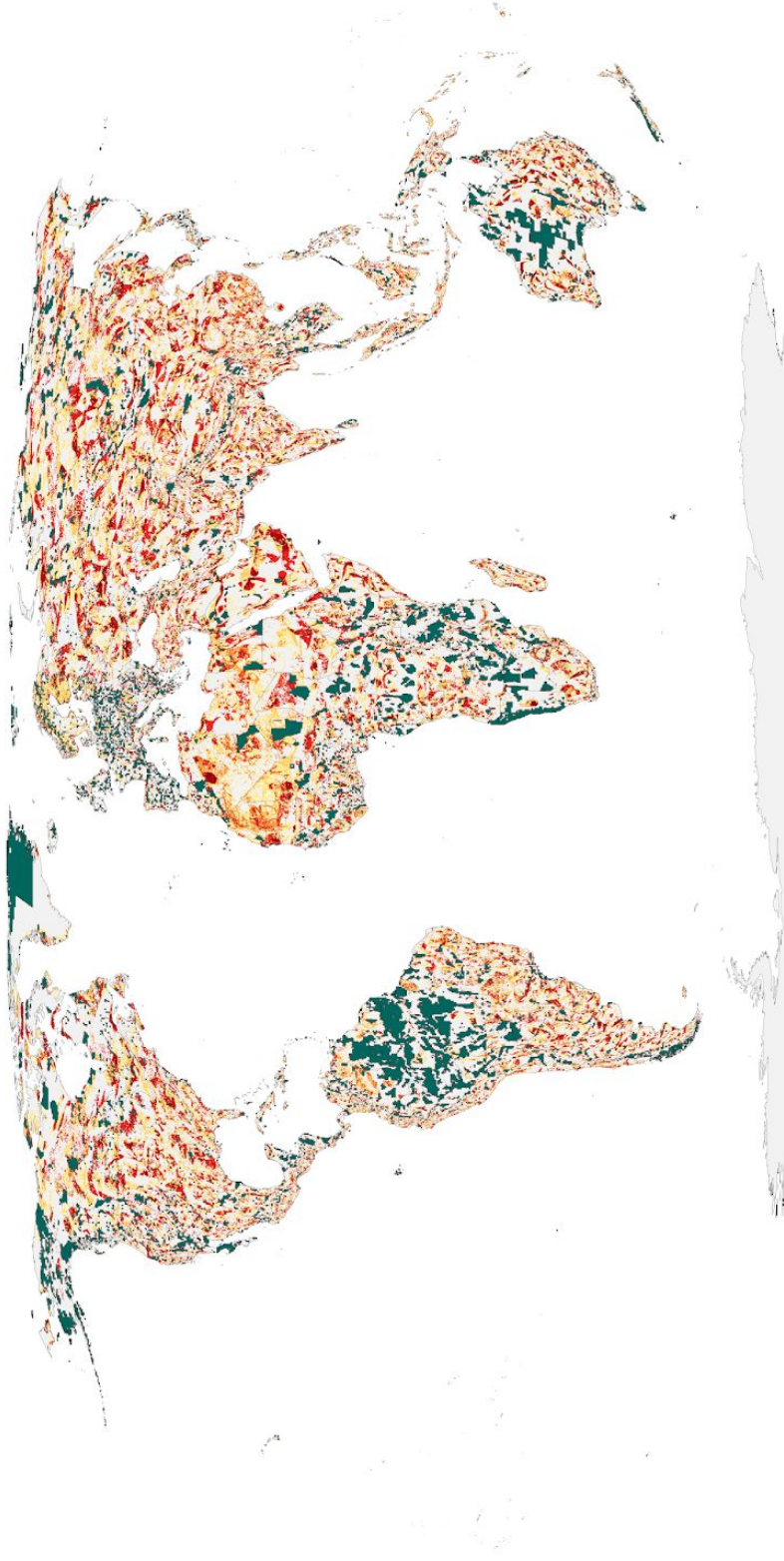


Figure S5. Areas of high scenario overlap (>10 scenarios, green) compared to Meyers et al. biodiversity hotspots (blue).

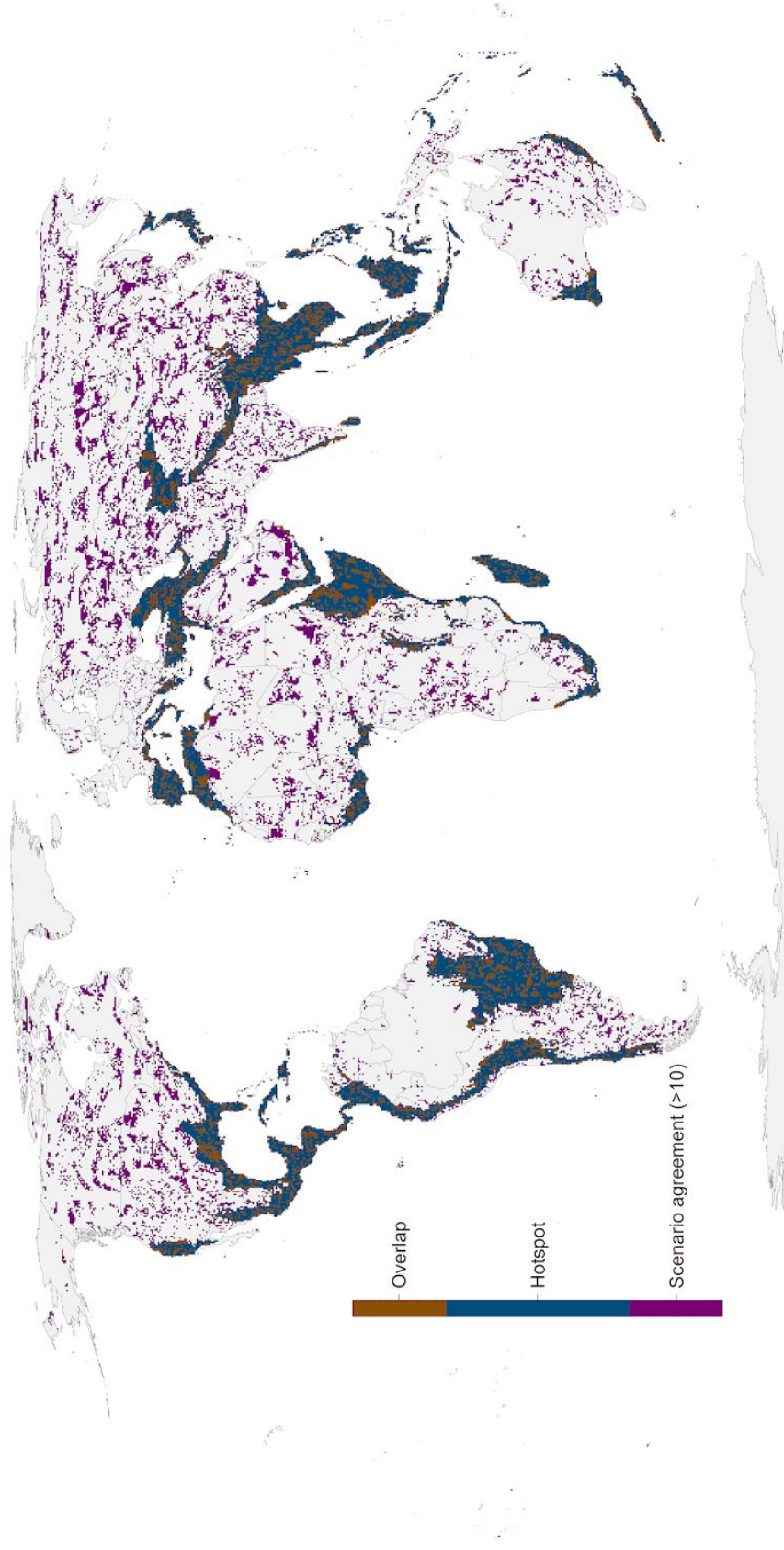


Figure S6: Influence of average country specific risk factors on the optimization outcomes compared between null scenario and the scenarios including one of the risk factors. Each data point represents the results for one country. The fitted blue lines and 95% confidence bands are from ordinary least-squares regression.

